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Title: Selective copper extraction by multi-modified mesoporous silica material, SBA-15

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Abstract: Selective copper (Cu) recovery from wastewater mitigates environmental pollution and is economically valuable. Mesoporous silica adsorbents with manganese loading and amine-grafting, SBA-15-NH₂ and Mn-SBA-15-NH₂, were synthesized. The characteristics of the synthesized adsorbents were evaluated in detail in terms of its crystal structure peaks, surface area and pore size distribution, transmission electron microscope and X-ray photoelectron spectroscopy. The results established the high selective Cu adsorption capacity of Mn-SBA-15-NH₂ in a mixed heavy metal solution, maintaining 96% competitive adsorption as that of a single Cu solution. Comparatively, Cu adsorption on SBA-15-NH₂ decreased by half due to high competition with other heavy metals. Optimal Cu adsorption occurred at pH 5. This pH condition enabled grafted amine group in Mn-SBA-15-NH₂ to form strong chelating bonds with Cu, avoiding protonation of amine group (below pH 5) as well as precipitation (above pH 5). The adsorption equilibrium well fitted to Langmuir and Freundlich isotherm models, while kinetic results were represented by models of linear driving force approximation (LDFA) and pore diffusion model (PDM). High regeneration and reuse capacity of Mn-SBA-15-NH₂ were well established by its capacity to maintain 90% adsorption capacity in a multiple adsorption-desorption cycle. Cu was selectively extracted from Mn-SBA-15-NH₂ with an acid solution.

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Prof Ravi Naidu
Guest Editor
Science of the Total Environment

UTS CRICOS PROVIDER CODE 00099F

Dear Prof. Ravi,

Please find attached our paper on “**Selective copper extraction by multi-modified mesoporous silica material, SBA-15**” by Seongchul Ryu, Gayathri Naidu, Hee Moon, and Saravanamuthu Vigneswaran for possible publication **in a special issue of the 1st Global CleanUp Congress 2018 in the Science of the Total Environment**. I declare that this paper is the original work of the authors and all the authors agree that it should be submitted to this journal. The manuscript was not previously been submitted to your journal or submitted concurrently to another journal.

The paper abstract is as below

Selective copper (Cu) recovery from wastewater mitigates environmental pollution and is economically valuable. Mesoporous silica adsorbents with manganese loading and amine-grafting, SBA-15-NH₂ and Mn-SBA-15-NH₂ was synthesized. The physical and chemical characteristics of the synthesized adsorbents were evaluated in detail in terms of its crystal structure peaks, surface area and pore size distribution, transmission electron microscope and X-ray photoelectron spectroscopy. The results established the high selective Cu adsorption capacity of Mn-SBA-15-NH₂ in a mixed heavy metal solution, maintaining 96% competitive adsorption as that of a single Cu solution. Comparatively, Cu adsorption on SBA-15-NH₂ decreased by half due to high competition with other heavy metals. Optimal Cu adsorption occurred at pH 5. This pH condition enabled grafted amine group in Mn-SBA-15-NH₂ to form strong chelating bonds with Cu, avoiding protonation of amine group (below pH 5) as well as precipitation (above pH 5). The adsorption equilibrium well fitted to Langmuir and Freundlich isotherm models, while kinetic results were represented by models of linear driving force approximation (LDFA) and pore diffusion model (PDM). High regeneration and reuse capacity of Mn-SBA-15-NH₂ were well established by its capacity to maintain 90% adsorption capacity in a multiple adsorption-desorption cycle. Cu was selectively extracted from Mn-SBA-15-NH₂ with an acid solution.

The authors believe that this paper fits the aim and scope of the journal on total environment, specifically covering the spheres of *hydrosphere and antrophere*. This is because this study evaluated the potential of remediating acidic mining wastewater (*hydrosphere*) containing heavy metals. Mining wastewater an environmental pollutant that affects the biosphere, hydrosphere, geosphere and antrophere. In this study, a green technology (mesoporous silica nanoparticle with high regenerative/reuse capacity) is used to selectively recover copper from the wastewater. Selective recovery of copper with a green technology is a sustainable and economically beneficial approach that can be adopted by the wastewater industry to attain resource recovery, rather than simply treat and dispose of the wastewater (*antrophere*).

Type of contribution: original research paper.

Yours faithfully,

S Vigneswaran

Title page

Selective copper extraction by multi-modified mesoporous silica material, SBA-15

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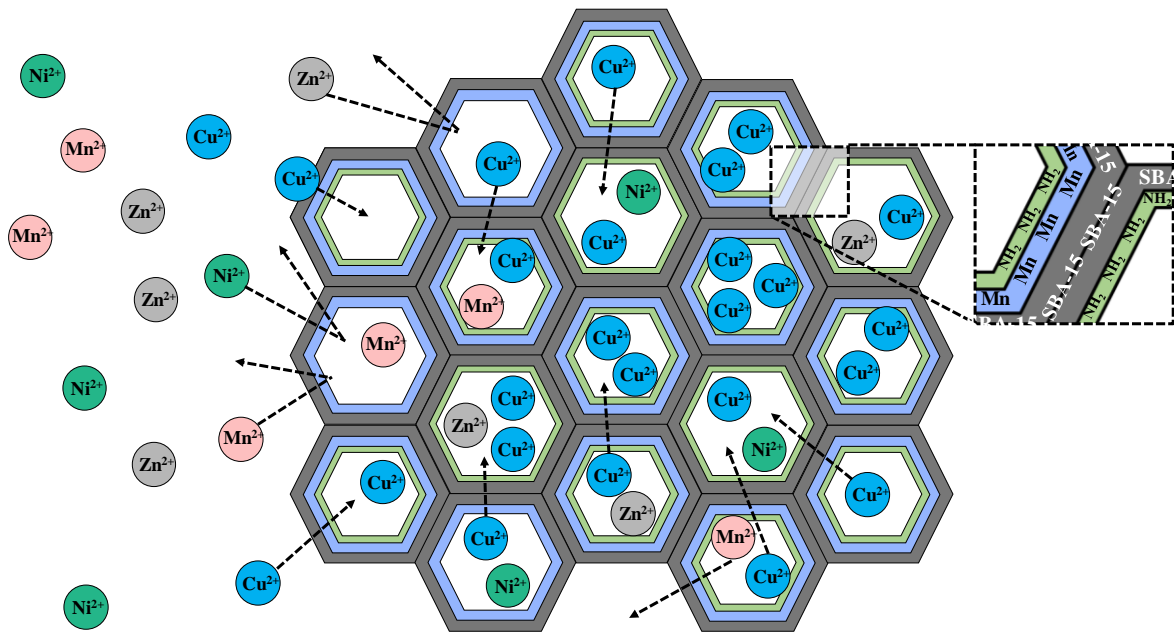
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Graphical abstract



Highlights

- Amine grafted Mn-SBA-15-NH₂ show high Cu adsorption unlike SBA-15-NH₂ and SBA-15.
- Mn-SBA-15-NH₂ maintained 96% selective Cu extraction in mixed heavy metal solution.
- Optimum pH is important for grafted amine group to form complexes with Cu.
- LDFA and PDM models represented kinetics of Cu diffusion satisfactorily.
- Mn-SBA-15-NH₂ maintained 90% adsorption capacity in multiple regeneration cycle

1 **Selective copper extraction by multi-modified mesoporous silica material,**
2 **SBA-15**

3
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8
9

10 **Abstract**

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12 economically valuable. Mesoporous silica adsorbents with manganese loading and amine-
13 grafting, SBA-15-NH₂ and Mn-SBA-15-NH₂, were synthesized. The characteristics of the
14 synthesized adsorbents were evaluated in detail in terms of its crystal structure peaks, surface
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28

29

30 **Keywords**

31 Mesoporous silica; Amine-grafting; Manganese loading; Adsorption; Copper; Heavy metals

32

33 1. Introduction

34 The discharge of industrial and mining wastewater containing elevated concentration of
35 heavy metal results in severe water contamination (Da'na and Sayari, 2012; Mureseanu et al.,
36 2008; Naidu et al., 2019; Zhao et al., 2017). Heavy metals are not biodegradable, and
37 therefore, tend to accumulate and remain in water bodies for a long time. This poses long
38 term detrimental effect to ecosystem and human health. Amongst the heavy metals, copper
39 (Cu) tends to be present in high concentrations in wastewater. This is because Cu is regularly
40 used in various industrial applications, such as printed circuit boards, semiconductor devices,
41 pipelines, and battery manufacturing (Cui and Zhang, 2008; Hadi et al., 2015; Lee et al., 2016;
42 Niu and Li, 2007; Xue et al., 2012). Wastewater containing Cu must be treated prior to
43 discharge as Cu is highly toxic, even if present in trace concentration (Nagajyoti et al., 2010).
44 At the same time, Cu is an economically valuable metal (Glöser et al., 2015) given its vast
45 industrial application. Thus, removing and selectively recovering Cu from wastewater, will be
46 both environmentally and economically beneficial.

