

23 Abstract

24 Due to the high Brunauer-Emmet-Teller (BET) surface area of zeolitic imidazolate framework (ZIF)-8,
25 secondary crystallization method was used to prepare particle electrode γ -Al₂O₃@ZIF-8. According to the results
26 from field emission scanning electron microscope (SEM) and X-ray diffractometer (XRD), the particle electrode
27 γ -Al₂O₃ was successfully loaded with ZIF-8, and the BET surface area (1433 m²/g) of ZIF-8 was over 10 times
28 than that of γ -Al₂O₃. The key operation parameters of cell voltage, pH, initial RhB concentration and electrolyte
29 concentration were all optimized. The observed rate constant (k_{obs}) of pseudo-first-order kinetic model for the
30 electrocatalytic oxidation (ECO) system with particle electrode γ -Al₂O₃@ZIF-8 ($15.2 \times 10^{-2} \text{ min}^{-1}$) was over 5 times
31 higher than that of the system with traditional particle electrode γ -Al₂O₃ ($2.6 \times 10^{-2} \text{ min}^{-1}$). The loading of ZIF-8 on
32 the surface of γ -Al₂O₃ played an important role in improving electrocatalytic activity for the degradation of
33 Rhodamine B (RhB), and the RhB removal efficiency of the 3D electrocatalytic system with particle electrode of
34 γ -Al₂O₃@ZIF-8 was 93.5% in 15 min, compared with 27.5% in 15 min for particle electrode γ -Al₂O₃. The RhB
35 removal efficiency was kept over 85% after five cycles of reuse for the three-dimensional (3D) electrocatalytic
36 system with particle electrode of γ -Al₂O₃@ZIF-8.

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38 **Keywords:** particle electrode; electrocatalytic oxidation (ECO); zeolitic imidazolate framework (ZIF)-8; RhB

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45 **1. Introduction**

46 Printing and dyeing mill generates a large amount of dye-rich effluents which contain some bio-refractory
47 organic complicated compounds. Even after the biodegradation or chemical flocculation treatment(Holkar *et al.*
48 2016), there still exists some of these substances, such as Rhodamine B (RhB), methyl orange and malachite
49 green(Yu *et al.* 2018). Considering the aesthetic and toxicity problem, it is indispensable to degrade these pigments
50 completely. So far, attempts have been made to degrade and destroy those pigments using diverse advanced
51 oxidation processes (AOPs), including photochemical or electrochemical reactions, ozonation, Fenton oxidation,
52 and catalytic wet oxidation(He *et al.* 2019). In comparison to other AOPs, electrocatalytic oxidation (ECO)
53 technologies have made great progress in wastewater treatment due to its high efficiency, environmental friendly
54 and versatility(Rana *et al.* 2019).

55 The present ECO technologies have developed from the two-dimensional (2D) system to the three-dimensional
56 (3D) electrocatalysis oxidation system, which has become a promising technology and is widely employed in
57 wastewater treatment. In 3D electrocatalysis oxidation systems, particle electrodes can be prepared using
58 mesoporous material, such as granular active carbon (GAC), carbon aerogel (CA), metal particles, modified kaolin,
59 and metal oxides(Zhang *et al.* 2010; Hardjono *et al.* 2011; Chu *et al.*,2016; Zhang *et al.* 2013; Moreira *et al.* 2017;
60 Long *et al.* 2019).

61 Considering the large Brunauer-Emmet-Teller (BET) surface area and electro-adsorption capacity of metal
62 organic frameworks (MOFs), the preparation of particle electrode using MOFs may be an attractive option(Duan *et*
63 *al.* 2016; Ren *et al.* 2017). However, on account of nanostructure of most MOFs materials, it is difficult to prepare
64 the dimensionally stable particle electrodes using MOFs. Thus, as one of the mesoporous materials, the
65 dimensionally stable γ -alumina (γ -Al₂O₃) can be potentially chosen as the carrier of MOFs. Zeolitic imidazolate
66 frameworks (ZIFs) belong to a special subclass of MOFs. In comparison to other MOFs, ZIFs morphology is