47 Various remediation approaches are generally applied for heavy metal removal from
48 wastewater such as chemical precipitation, ion exchange/adsorption, membrane filtration, and
49 electrochemical methods (Fu and Wang, 2011). Adsorption is a promising approach due to its
50 convenience, ease of operation and cost effectiveness (Cai et al., 2019; Da'na and Sayari,
51 2012). More importantly, adsorption possess the capacity to selectively recover valuable
52 metals such as Cu from wastewater containing mixed constituent of heavy metals.
53 Conventionally, absorbents such as zeolite (Ryu et al., 2019), pumice (Yavuz et al., 2008),
54 activated carbon (Anirudhan and Sreekumari, 2011), and chitosan (Feng et al., 2019; Xiao et
55 al., 2019) are used for removal of heavy metals from wastewater. However, these adsorbents

56 exhibit low heavy metal adsorption capability for practical application and do not possess
57 selective adsorption capacity in mixed solutions. One potential method to enhance heavy
58 metal removal efficiency is by increasing the surface area of absorbents.

59 In this regard, mesoporous silica is a promising material that possess large surface area. This
60 apart, mesoporous silica exhibit several other advantageous features such as high thermal and
61 physical stabilities, tuneable and uniform porosity and flexibility towards surface chemical
62 modification (McManamon et al., 2012; Szegedi et al., 2011). In particular, mesoporous
63 materials of SBA series consist of thick pore walls, large and tuneable pore size (2 – 30 nm)
64 (Anbia et al., 2010). The triblock copolymer surfactant acts as a structure-directing agent
65 (SDA) for SBA-15 preparation (Kim et al., 2014). However, in spite of the various beneficial
66 features of SBA-15, in its natural condition, it cannot selectively adsorb contaminants such as
67 heavy metals from wastewater as it lacks active sites and attractive functional groups for ions
68 (Jiang et al., 2007; Liang et al., 2017).

69 Thus, chemical functionalization of SBA-15 is necessary to produce active sites for
70 adsorption and enhance selective removal of heavy metals. A number of different SBA-15
71 modification methods have been explored (Ezzeddine et al., 2015; Hernández-Morales et al.,
72 2012; Jiang et al., 2007; Liang et al., 2017). For instance, Saad et al. (2007) synthesized and
73 modified SBA-15 using an amino-group source of (3-aminopropyl) triethoxysilane (APTES)
74 with two different modification methods, co-condensation and post synthesis (Saad et al.,
75 2007). Amine grafted SBA-15 enabled to achieve high heavy metal removal because amino-
76 propyl moieties reacts with silanol groups on the surface (Benhamou et al., 2013; Da'na et al.,
77 2011; Yokoi et al., 2006). In another study, Lei et al (2015) carried out a combined
78 modification of manganese loading and amine-grafted SBA-15 for heavy metal adsorption.
79 Amine grafting on SBA-15 with manganese loading enabled to achieve high selective Cu

80 extraction in a competitive adsorption study (Lei et al., 2015). The mechanism of manganese
81 modification with amine on SBA-15 has been evaluated in detail for applications such as
82 catalyst (Saikia and Srinivas, 2009; Saikia et al., 2007). However, there is still limited
83 research with regards to the mechanism of manganese modified SBA-15 with amine for
84 heavy metal adsorption. Further, theoretical analysis based on isotherm parameters such as
85 surface and pore diffusion models and selective capacity of Cu in a mixed heavy metal
86 solution has not been explored in detail.

87 In this study, the role of manganese loading and grafted amine-groups on SBA-15 towards
88 selective Cu removal is investigated in detail. For this reason, SBA-15 synthesis was carried
89 out in a batch hydrothermal reaction and modified by manganese loading and amine-grafting.
90 The characteristics and adsorption capacity of the modified SBA-15 for heavy metal (Cu, Zn,
91 Ni and Mn) removal were evaluated in detail. Experimental data were fitted to equilibrium
92 models such as Langmuir and Freundlich and kinetic models such as linear driving force
93 approximation and pore diffusion models. Further, regeneration and reusability of the
94 adsorbents were also evaluated to establish the cost effectiveness and practical applicability
95 of the adsorbents.

96

97 2. Material and methods

98 2.1. Materials

99 Tetraethyl orthosilicate (TEOS) (98%, Sigma-Aldrich) and poly(ethylene glycol)-block-
100 poly(propylene glycol)-block-poly(ethylene glycol) (P123) were used as a silica precursor
101 and a structure directing agent (SDA), respectively. 2 M of hydrochloric acid (HCl) was
102 prepared by using 34% HCl solution purchased from Sigma-Aldrich to hydrolyze the silica
103 precursor, TEOS, for self-assembly reaction onto the SDA, P123. KMnO_4 (Sigma-Aldrich)

104 was used for manganese modification. (3-aminopropyl) triethoxysilane (APTES, Sigma-
105 Aldrich) and toluene were used for amine-grafting onto the mesoporous silica material, SBA-
106 15. CuSO₄, ZnSO₄, MnSO₄ and NiNO₃ were used to prepare heavy metal solutions for batch
107 adsorption experiments. 0.01 M of sodium hydroxide (NaOH) (>98%, Sigma-Aldrich) and
108 HCl solutions were used for pH adjustment. 0.01 M HCl solution was used in desorption
109 experiments.

110

111 2.2. Preparation of adsorbent

112 SBA-15 was synthesized via a hydrothermal reaction based on a previously reported method
113 (Zucchetto et al., 2018). 12 g of P123 was added in a mixture of 90 ml of Milli-Q water and
114 480 ml of 2 M HCl at a room temperature (24 ± 1 °C). 25.5 g TEOS was then added slowly
115 into the mixture for the hydrolysis by acidic condition (Wei et al., 2006). Thereafter, the
116 mixture was kept at 35 °C for 20 h and 100 °C for 24 h in a polypropylene bottle for
117 hydrothermal reaction. The precipitated powder was then dried at 70 °C in an oven after
118 filtration and washed with Milli-Q water. It was then calcined at 550 °C for 3 h in an air
119 condition to get rid of the surfactant, P123, from the ordered structure with silica sources.

120

121 2.3. Modifications of adsorbent

122 2.3.1. Manganese loading

123 Modification of SBA-15 with manganese loading was carried out by adding 1 g of the
124 prepared SBA-15 into 200 ml of 0.1 M KMnO₄ aqueous solution (Lei et al., 2015). The
125 mixture of SBA-15 and KMnO₄ was mixed for 3 h at ambient temperature. The resultant
126 solid powder was filtered and washed with Milli-Q water thoroughly and thereafter oven
127 dried at 70 °C, to produce manganese doped mesoporous silica, Mn-SBA-15.

128

129 2.3.2. Amine-grafting

130 Surface modification of mesoporous silica material by amine-groups was carried out using
131 post-grafting method based on a previously reported method (Da'na and Sayari, 2011; Saad
132 et al., 2007). 1.0 g of prepared SBA-15 and Mn-SBA-15 were dispersed into 100 ml of dry
133 toluene. Thereafter, 1 ml of APTES was added. The solution was refluxed for 10 h at 110 °C
134 to enable reaction to occur between OH groups of mesoporous silica's surface and amino-
135 propyl groups of the APTES. The final products SBA-15-NH₂ and Mn-SBA-15-NH₂ were
136 filtered, washed thoroughly with ethanol and oven-dried at 70 °C.

137

138 2.4. Absorbent characterizations

139 2.4.1. Crystal structure

140 Powder X-ray diffraction (XRD) analysis was used to determine the structure of the
141 adsorbents prepared. XRD analysis was carried out with a X-ray diffractometer with high
142 resolution (X'Pert PRO Multi-Purpose X-Ray Diffractometer) operated with the radiation
143 source of Cu-K α . The system was operated at room temperature in 2 theta range of 0 to 10 °
144 with a 2°/min scan rate at 20 mA and 40 kV.

145

146 2.4.2. Surface area and pore size distribution

147 The physical properties (surface area and pore size distribution) of the prepared adsorbents
148 were determined after performing nitrogen adsorption/ desorption isotherm experiments.
149 They were measured at 77 K (Nanoporosity, Mirae SI Korea). The mesopore volume and the
150 pore size distribution were estimated by Barrett-Joyner-Halenda (BJH) method, while surface
151 areas of the samples were obtained from the Brunauer-Emmett-Teller (BET) method.

152

153 2.4.3. Surface morphology and element contents

154 The surface morphologies of adsorbents were investigated using a field emission
155 transmission electron microscope (FETEM, JEM-2100F) at 200 kV and the element contents
156 of all samples were identified by X-ray photoelectron spectroscopy (XPS) using $K\alpha$ radiation
157 with seven Channeltron detectors. Fourier transform infrared spectroscopy (FTIR)
158 spectrometer (FTIR-410, Jasco Co.) was used to obtain for investigating the chemical bonds
159 and vibrations.

160

161 2.5. Influence of pH

162 pH effect on heavy metal adsorption was investigated using solutions of different pH (1 to 5).
163 They were prepared with 0.1 M NaOH and HCl aqueous solutions. 0.01 g of the prepared
164 adsorbent was suspended in beakers containing 50 ml of Cu solution (300 mg/L) with varied
165 initial pH. The flasks were agitated at 120 rpm for 24 h in a flat shaker at ambient
166 temperature (24 ± 1 °C).

167

168 2.6. Adsorption study

169 2.6.1. Equilibrium

170 Batch equilibrium adsorption of heavy metals (Cu, Zn, Ni and Mn) on the prepared pristine,
171 manganese-doped, and amine-grafted adsorbents was carried out. 0.01 g of each adsorbents
172 and 50 ml of heavy metal solution were mixed into beakers. The beakers were placed in a flat
173 shaker at room temperature (24 ± 1 °C) for a day to reach equilibrium. The preliminary
174 experiment showed that a period of 24 h was sufficient to reach equilibrium. Single-
175 component adsorption experiments were conducted with 10 – 300 mg/L of initial

176 concentrations of heavy metals, while 10 to 100 mg/L of metals were used for multi-
177 component adsorption experiment. These concentrations selected were based on the presence
178 of heavy metals in wastewater (Jalali and Moradi, 2013; Özdemir et al., 2009; Šćiban et al.,
179 2007; Singh and Agrawal, 2007). Microwave Plasma-Atomic Emission Spectrometer (MP-
180 AES, MP-AES-4100 Agilent) was used to determine the heavy metal concentrations in
181 treated solutions. The equilibrium adsorption capacity, Q_e (mmol/g), was determined using
182 Eq. (1):

183

$$184 \quad Q_e = V(C_i - C_e)/M \quad (1)$$

185

186 where, V is the solution volume (L) and M is sorbent mass (g). C_i and C_e are the
187 concentrations (mg/L) of the heavy metals at initial and equilibrium, respectively.

188 Langmuir and Freundlich isotherm models were employed to obtain equilibrium adsorption
189 data:

190

$$191 \quad \text{Langmuir isotherm: } Q_e = Q_m b C_e / (1 + b C_e) \quad (2)$$

192

$$193 \quad \text{Freundlich isotherm: } Q_e = K_F C_e^{1/n} \quad (3)$$

194

195 where, Q_m and b are the maximum sorption capacity (mmol/g) and Langmuir constant
196 (L/mmol) related to the affinity of the binding sites, respectively. Freundlich constant is
197 related to sorption affinity, K_F ($\text{g}^{1-n} \text{L}^n \text{g}^{-1}$) and 1/dimension less parameter ($1/n$) was related to
198 surface heterogeneity (Ali et al., 2016; Foo and Hameed, 2010; Gonzales et al., 2016) were
199 also used in adsorption modeling.

200

201 2.6.2. Kinetics

202 The kinetic adsorption experiments were carried out in batch systems. Each beaker contained
203 0.01 g adsorbent and 50 ml of 50 mg/L Cu solution. The beaker with the solution was
204 agitated at room temperature (24 ± 1 °C) and samples were collected at different time
205 intervals until 24 h for metal analysis. In this study, the adsorption data were analyzed by
206 linear driving force approximation (LDFA) model due to its simplicity and feasibility in
207 describing adsorption of Cu ions from aqueous solution on modified SBA-15 samples (Kim
208 et al., 2008). The mass transfer rate between solid and liquid phases can be described by the
209 LDFA model as:

210

$$211 \quad d\bar{Q}/dt = 3k_f(C_b - C_s)/R_p\rho_p = k_s(Q_s - \bar{Q}) \quad (4)$$

212

213 where

214

$$215 \quad k_s = 15D_s/R_p^2 \quad (5)$$

216

217 where k_s is the mass transfer coefficient of solid side (1/s), k_f is the film mass transfer
218 coefficient (m/s), R_p is the radius of adsorbent particle (m), C_b is bulk concentration (mg/L),
219 C_s is equilibrium concentration (mg/L), \bar{Q} is the average value of adsorbed particle amounts,
220 t is the time passed in a batch reactor, Q_s is the concentration of adsorbed phase at the
221 external surface of adsorbent particle (mmol/g), ρ_p is the particle density (kg/m^3), and D_s is
222 effective surface diffusion coefficient (m^2/s).

223 Adsorption occurs both through external film and internal pore (Hariharan et al., 2012). Thus,

224 the kinetic data were also determined by pore diffusion model (PDM) using the following
225 equations:

226

$$227 \quad \varepsilon_p(\partial C_p/\partial t) + \rho_p(\partial Q/\partial t) = D_p[(\partial^2 C_p/\partial r^2) + (\partial C_p/\partial r)/r] \quad (6)$$

228

229 Corresponding boundary conditions and initial concentration are:

230

$$231 \quad \partial C_p/\partial r = k_f(C_b - C_p)/D_p\varepsilon_p \text{ at } r = R_p \quad (7)$$

$$232 \quad C_p = C_0 \text{ at } t = 0 \quad (8)$$

$$233 \quad \partial C/\partial r = 0 \text{ at } r = 0 \quad (9)$$

234

235 where ε_p is the particle porosity, C_p is the concentration of ions at the inside of particle
236 (mg/L), D_p is the effective pore diffusion coefficient (m^2/s), and r is the radial distance (m).

237

238 2.7. Metal desorption and adsorbent regeneration

239 Reusability of the manganese doped and/or amine-grafted SBA-15 was tested with 10-time

240 adsorption and desorption recycle experiments. After each adsorption experiments, spent

241 adsorbents were recovered by filtration from the working solution. For the desorption step,

242 the recovered powder was treated by 0.01 M HCl solution for Cu recovery (Da'na and Sayari,

243 2011). Acidic solution can bring about heavy metal desorption through protonation (NH_3^+) of

244 the grafted amine groups (NH_2). The recovered adsorbents were dried at 80 °C after filtration

245 and washing with Milli-Q water. Prior to the use, the dried white powder has to be neutralized.

246 Therefore, the powder was stirred in a solution of 0.01 M NaOH for 1 h. For the further

247 adsorption experiments, the regenerated silica powder was used with another heavy metal

248 solution.

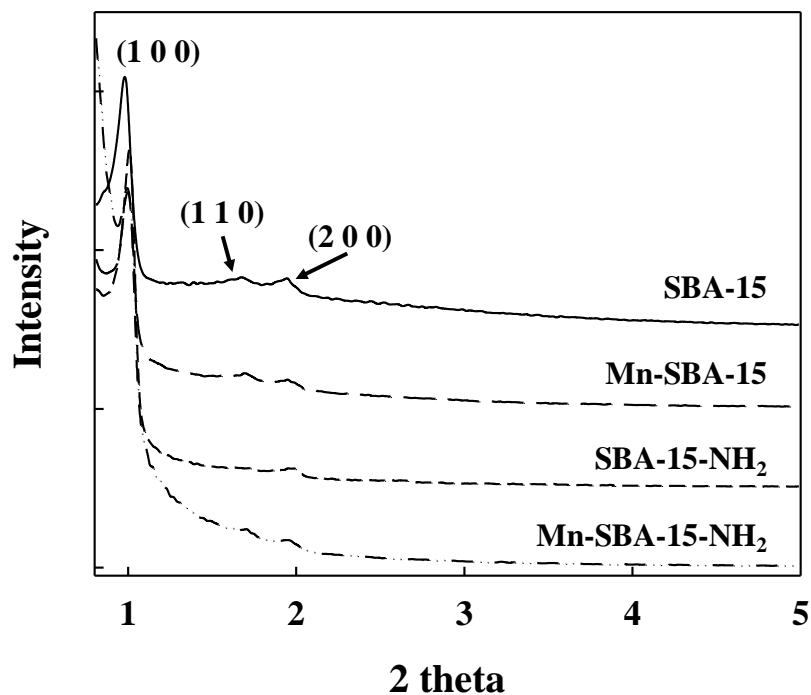
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250 3. Results and discussion

251 3.1. Adsorbent characterization

252 3.1.1. Crystal structure

253 X-ray diffraction pattern by XRD analysis revealed that the crystal structure pattern of
254 adsorbents SBA-15, Mn-SBA-15, SBA-15-NH₂ and Mn-SBA-15-NH₂ contained three peaks
255 in the low 2 theta range of 0 – 10° (Fig. 1). Previous studies have established the crystal
256 structure of a virgin SBA-15 contain one high peak (1 0 0), and two tiny peaks of (1 1 0) and
257 (2 0 0) (Kim et al., 2015). The XRD results implies that the prepared SBA-15 in this work
258 was synthesized well with highly ordered hexagonal structured material. The three crystal
259 peaks were still maintained even after Mn modification on the virgin SBA-15. This indicates
260 that the 2-dimensional (2D) hexagonal structure of SBA-15 was stable during Mn loading
261 (Lei et al., 2015). All the three adsorbents showed closely similar intensity of the main peak
262 (1 0 0), indicating that hexagonal structures of SBA-15 was not collapsed after the
263 modification processes (Ryu et al., 2018). However, 2 peaks of (1 1 0) and (2 0 0)
264 significantly decreased after amine-grafting on SBA-15 and Mn-SBA-15. Other studies
265 (Aguado et al., 2009; Calleja et al., 2011; Chang et al., 2009) also showed the decrease in
266 minor peaks of SBA-15-NH₂ (Shahbazi et al., 2011). The decrease is associated to pore
267 blocking by amine (–NH₂) groups bonded on the surface of SBA-15.