67 usually stable in aqueous solution and organic solvents, especially ZIF-8(Wang *et al.* 2008). To the best of our
68 knowledge, the application of MOFs in wastewater treatment was reported mainly on adsorptive removal of methyl
69 orange and methylene blue from aqueous solution (Haque *et al.* 2010). The active roles of ZIF-8 on the enhanced
70 visible photocatalytic activity of Ag/AgCl were also examined for RhB removal, which can be attributed to both
71 adsorption and generation of superoxide radical (Liu *et al.* 2017). An iron terephthalate metal-organic framework
72 MIL-53(Fe) synthesized by a facile solvothermal reaction was capable of activating hydrogen peroxide (H₂O₂) to
73 achieve high efficiency in photocatalytic process (Ai *et al.* 2014). However, there are still some shortcomings
74 limiting their industrial application of the ECO technologies, such as short lifetime of electrode materials and low
75 current efficiency due to some intrinsic drawbacks such as mass transfer limitation, small space-time yield, and low
76 area-volume ratio. Therefore, it is important to develop new particle electrode for improving area-volume ratio and
77 current efficiency.

78 In this study, particle electrode developed from ZIF-8 material was first proposed and the preparation process
79 was optimized. Characterization and performance research of the particle electrode were also carried out. Moreover,
80 this study was committed to supplying a useful base and reference for practical application of 3D electrochemical
81 degradation on dyeing wastewater treatment. RhB was used as the representative target substance to prepare the
82 simulated wastewater to evaluate the electrochemical performance of prepared particle electrodes. The effects of
83 operating parameters, such as cell voltage, pH, initial RhB concentration and electrolyte concentration, on the RhB
84 degradation were all examined. The mechanisms for RhB removal and the reusability of particle electrode were
85 both investigated.

86

87 **2. Experimental section**

88 *2.1. Materials*

89 Zinc nitrate hexahydrate ($\text{Zn}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$), methanol, zinc chloride, sodium formate and sodium sulphate
90 anhydrous were purchased from Guangzhou chemical reagent factory. 2-methylimidazole was purchased from
91 Macklin. All of these reagents were of analytical purity grade. $\gamma\text{-Al}_2\text{O}_3$ (BET surface area: $140 \text{ m}^2/\text{g}$, pore volume:
92 $0.386 \text{ cm}^3/\text{g}$) with the diameters of 3-5 mm was purchased from Dongzi science and technology co., LTD, China.
93 Titanium electrode ($5 \text{ cm} \times 5 \text{ cm} \times 1 \text{ mm}$) and Ti/RuO₂($5 \text{ cm} \times 5 \text{ cm} \times 1 \text{ mm}$) electrode were purchased from Zhao bond
94 metal composite material co., LTD, China.

95 2.2. Preparation of particle electrode $\gamma\text{-Al}_2\text{O}_3@ \text{ZIF-8}$

96 In this study, $\gamma\text{-Al}_2\text{O}_3$ was used as the support for the synthesis of particle electrode. The $\gamma\text{-Al}_2\text{O}_3$ particles
97 were washed with tap-water and deionized water for three times under ultrasonication, and then were baked in the
98 muffle furnace at 823 K for 6 h with 1 k/min heating rate.

99 In a standard synthesis of ZIF-8 crystals (Venna *et al.* 2010), 70mL methanol containing 3.3 g
100 2-methylimidazole was poured into the equal volume solution containing 1.5 g ($\text{Zn}(\text{NO}_3)_2 \cdot 6\text{H}_2\text{O}$). The mixture was
101 stirred for 24 h at room temperature ($298 \pm 1 \text{ K}$). The product was collected by centrifuging, washed with methanol
102 for three times, and then dried at 353 K overnight in a drying oven for further research.

103 The powdered sample ZIF-8 was dissolved in methanol in a certain proportion and the well-distributed
104 suspended solution was obtained under ultrasound for 10 minutes. The dry $\gamma\text{-Al}_2\text{O}_3$ was dipped into the
105 above-mentioned suspended solution for 5 min. After that period, the particles were taken out from the solution and
106 dried at room temperature for 3 hours, then dried at 393 K for 24 h. Thus, the particles loading ZIF-8 seed layer
107 were obtained, used as the support for secondary growth of ZIF-8 membranes (Xu *et al.* 2011).

108 The particles loading ZIF-8 seed layer were tiled at the bottom in a Teflon-lined autoclave, followed by slowly
109 pouring ZIF-8 synthesis solution containing ZnCl_2 , 2-methylimidazole, sodium formate and methanol into the
110 autoclave, and then sealed and heated in a stainless steel autoclave at 393 K for 4 h, then cooled down to room

111 temperature. The resulting particles were washed with methanol for three times and dried at room temperature for
112 24 h, then $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ was obtained.