268

269 Fig. 1. The XRD patterns of SBA-15, Mn-SBA-15, SBA-15-NH₂ and Mn-SBA-15-NH₂

270

271 3.1.2. Chemical properties

272 Fig. 2. Presents FT-IR spectra of the SBA-15, Mn-SBA-15, SBA-15-NH₂ and Mn-SBA-15-

273 NH₂. Typically, SBA-15 consists of Si-O-Si bands that appear around 1080 and 800 cm⁻¹ and

274 hydroxyl group (-OH) attributed to silica network shown at around 960 cm⁻¹ (Wang et al.,

275 2015). After Mn modification, a decrease in the vibration band for Si-OH at 970 cm⁻¹ was

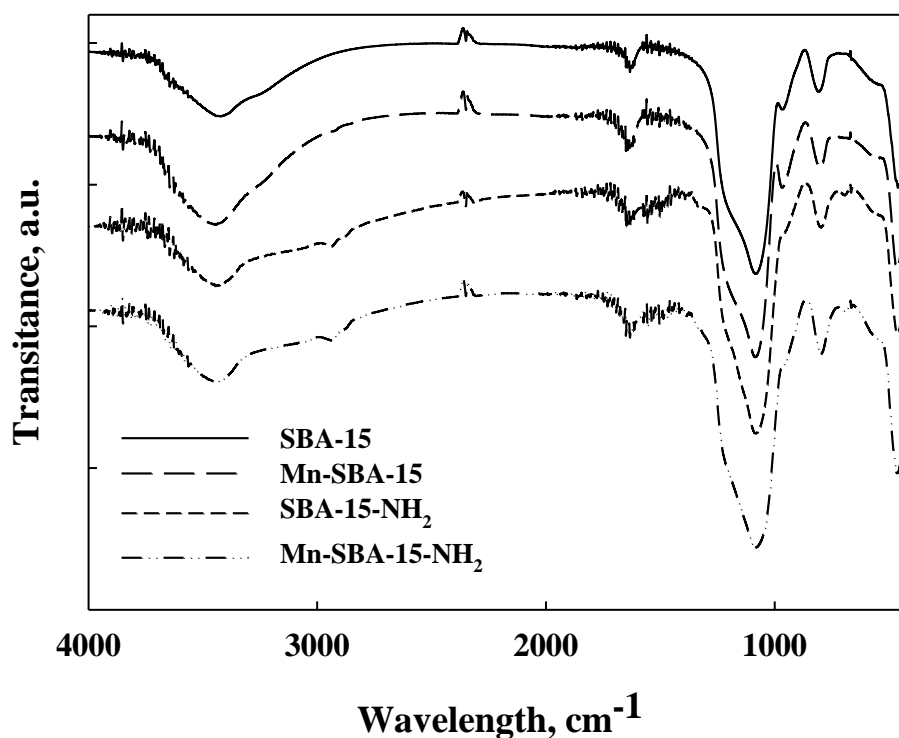
276 observed. This implies the successful reaction between Si-OH and Mn (Li et al., 2007).

277 Additional spectra at around 2900 and 1500 cm⁻¹ were observed on the modified SBA-15-

278 NH₂ and Mn-SBA-15-NH₂. They are attributed to N-H vibration band by amine-grafting on

279 SBA-15 (Szegedi et al., 2011; Yan et al., 2011). The FT-IR spectra enabled to confirm

280 whereas the modifications of SBA-15 were successfully carried out or not.



281

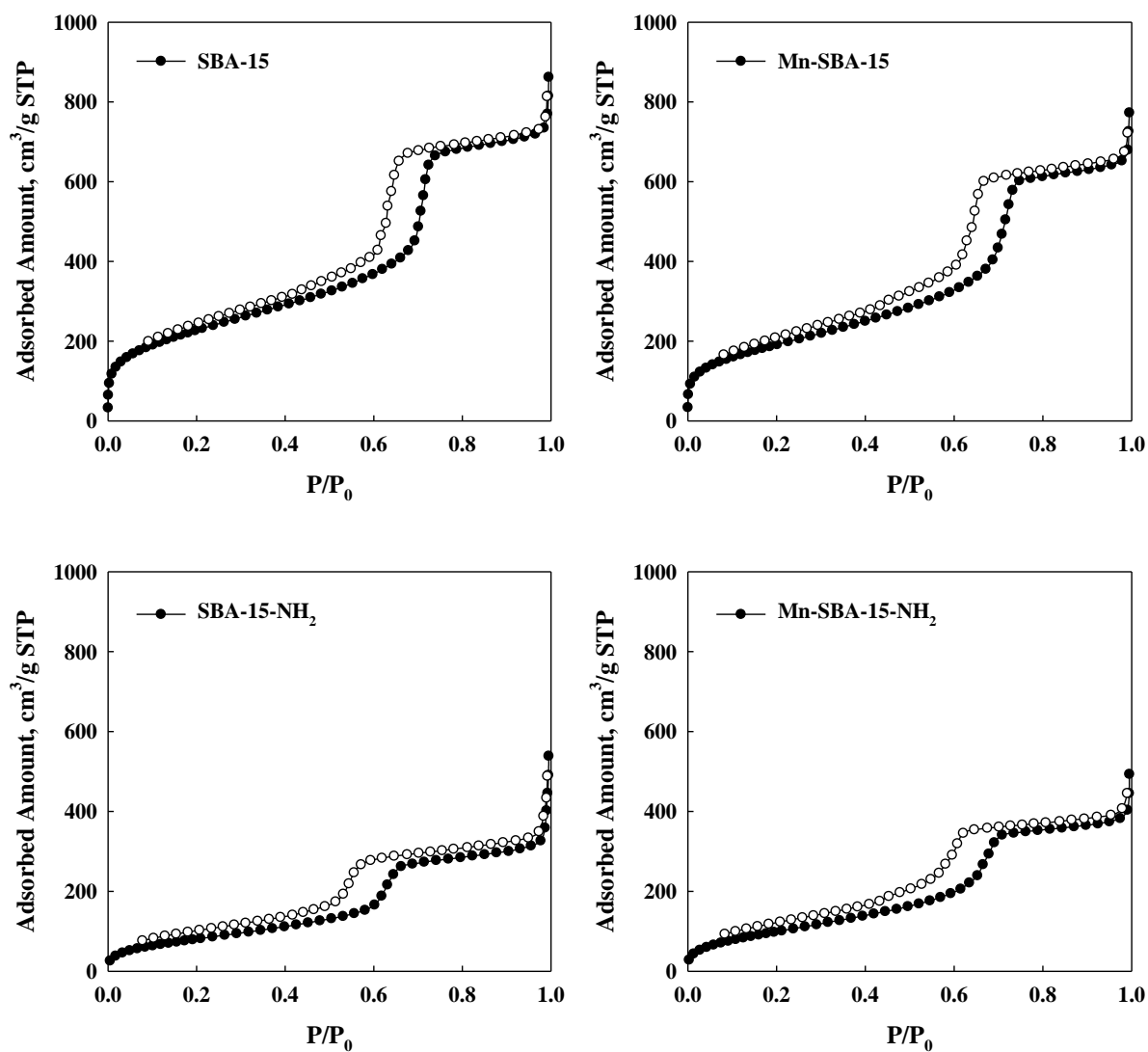
282 Fig. 2. FT-IR spectra of prepared SBA adsorbents

283

284 3.1.3. Surface area and pore size distribution

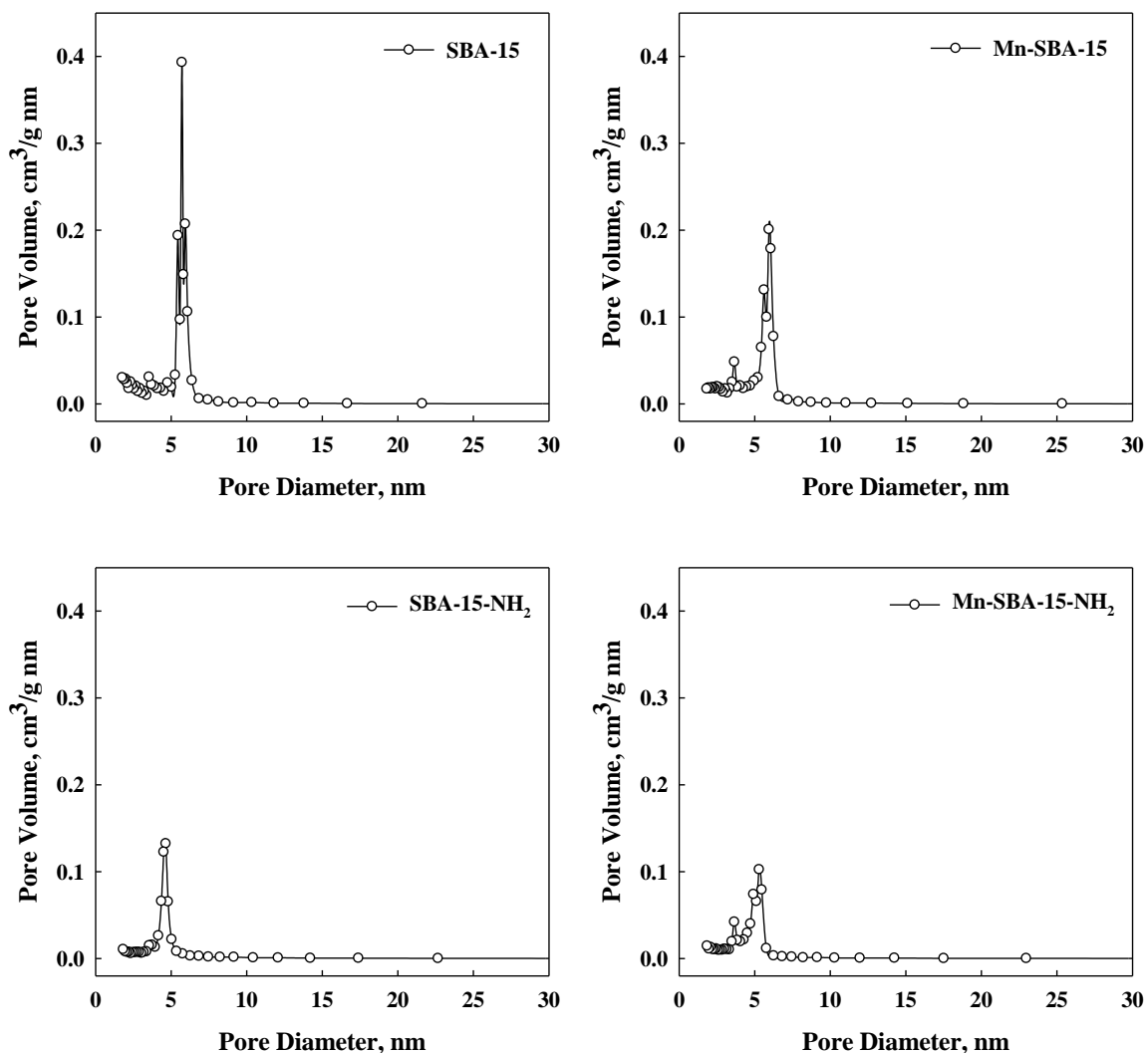
285 The physical properties (surface area and pore size distribution) of the adsorbents are
 286 summarized in Table 1. The adsorption and desorption isotherm of nitrogen at 77 K on
 287 adsorbents (Fig. S1), show that all graphs exhibit type IV adsorption based on the BDDT
 288 classification. This is in line with the results on mesoporous materials (Lombardo et al.,
 289 2012). The specific surface area of virgin SBA-15 measured by BET method was 830 m²/g
 290 for the virgin SBA-15. The BET surface area of Mn-SBA-15, SBA-15-NH₂ and Mn-SBA-15-
 291 NH₂ decreased to 700, 310 and 310 m²/g due to pore blocking by chemical bonding of SBA-
 292 15 with Mn and amine functional groups. These results can be explained by the porosity of
 293 adsorbents. The pore size distributions of samples are shown in Fig. S2. The peak position

294 shifted from around 6 nm to 5 nm after amine-grafting on SBA-15. On the other hand, an
295 additional minor peak was visible at around 4 nm with Mn modified adsorbents (Mn-SBA-15
296 and Mn-SBA-15-NH₂). It implies that some portion of pores were blocked by Mn application
297 which caused in decrease of the pore size. Total pore volume of virgin SBA-15 was 1.24
298 cm³/g and it decreased to 1.11, 0.73, and 0.76 cm³/g for Mn-SBA-15, SBA-15-NH₂, and Mn-
299 SBA-15-NH₂, respectively. Likewise, mean pore diameter of SBA-15 and Mn-SBA-15
300 reduced from 4.6 and 4.5 to 4.0 nm after amine grafting. Overall, the physical properties of
301 the adsorbents demonstrate the success in modification of SBA-15.



302

303 Fig. S1. Isotherms for nitrogen adsorption and desorption of prepared samples



304

305 Fig. S2. Adsorbent BJH pore size distribution profiles

306

307 Table 1. Adsorbent physical properties

Adsorbents	BET surface area m ² /g	Total pore volume cm ³ /g	Mean pore diameter nm
SBA-15	830	1.24	4.6
Mn-SBA-15	700	1.11	4.5
SBA-15-NH ₂	310	0.73	4.0

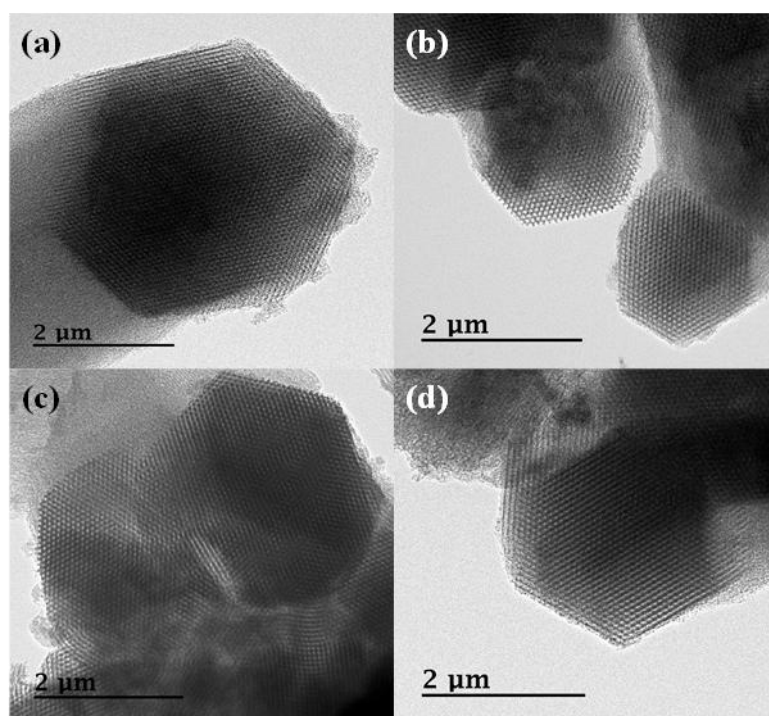
Mn-SBA-15-NH ₂	310	0.76	4.0
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308

309 3.1.4. Surface morphology and element content

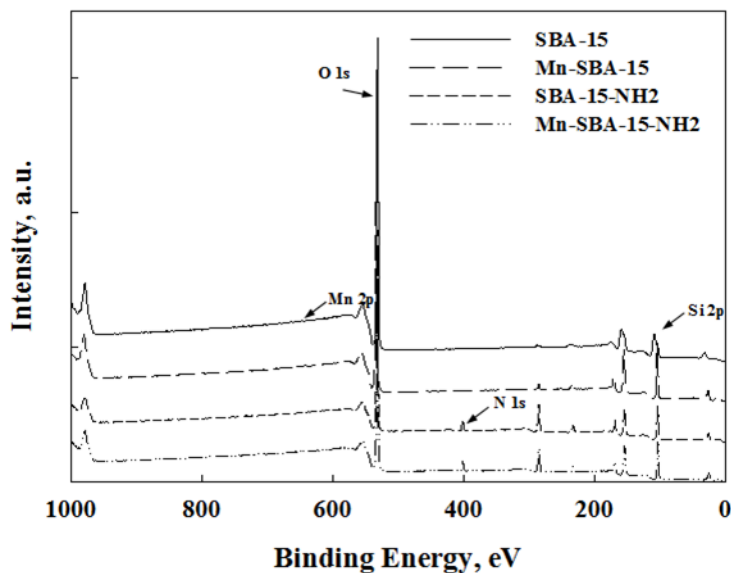
310 TEM images of the virgin and modified SBA-15 adsorbents (Fig. 3) exhibit the well-ordered
311 and 2D hexagonal structured array of the mesoporous material. The surface morphology
312 indicated that the modified adsorbents were able to maintain well-ordered structure as that of
313 the virgin SBA-15. The XPS spectra revealed element contents of Si, O, N and Mn of the
314 adsorbents (Fig. 4). In particular, high resolution XPS curves of O 1s and Si 2p in the region
315 of 530 – 535 eV and 103 – 110 eV were attributed to Si-O-Si bonds (Table 2). The presence
316 of N 1s peaks in the range of 400 – 405 eV on SBA-15-NH₂ and Mn-SBA-15-NH₂ indicated
317 the presence of NH₂ upon amine grafting (Gao et al., 2007). Further, minor peaks around 645
318 eV were observed on Mn-SBA-15-NH₂ and this was attributed to Mn loading into the pore of
319 SBA-15.

320



321

322 Fig. 3. TEM images of adsorbents (a) SBA-15, (b) Mn-SBA-15, (c) SBA-15-NH₂ and (d)
 323 Mn-SBA-15-NH₂



324
 325 Fig. 4. XPS spectra of virgin and modified SBA-15

326

327 Table 2. XPS binding energy (eV) of virgin and modified SBA-15

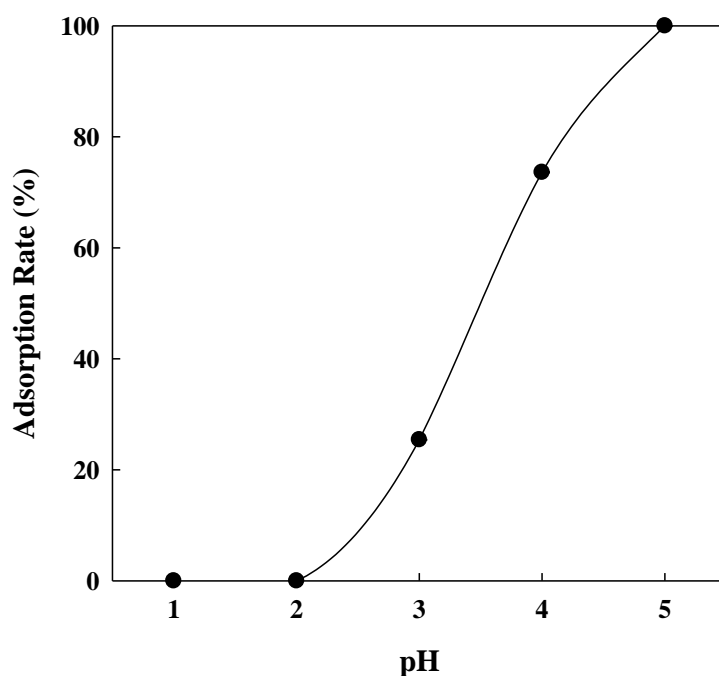
Adsorbents	Si 2p	O 1s	N 1s	Mn 2p
SBA-15	108.6	531.9	-	-
Mn-SBA-15	104.1	533.2	-	646.4
SBA-15-NH ₂	103.3	532.3	401.76	-
Mn-SBA-15-NH ₂	103.2	532.2	401.33	645.1