113 *2.3. Electrocatalytic setup and operation*

114 The 3D-electrochemical degradation of RhB was carried out in a rectangular undivided organic glass cell of
115 585 mL (13cm \times 9cm \times 5cm) capacity. In the system, a Ti/RuO₂ electrode and a Ti electrode were used as the anode
116 and the cathode, respectively. The available working surface of two electrodes was 15cm² (5cm \times 3cm). The anode
117 and cathode were situated vertically and parallel to each other with an inner gap of 4 cm to position a given volume
118 of particle electrodes, which were immersed in the simulated wastewater for 24 h prior to its use. Every particle
119 with high independence would serve as an electrolysis cell in the system, thus the electrodes areas were increased
120 greatly. The particle electrodes were placed in a support netted container in the 3D electrocatalytic system. 450 mL
121 of RhB solution was transferred into the reactor. Na₂SO₄ was added into the RhB solution to enhance the
122 conductivity and adjust the initial solution pH before electro-degradation. A DC power supply (RXN 305D, China)
123 was employed to connect with anode and cathode. The compressed air was sparged into 3D electrode reactor from
124 the bottom at 0.04 m³/h, the compressed air was used to provide oxygen and accelerate mass transfer of RhB.

125 *2.4. Optimization of influence factors for RhB degradation*

126 The effects of cell voltage, pH, initial RhB concentration and electrolyte concentration on the electrocatalytic
127 degradation of RhB were examined, the design of laboratory batch tests was shown in Table 1. Typically, 450 mL
128 of RhB solution was used for tests and all samples were measured in triplicates. To exclude the effect of adsorptive
129 removal on RhB degradation, the 3D electrocatalytic system was immersed in the simulated wastewater for 24 h
130 each test prior to its use. And the RhB samples were collected at every 5 min after the start of the treatment until
131 complete removal of RhB. (The results are shown in Supplementary Material).

132 *2.5. Comparison of electrocatalytic activity for RhB degradation and kinetics analysis*

133 Under the obtained optimum operation parameters, the degradation of RhB was comparatively investigated with
 134 particle electrode of $\gamma\text{-Al}_2\text{O}_3$ and $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$. In order to compare the electrocatalytic activity, the degradation
 135 of RhB was fitted using a pseudo-first-order kinetic model, as described by the following equation (Diao *et al.*
 136 2017):

$$137 \quad \ln(C_t/C_0) = -k_{obs} \cdot t \quad (1)$$

138 where C_t is the concentration of RhB at selected time (mg/L), C_0 is the initial RhB concentration (mg/L), k_{obs} is the
 139 observed rate constant (min^{-1}), and t is time (min).

140 Table 1 Design of the batch tests for study on influence factors of RhB degradation

Group	Influence factors	Other operation parameters
1. Effect of cell voltage/ V	8	Initial pH of 2.0, 20 mg/L RhB, 8 g/L of Na_2SO_4
	10	
	15	
	20	
	25	
2. Effect of pH	2.0	Cell voltage of 20 V, 20 mg/L RhB, 8 g/L of Na_2SO_4
	4.0	
	6.0	
	9.0	
	11.0	
3. Effect of initial RhB concentration/ mg/L	10	Cell voltage of 20 V, initial pH of 2.0, 8 g/L of Na_2SO_4
	20	
	40	
	60	
4. Effect of electrolyte concentration/ g/L	2	Cell voltage of 20 V, initial pH of 2.0, 20 mg/L RhB
	4	
	8	
	16	

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142 2.6. Experiments for mechanisms of RhB degradation

143 In order to demonstrate the mechanisms of RhB degradation with particle electrode of $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$, two
 144 identical electrocatalytic systems were comparatively examined. One was used as control without Tertiary butyl

145 alcohol (TBA) addition, the other was added 1% (v/v) of TBA as a hydroxyl radical ($\cdot\text{OH}$) quencher. The
146 experiment was carried out at the optimum conditions that determined based on the above experiments: cell voltage
147 of 20 V, pH of 2.0, initial RhB concentration of 20 mg/L and electrolyte concentration of 8 g/L. The collection and
148 measurements of samples were same as above-mentioned steps.