328

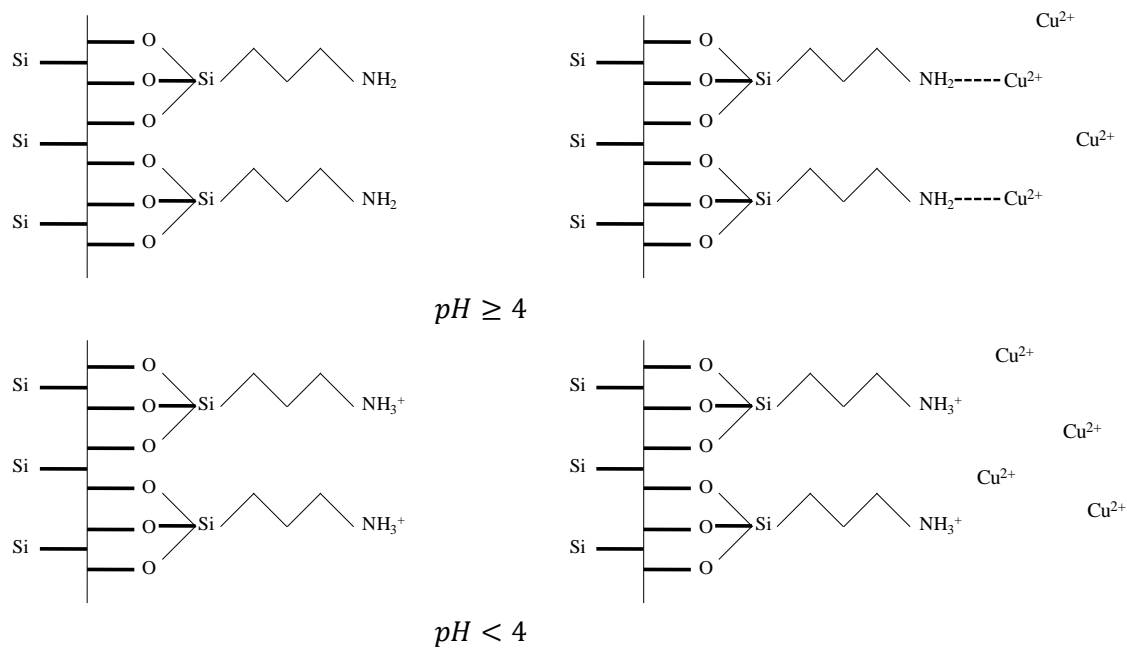
329 3.1.5. Influence of pH

330 Cu adsorption by SBA-15-NH₂ was highly influenced by pH of the solution (Fig. 5). Minimal
 331 Cu adsorption (close to zero) occurred at pH of 1 – 2. This was due to protonation of amine
 332 groups, –NH₂ to –NH₃ (Da'na and Sayari, 2011). A trend of higher Cu adsorption capacity
 333 was observed as the solution pH was increased above 2. A maximum Cu adsorption was

334 achieved at pH 5. The mechanisms describing the influence of pH towards Cu adsorption by
335 amine grafted SBA-15 is depicted in Fig. 6 (Da'na and Sayari, 2011). The grafted amine
336 groups (NH_2) on SBA-15 easily form complexes with Cu ions at pH above 4. However, the
337 protonated amine groups, $-\text{NH}_3^+$, do not form bonds with Cu ions easily when the pH value is
338 below 4. Meanwhile, above pH 5, a tendency of copper precipitation occurs. Hence all heavy
339 metal equilibrium experiments were carried out at optimum pH 5.
340



341
342 Fig. 5. Adsorption rate of Cu on SBA-15-NH₂ as a function of solution pH (experimental
343 conditions: adsorbent dose: 0.2 g/L, Cu concentration: 300 mg/L, equilibrium time: 24 h)
344



345

346 Fig. 6. Depiction of Cu adsorption mechanisms on the amine-grafted SBA-15

347

348 3.2. Adsorption study

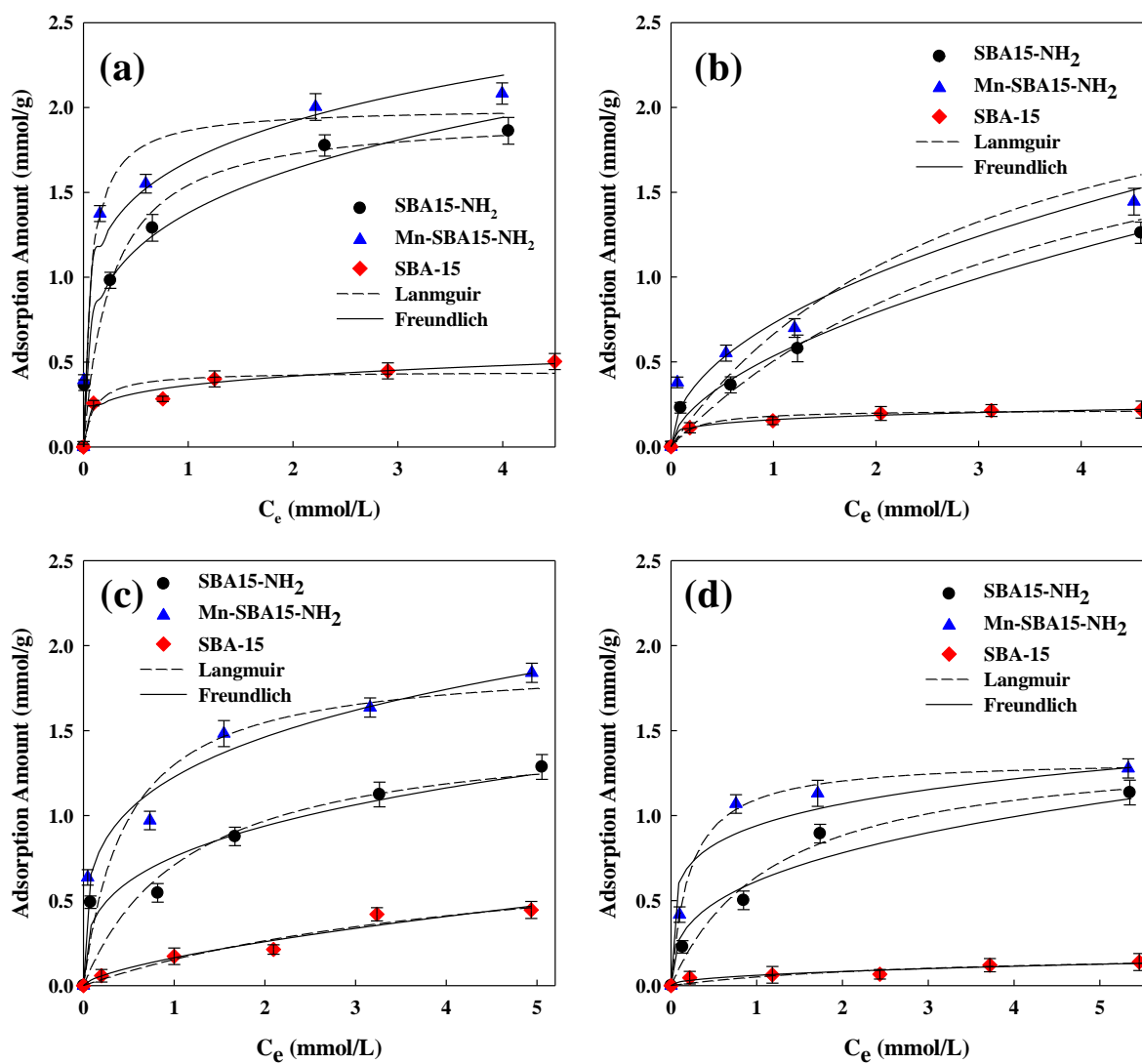
349 3.2.1. Equilibrium

350 3.2.1.1. Single component adsorption

351 Fig. 7 shows the adsorption isotherms at room temperature (24 ± 1 °C) for individual heavy
 352 metals (Cu, Zn, Ni and Mn) on prepared adsorbents (SBA-15, SBA-15-NH₂, and Mn-SBA-
 353 15-NH₂) at pH 5. All data were satisfactorily fitted with both Langmuir and Freundlich
 354 equations according to R^2 values (Table 3). The Langmuir model describes monolayer
 355 adsorption, thus Freundlich isotherm is the practical model on adsorption with heterogeneous
 356 surface (Popuri et al., 2009; Shahbazi et al., 2011). The heavy metal adsorption capacity by
 357 SBA-15 was minimal (0.14 – 0.50 mmol/g). Comparatively, the adsorption capacity of heavy
 358 metals by SBA-15-NH₂ were 4 – 9 times higher. Specifically, adsorption capacity of Cu, Zn,
 359 Ni and Mn were 1.86, 1.26, 1.28 and 1.14 mmol/g with SBA-15-NH₂. Meanwhile, Mn-SBA-
 360 15-NH₂ showed improved adsorption capacities of heavy metals than SBA-15-NH₂ (12 – 43%