149 *2.7. Reuse tests*

150 In order to examine the reusability of novel particle electrode of $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ in a 3D electrocatalytic system
151 for RhB degradation, as the typical steps for preparation of 450 ml of RhB solution and the 3D electrocatalytic
152 system, the optimum operation parameters of cell voltage of 20 V, pH of 2.0, initial RhB concentration of 20 mg/L,
153 8 g/L of Na_2SO_4 were used and the treatment time was 60 min. The RhB degradation tests were carried out
154 repeatedly for 5 times, the collection and measurements of samples were same as above-mentioned steps. (The
155 results are shown in Supplementary Material).

156 *2.8. Analytical methods*

157 The morphology and energy dispersive spectrometer (EDS) of particle electrode were characterized by field
158 emission scanning electron microscope (SEM, ZEISS Ultra 55, Germany, Carl Zeiss). BET surface area and
159 micropore distribution of ZIF-8 were measured by N_2 adsorption at 77 K on an ASAP2020 equipment. X-ray
160 diffractometer (XRD, Rigaku MiniFlex 600) was employed to analyze the crystal structures of the particle
161 electrodes. Fourier transform infrared (FTIR) spectra of the particle electrodes were investigated with fourier
162 transform infrared spectrometer (FTIR Spectrometer, Nicolet 6700, Thermo Nicolet). Before measurement, the
163 particle electrodes were dehydrated under vacuum at 393 K for 12 h. The RhB concentration was measured with
164 UV-vis spectrometer (UV6000PC, China) at its maximum absorption wavelength of 554 nm. The initial pH of the
165 simulated wastewater was measured with a DZS-706 multi-parameter water quality meter (Rex electric Chemical,
166 Shanghai, China).

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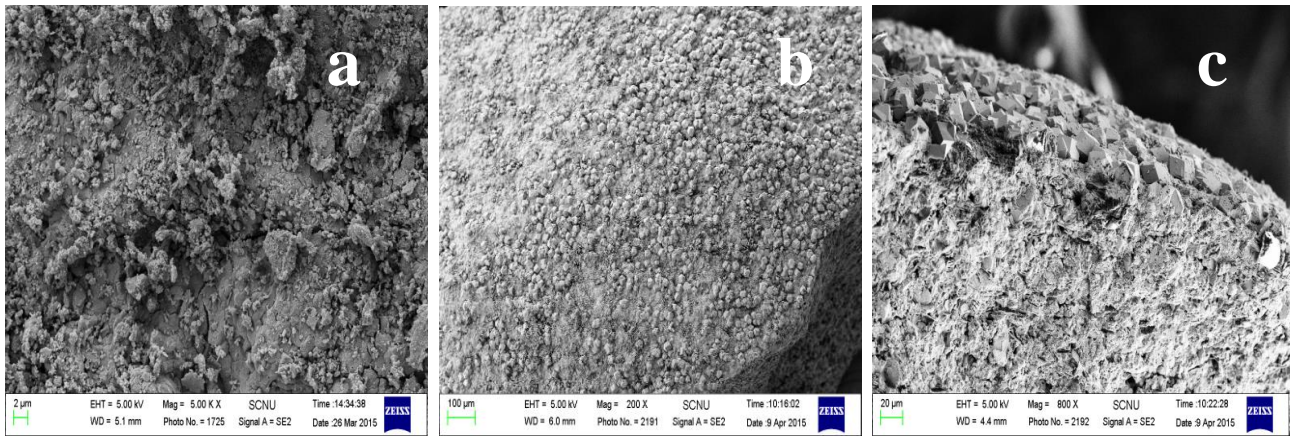
168 **3. Results and discussion**

169 *3.1. Characterization*

170 The SEM images of the particle electrode are shown in Fig. 1. Apparently, no matter from the top view or the
171 cross-section view, compact hexahedral crystals could be observed on the outer surfaces of the γ -Al₂O₃@ZIF-8.
172 The images also display that the average crystal sizes and the film thickness are about 15 μ m and 30 μ m,
173 respectively. That indicated a large quantity of ZIF-8 crystals had been growing on the surface of the γ -Al₂O₃
174 successfully. Fig. 2 shows N₂ adsorption-desorption isotherms (a) and micropore distribution (b) of ZIF-8. On
175 account of the nano-structure and high BET surface area of ZIF-8 (1433 m²/g) compared with that of γ -Al₂O₃ (140
176 m²/g), the adsorption capacity of the γ -Al₂O₃@ZIF-8 was increased, correspondingly, the performance of particle
177 electrode might be improved.