361 higher). Specifically, adsorption capacity of Cu, Zn, Ni and Mn were 2.08, 1.44, 1.84 and 1.28
362 mmol/g with Mn-SBA-15-NH₂. The amine group grafted on the modified adsorbent is a
363 major factor that enhances adsorption capacity of heavy metals compared to virgin SBA-15
364 (Aguado et al., 2009). Further, in the case of Mn-SBA-15-NH₂, a large portion of amine
365 groups is able to be grafted on Mn-SBA-15. This is because manganese oxide reacts well
366 with amine (Zhao et al., 2015; Zhu and He, 2012). Thus, Mn-SBA-15-NH₂ was able to
367 achieve higher heavy metal adsorption capacity compared to SBA-15-NH₂.
368 In individual single metal solutions, the heavy metal adsorption capacity with SBA-15, SBA-
369 15-NH₂ and Mn-SBA-15-NH₂ followed the order of Cu>Ni>Zn>Mn. Likewise, a previous
370 study (Da'na and Sayari, 2012; Mureseanu et al., 2008; Sreenu et al., 2016) showed higher Cu
371 adsorption compared to other heavy metals such as Ni and Zn when mesoporous silica
372 material with amine grafting was used.

373



374
375

376 Fig. 7. Adsorption isotherm graphs on synthesized adsorbents for (a) Cu, (b) Zn, (c) Ni and (d)
377 Mn (experimental conditions: 0.2 g/L adsorbent dose, pH=5, equilibrium time: 24 h)

378 Table 3. Isotherm parameters for heavy metal adsorption on prepared adsorbents

Equation	Parameter	Sample											
		Cu adsorption			Zn adsorption			Ni adsorption			Mn adsorption		
		SBA-15	SBA-15-NH ₂	Mn-SBA-15-NH ₂	SBA-15	SBA-15-NH ₂	Mn-SBA-15-NH ₂	SBA-15	SBA-15-NH ₂	Mn-SBA-15-NH ₂	SBA-15	SBA-15-NH ₂	Mn-SBA-15-NH ₂
	q_m	0.44	1.96	2.00	0.22	2.49	2.65	0.92	1.53	1.91	0.20	1.42	1.33
Langmuir	b	9.99	3.70	13.55	4.64	0.25	0.33	0.20	0.86	2.12	0.34	0.82	4.60
	R^2	0.99	0.99	0.98	0.94	0.99	0.98	0.97	0.99	0.99	0.91	0.99	0.99
	K_F	10.03	31.40	48.39	4.26	3.29	6.21	0.66	12.66	25.72	0.54	8.41	24.64
Freundlich	n	4.98	4.05	5.24	4.68	1.77	2.05	1.52	3.24	3.95	2.21	2.88	5.41
	R^2	0.99	0.99	0.97	0.99	0.99	0.99	0.96	0.99	0.99	0.92	0.97	0.92

379 3.2.1.2. Multi components adsorption

380 The effect of metal competition and selective Cu extraction was also evaluated in a multi
381 component solution containing heavy metals (Cu, Zn, Ni and Mn) mixed in equal
382 concentration ranging from 10 mg/L to 100 mg/L.

383 Adsorption capacity for competitive heavy metal removal in terms of equilibrium
384 concentration (Fig. 8) showed that the adsorption capacity for Cu on SBA-15-NH₂ (1.01
385 mmol/g) decreased by 54% compared to that of single-component adsorption experiment (Fig.
386 7). This is attributed to competition with other heavy metals (Zn, Ni and Mn). On the other
387 hand, Mn-SBA-15-NH₂ was able to maintain 96% of adsorption capacity for Cu (2.01
388 mmol/g). The adsorption selectivity cannot be determined by single component adsorption
389 capacities for those metal ions (Koong et al., 2013).

390 Generally, the capacity of an adsorbent to maintain high selective ion extraction in a mixed
391 solution is associated to ion characteristics such as ion radius, absolute hardness and absolute
392 electronegativity. In this regard, the ion radius of Cu (0.71 Å), Zn (0.74 Å), Ni (0.69 Å) and
393 Mn (0.80 Å) (Table 4) do not exhibit significant difference, suggesting that ion radius does
394 not play a role towards the selective Cu adsorption capacity exhibited by Mn-SBA-15-NH₂.
395 Likewise, the absolute electronegativity is also similar for Cu²⁺ and Zn²⁺ (28.6 and 28.8) with
396 amine groups (Liu et al., 2008). However, the absolute hardness of Zn (10.8) is higher than
397 that of Cu (8.3). Theoretically, this implies that Mn-SBA-15-NH₂ should have a high
398 adsorption affinity for Zn compared to Cu, because amine is regarded as a harder donor atom
399 (Lee et al., 2016; Yu et al., 2007). Hence, the high selective Cu adsorption exhibited by Mn-
400 SBA-15-NH₂ in mixed metal solution is not entirely related to ion characteristics.

401

402 Table 4. Element characteristics

Metals	Ion radiuses (Å)	Absolute hardness	Absolute electronegativity
Cu	0.71	8.3	28.6
Zn	0.74	10.8	28.8
Ni	0.69	8.5	26.7
Mn	0.80	9.3	24.4

403

404 The affinity of high Cu selectivity by Mn-SBA-15-NH₂ could be related to chelating effect.

405 This is because chemical groups such as amine, amide and aldehyde have different chelating

406 ability with heavy metals. The amine group is regarded as a chemical group that can chelate

407 well with Cu (Koong et al., 2013). Thus, it can be deduced that presence of amine groups on

408 SBA-15 enables to form strong complexation with heavy metals, especially for Cu.

409 Specifically, the manganese loaded and amine-grafted SBA-15 (Mn-SBA-15-NH₂) has an410 enhanced Cu selectivity than amine-grafted SBA-15 (SBA-15-NH₂) due to larger amount of

411 grafted amine groups.

412 The adsorption capacity of the adsorbents used showed an increasing trend as the

413 concentration of heavy metals was increased from 0 – 50 mg/L (Fig. 8). This is most likely

414 due to the presence of higher amounts of other heavy metals (Zn, Ni and Mn) that reduced the

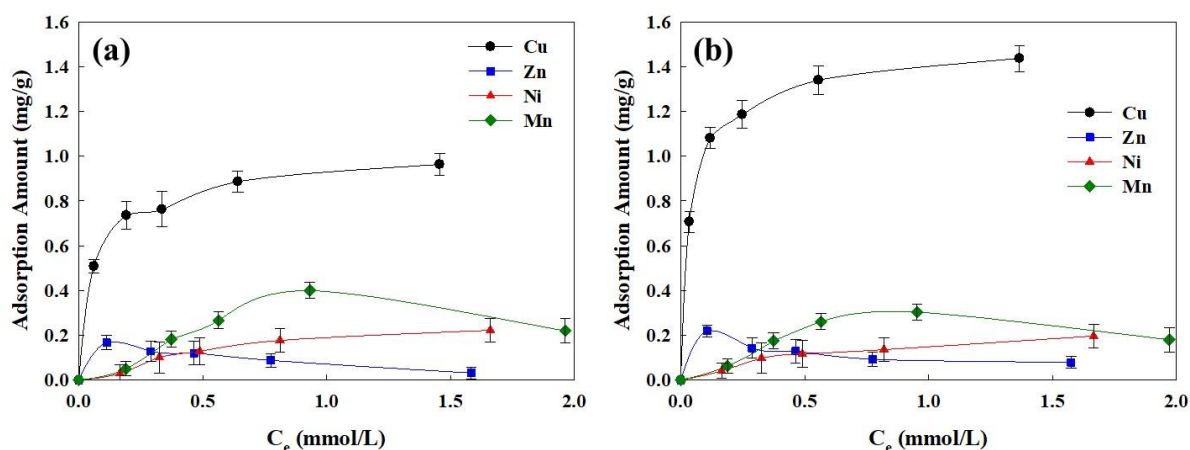
415 adsorption of Cu. However, when the initial concentration of heavy metals was over 50 mg/L,

416 only the adsorption of Cu increased, while the adsorption capacities for Zn, Ni and Mn did

417 not increase any further. In fact, a decrease in adsorption capacities of these heavy metals was

418 observed. Similar trends have been reported in previous studies (Koong et al., 2013; Liu et al.,

419 2008).