178 It is well known that as one of MOFs, ZIF-8 has the property of permanent microporosity and tremendous BET
179 specific surface areas. Considering the theory of carbon fiber felt (ACF) used as electrodes in electro-catalysis
180 (Huang & Su 2010), large BET specific surface areas and electrical conductivity could improve the efficiency and
181 reproducibility.

182 The impregnation process is the key to obtain γ -Al₂O₃@ZIF-8 in our synthesis method. To ensure that the ZIF-8
183 crystals have been successfully loaded on the γ -Al₂O₃ particle, XRD and IR were employed to analyze the
184 structures and functional groups of the γ -Al₂O₃ and γ -Al₂O₃@ZIF-8. Fig. 3 and Fig. 4 display XRD patterns and IR
185 spectrograms of the particles before and after loading, respectively. Both of the spectrograms of γ -Al₂O₃@ZIF-8
186 seem to be homologous with γ -Al₂O₃ substantially and yet appear some main characteristic peaks of ZIF-8, which
187 were pointed out in the figures. The diffraction peaks of γ -Al₂O₃@ZIF-8 are relatively weak, which results from the
188 relatively lower content of ZIF-8. Another possible reason is weak crystalline during secondary growth.

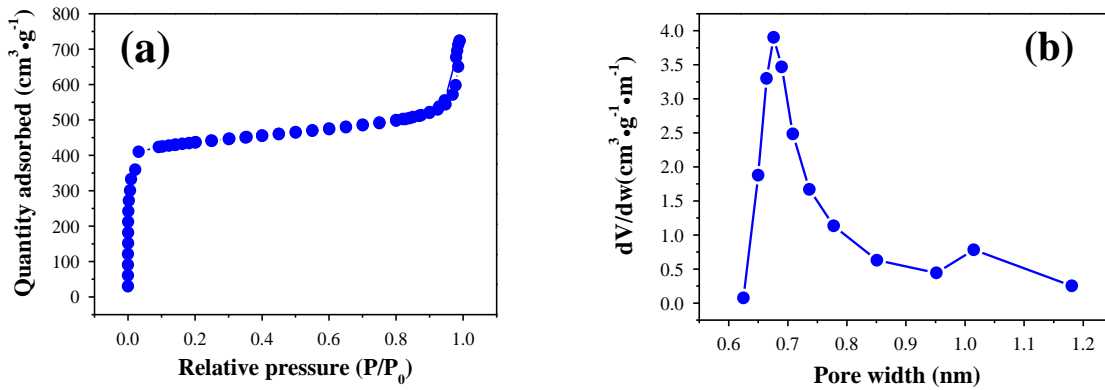


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190 (a) Top view of the γ -Al₂O₃ (b) Top view of the γ -Al₂O₃@ZIF-8 (c) Cross-section of the γ -Al₂O₃@ZIF-8

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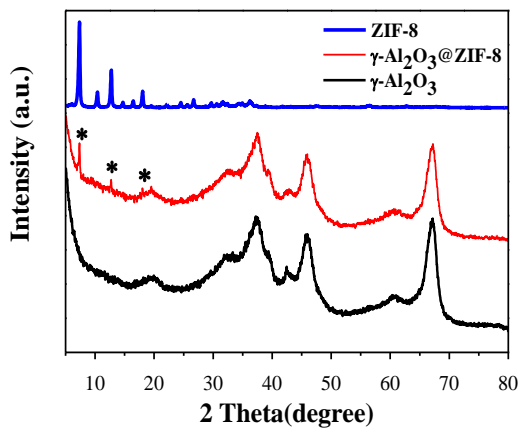
Fig. 1 SEM images of the particles



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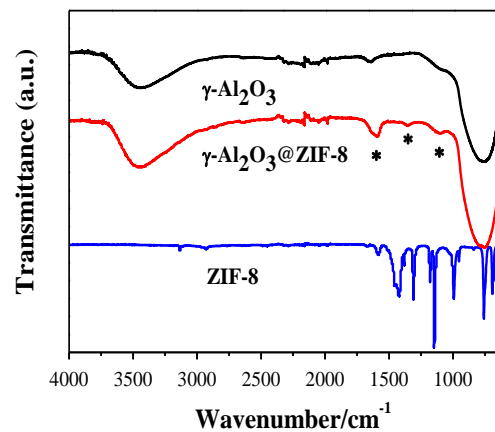
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Fig. 2 N₂ adsorption-desorption isotherms (a) and micropore distribution (b) of ZIF-8



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195 Fig. 3 X-ray diffraction patterns of the particles

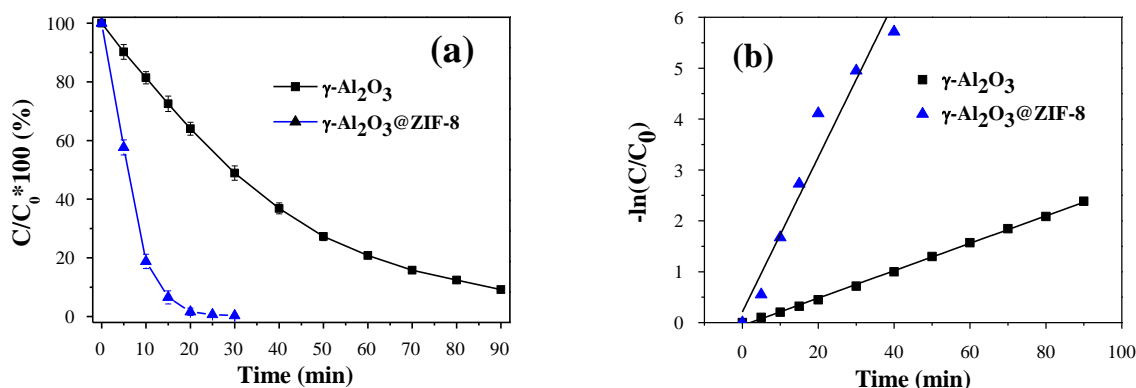


196 Fig. 4 IR spectra of the particles

197 3.2. Comparison of RhB degradation with $\gamma\text{-Al}_2\text{O}_3$, $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ and kinetics analysis

198 The degradation efficiency of RhB with the prepared particle electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ was investigated,
199 compared with traditional particle electrode $\gamma\text{-Al}_2\text{O}_3$. As shown in Fig. 5(a), the particle electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$
200 decolorized RhB more effectively than electrode $\gamma\text{-Al}_2\text{O}_3$. The RhB removal rate for the system with particle
201 electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ reached over 93.5% in 15 min. In contrast, the removal rate of RhB for the
202 electrocatalytic system with particle electrode $\gamma\text{-Al}_2\text{O}_3$ only reached 27.5% in 15 min and reached 90% after the
203 treatment for 90 min. Ai et al. employed an iron terephthalate metal-organic framework MIL-53(Fe) to achieve
204 high efficiency in photocatalytic degradation of RhB, with the presence of hydrogen peroxide (H_2O_2), under visible
205 light irradiation within 50 min, and found that the removal efficiency of RhB was less than 85% in 20 min(Ai *et al.*
206 2014). Liu et al. prepared a novel Ag/AgCl/ZIF-8(50%) composite photocatalyst for the removal of RhB from
207 aqueous solution, and found that the removal efficiency of RhB was less than 80% in 30 min with the synthesized
208 Ag/AgCl/ZIF-8(50%). It was also found that both the adsorption ability of ZIF-8 and the ability for the formation
209 of $\text{O}_2^{\cdot -}$ contributed to the degradation of RhB (Liu *et al.* 2017). Compared with the previous studies, the higher
210 RhB removal efficiency was obtained at shorter reaction time for the 3D electrocatalytic system with particle
211 electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ addition.

212 As shown in Fig. 5(b), the kinetics of electrocatalytic degradation of RhB were pseudo-first order, and the
213 observed rate constant (k_{obs}) of the ECO system with particle electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ ($15.2 \times 10^{-2} \text{ min}^{-1}$) was over
214 5 times higher than that of the system with traditional particle electrode $\gamma\text{-Al}_2\text{O}_3$ ($2.6 \times 10^{-2} \text{ min}^{-1}$), which suggested
215 that the loading of ZIF-8 on the surface of particle electrode $\gamma\text{-Al}_2\text{O}_3$ enhanced greatly the electrocatalytic activity
216 of 3D ECO system.



217

218 Fig. 5 Time-course variation with different particle electrodes (a) C/C_0 of RhB solution and (b) Pseudo-first order

219 kinetics for the degradation of RhB

220

221 3.3. Mechanisms for RhB degradation with particle electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$

222 As shown in Fig. 6, the absorption spectrum of RhB was characterized by its maximum absorbance at 554nm

223 and the absorbance peaks declined obviously with prolonged reaction time due to electrochemical degradation.