420

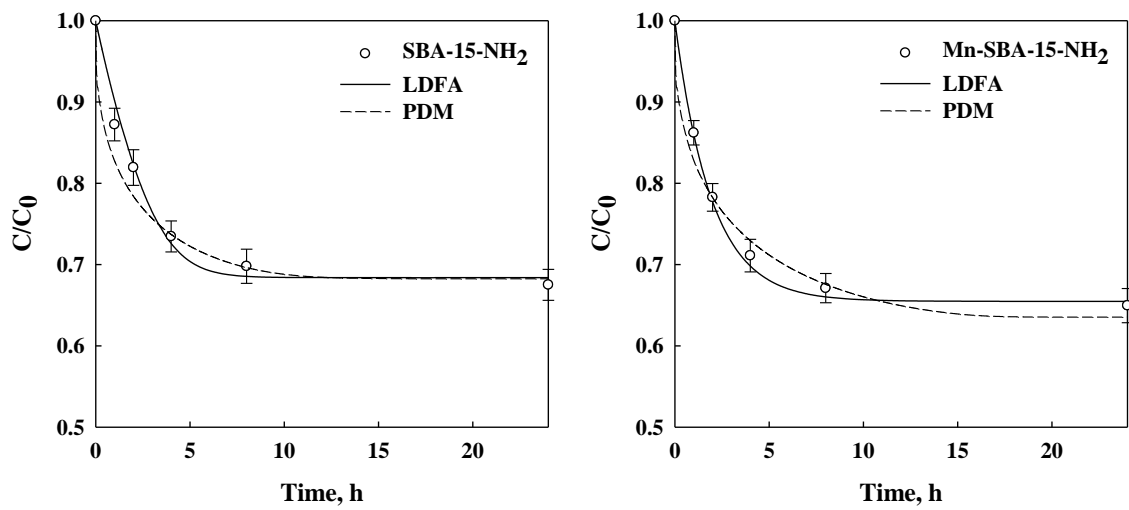
421 Fig. 8. Competitive adsorption of heavy metals on (a) SBA-15-NH₂ and (b) Mn-SBA-15-NH₂
 422 (experimental conditions: 0.01 g of adsorbent dose in 50 ml of heavy metal solution at pH 5,
 423 equilibrium time: 24 h)

424

425 3.2.2. Adsorption kinetics

426 In this work, film mass transfer coefficient (k_f), surface diffusion coefficient (D_s), and pore
 427 diffusion coefficient (D_p) were estimated by LDFA and PDM models to represent the
 428 adsorption kinetics of SBA-15-NH₂ and Mn-SBA-15-NH₂ (Table 5). The kinetic adsorption
 429 phenomenon of the adsorbents was described using ordinary differential equations (ODEs)
 430 and polite dignified effective practical enthusiastic (PDEPE) in MATLAB program,
 431 numerically. Fig. 9 shows the good fitting of experimental values with theatrical values. All
 432 the kinetic coefficients (k_f , D_s and D_p) of SBA-15-NH₂ were observed to be higher than those
 433 of Mn-SBA-15-NH₂. This results indicated higher Cu adsorption rate on the surface of SBA-
 434 15-NH₂ compared to that of Mn-SBA-15-NH₂. The slower adsorption rate on Mn-SBA-15-
 435 NH₂ is attributed to the small pore size upon manganese modification. This was in line with
 436 the pore size evaluation described in Section 3.1.3 (Fig. S2, Table 1). Typically, the smaller

437 pore size of Mn-SBA-15-NH₂ compromised Cu diffusion rate but it increased the Cu
 438 selective adsorption capacity.
 439



440
 441 Fig. 9. Experimental adsorption kinetics of Cu on SBA-15-NH₂ and Mn-SBA-15-NH₂
 442 together with theoretical curves calculated using LDFA and PDM models (experimental
 443 conditions: 0.01 g adsorbent dose in 50 ml of 50 mg/L Cu solution at pH 5)
 444

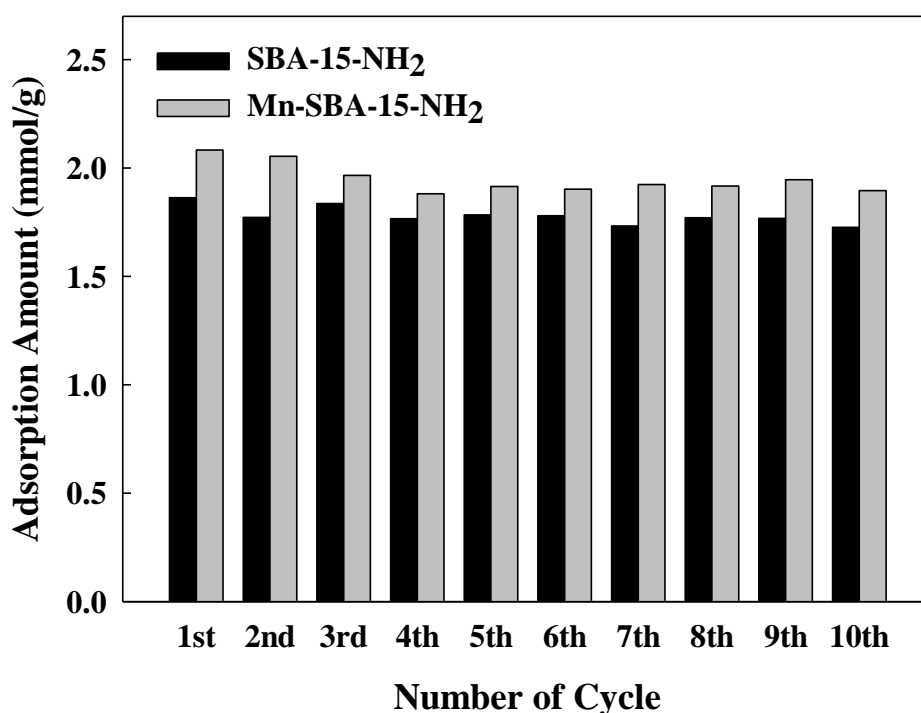
445 Table 5. Adsorption kinetic parameters of SBA-15-NH₂ and Mn-SBA-15-NH₂ (LDFA and
 446 PDM models)

Adsorbents	k_f	D_s	D_p
	m/sec	m ² /sec	m ² /sec
SBA-15-NH ₂	5.35×10^{-5}	2.50×10^{-10}	2.50×10^{-8}
Mn-SBA-15-NH ₂	8.90×10^{-6}	7.60×10^{-12}	1.75×10^{-8}

447
 448 3.2.3. Adsorption and desorption cycle experiment

449 The regenerative capacity of SBA-15-NH₂ and Mn-SBA-15-NH₂ was experimentally

450 evaluated after multiple cycles of adsorption and desorption (Fig. 10). Cu was desorbed from
451 the exhausted adsorbents using acid (HCl) followed by alkaline (NaOH) regeneration. Cu
452 desorption by acid is attributed to the detachment of Cu from amine group in acidic condition.
453 This is reflected by the pH evaluation (Fig. 5) which showed near zero Cu adsorption
454 occurred between pH 1-2. During the 10 times of adsorption and desorption cycle
455 experiments, the adsorption capacity of SBA-15-NH₂ and Mn-SBA-15-NH₂ for Cu were
456 stable with only minimal fluctuations, enabling to maintain more than 90% of the initial
457 adsorption capacity of Cu. The results established the feasibility of high regenerative capacity
458 of both the adsorbents.
459



460
461 Fig. 10. Repeated Cu adsorption on SBA-15-NH₂ and Mn-SBA-15-NH₂ during adsorption-
462 desorption experiment (experimental conditions: 0.2 g/L adsorbent dose in 300 mg/L of Cu

463 solution, pH = 5, equilibrium time: 24 h)

464 4. Conclusion

465 SBA-15, a 2D-hexagonal structured mesoporous silica material, was synthesized with
466 hydrothermal reaction, and subsequently modified by manganese loading and amine grafting,
467 to produce SBA-15-NH₂ and Mn-SBA-15-NH₂. Detailed physical and chemical analyses
468 were made to determine the crystal structure, chemical and surface morphology and surface
469 area and pore size of the synthesized SBA-15-NH₂ and Mn-SBA-15-NH₂. The selective Cu
470 adsorption capacity on SBA-15-NH₂ and Mn-SBA-15-NH₂ was evaluated. The results
471 showed that:

- 472 • The modified adsorbents SBA-15-NH₂ and Mn-SBA-15-NH₂ resulted in 3 – 9 times
473 higher adsorption capacity of heavy metals (Cu, Zn, Ni and Mn) compared to SBA-15
- 474 • Adsorption of heavy metals with both SBA-15-NH₂ and Mn-SBA-15-NH₂ were
475 satisfactorily fitted with Langmuir and Freundlich isotherm equations at optimum pH of
476 5. Mn-SBA-15-NH₂ achieved higher adsorption capacity of Cu (2.08 mmol/g), Zn (1.44
477 mmol/g), Ni (1.84 mmol/g) and Mn (1.28 mmol/g) than SBA-15-NH₂ (1.86, 1.26, 1.29
478 and 1.14 mmol/g).
- 479 • Mn-SBA-15-NH₂ exhibited high selective Cu adsorption (96% of the Cu single solution
480 adsorption capacity) in a mixed metal solution. This is attributed to the presence of
481 higher amine in Mn-SBA-15-NH₂ that formed strong chelating bonding with Cu
482 compared to other metals.
- 483 • Adsorption kinetic data were well fitted with LDFA and PDM models. The slower

484 diffusion rate of Mn-SBA-15-NH₂ compared to SBA-15-NH₂ was due to smaller pore
485 size upon manganese modification.

486 • Multiple cycle of adsorption and desorption experiments established the high
487 regenerative capacity of the modified adsorbent. Both SBA-15-NH₂ and Mn-SBA-15-
488 NH₂ were able to maintain 90% of Cu adsorption capacity even after 10 cycles of
489 adsorption and desorption experiments.

490

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