224 Natarajan et al. (2011) also reported the study on UV-LED/ TiO_2 process for degradation of RhB, and found that the

225 formed intermediate compounds were similar to previously reported literature on the degradation of RhB, the

226 formed oxidized products were mineralized into CO_2 , H_2O , NO_3^- and NH_4^+ . It was inferred that the degradation of

227 RhB by the photogenerated active species such as $\cdot\text{OH}$ and hole could attack the central carbon of RhB to

228 decolorize the dye and further degraded via N-de-ethylation process. The adsorptive removal of RhB was excluded

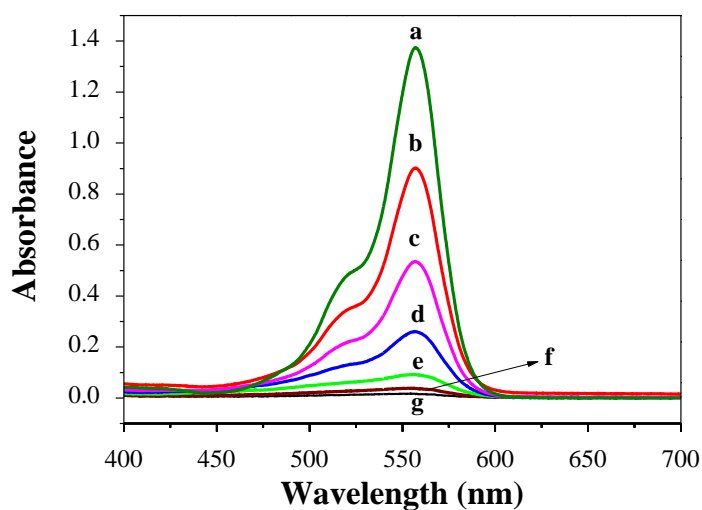
229 through immersing particle electrode for 24h prior to its use. However, the direct quantitative comparison of RhB

230 removal between electrocatalytic oxidation and the available technologies are difficult since lots of factors (e.g.

231 initial RhB concentration, operating conditions, among others) could affect the results. The higher RhB removal

232 efficiency of 3D electrocatalytic system with particle electrode of $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$ is favorable for increasing

233 treatment capacity for dyeing wastewater.



234

235 Fig. 6 UV-vis spectra of RhB with γ -Al₂O₃@ZIF-8 electrodes at various treating time: (a) 0 min; (b) 5 min; (c) 10
 236 min; (d) 15 min; (e) 20 min; (f) 25 min; (g) 30 min. Cell voltage of 20 V, pH of 2.0, initial RhB concentration of 20
 237 mg/L, Na₂SO₄ of 8 g/L as supporting electrolyte.

238

239 The TBA is a commonly used \cdot OH radical scavenger and reacts with hydroxyl radicals with a rate constant of
 240 $6 \times 10^8 \text{ M}^{-1} \cdot \text{s}^{-1}$ (Ikhlaiq et al., 2012). As shown in Fig. 7, the experimental results showed that the RhB degradation
 241 efficiency for the system with the addition of TBA was obviously lower than that with no TBA addition, which also
 242 suggested that hydroxyl radical played an important role in the electrocatalytic degradation of RhB. It could be
 243 attributed to the disintegration of the azo linkage (conjugated xanthene ring), which acted as the chromophore of
 244 RhB (Yan *et al.* 2011). This is probably because \cdot OH radicals attacked the N=N bond of the azo dye, which maybe
 245 the most active site. According to the SEM (Fig. 1) and XRD (Fig. 3) results, ZIF-8 was successfully supported on
 246 the outer surface of γ -Al₂O₃.

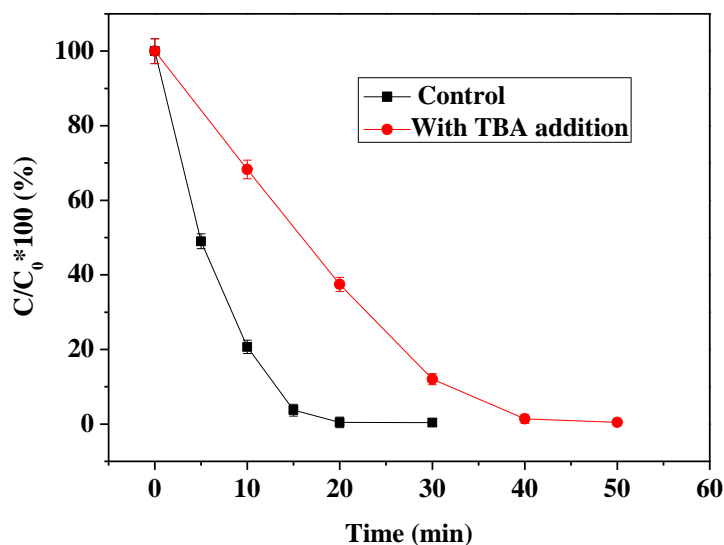


Fig. 7 Effect of 1% (v/v) TBA addition on RhB removal

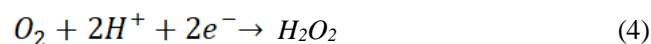
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250 Fig. 8 shows the mechanisms of hydroxyl radical generation in 3D electrocatalytic system with particle
 251 electrode $\gamma\text{-Al}_2\text{O}_3\text{@ZIF-8}$. There were two pathways to generate hydroxyl radical: on the anode and particle
 252 electrode. At an appropriate voltage, the particles were polarized to form lots of charged microelectrodes with one
 253 surface as anode while the other was charged the opposite. The generation of hydroxyl radical on the surface of
 254 particle electrode can be expressed as Formula 4 and Formula 5 (Zhang *et al.* 2013). On the other hand, the BET
 255 surface area of ZIF-8 was $1433\text{ m}^2/\text{g}$ and was much higher than that of the carrier $\gamma\text{-Al}_2\text{O}_3$, which could enhance
 256 greatly the adsorption capacity for RhB and accelerate the mass transfer rate of RhB on liquid-solid interface,
 257 resulting in high efficiency of RhB degradation.

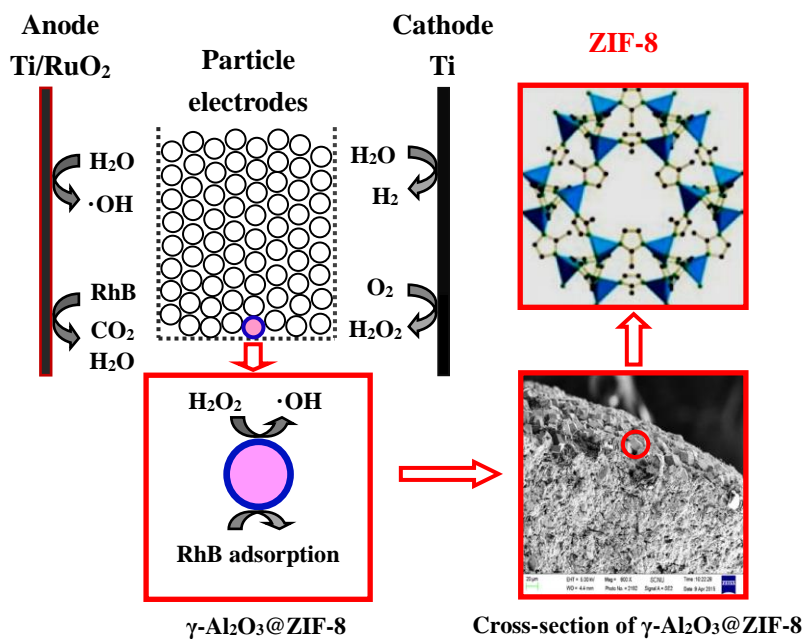
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262 Fig. 8 Mechanisms of hydroxyl radical generation in 3D electrocatalytic system with particle electrode

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γ-Al₂O₃@ZIF-8

264

265 4. Conclusions

266 In summary, we demonstrated that the novel particle electrode of γ-Al₂O₃@ZIF-8 showed good electrocatalytic

267 activity and reusability for RhB degradation in a 3D electrochemical system. The electrocatalytic activities

268 depended greatly on the various operating parameters. The results on hydroxyl radical quencher by TBA showed

269 that the loading of ZIF-8 on the surface of γ-Al₂O₃ played an important role in improving electrocatalytic activity

270 for RhB degradation, and the RhB removal efficiency of the electrocatalytic system with particle electrode of

271 γ-Al₂O₃@ZIF-8 was 93.5% in 15 min. The high RhB removal efficiency and reusability of particle electrode

272 γ-Al₂O₃@ZIF-8 in the ECO system make it become a promising treatment technology for dye wastewater.

273

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