

Local scale thermal environment and limited gene flow indicate vulnerability at warm-edge population of a habitat-forming macroalga

Jennifer S. Clark^{1, 2*}, Alistair G. Poore³, Melinda A. Coleman⁴, Martina A. Doblin⁵

¹University of British Columbia, Canada, ²Climate Change Cluster, University of Technology Sydney, Australia, ³University of New South Wales, Australia, ⁴National Marine Science Centre, School of Environment, Science and Engineering, Southern Cross University, Australia, ⁵University of Technology Sydney, Australia

Submitted to Journal:
Frontiers in Marine Science

Specialty Section:
Global Change and the Future Ocean

Article type:
Original Research Article

Manuscript ID:
545122

Received on:
31 Mar 2020

Revised on:
07 Jul 2020

Frontiers website link:
www.frontiersin.org

Conflict of interest statement

The authors declare that the research was conducted in the absence of any commercial or financial relationships that could be construed as a potential conflict of interest

Author contribution statement

JC conceived the idea for the manuscript, collected the data, curated and formally analysed the data and wrote the original manuscript. MD, AP, MC supervised the project, reviewed, wrote and edited the manuscript, MD collected some of the data and provided resources for funding the project.

Keywords

Thermal performance, Climate Change, Photosynthesis, microsatellites, Genetic diversity, quantitative breeding designs, Seaweed, *Hormosira banksii*

Abstract

Word count: 299

Species inhabiting warm-edge populations of their distribution are suggested to be at the forefront of global warming due to reduced fitness, limited gene flow and living close to their physiological thermal limits. Determining the scale that governs thermal niche and the functional responses of habitat-forming species to environmental stressors is critical for successful conservation efforts, particularly as coastal ecosystems are impacted by global change. Here, we examine the susceptibility of warm-edge populations to warming, in the habitat-forming macroalga, *Hormosira banksii*, from south-eastern Australia. We use a quantitative breeding design to quantify intraspecific variation in thermal performance (growth, ontogenic development and photosynthetic efficiency) of different genotypes sourced from sites at the equatorward distributional edge (warm-edge) and those towards the center of its distribution (non-edge). The genetic diversity and structure of *H. banksii* was also examined using microsatellite markers amongst the same sites. Our results found contrasting thermal performance in growth and development which depended on local scale thermal environment rather than distribution origin. Contrarily, warm-edge germlings grew optimally in lower temperatures and had narrower thermal breadth compared to non-edge germlings. Warm-edge germlings however, showed greater plasticity to tolerate high light indicated by a greater proportion of energy being dissipated as regulated nonphotochemical quenching (Y(NPQ)) than nonregulated nonphotochemical quenching (Y(NO)). Overall genetic diversity was lower at the warm-edge sites with evidence of increased structuring and reduced gene flow in comparison to the non-edge location. Evidence of genetic structuring was not found locally between high and low shore within sites. Together, these data suggest that non-edge populations may be “thermally buffered” from increased temperatures associated with ocean warming. Warm-edge populations of *H. banksii*, however, may be vulnerable to warming, due to narrower thermal breadth and sensitivity to higher temperatures, with genetic impoverishment through loss of individuals likely to further reduce population viability.

Contribution to the field

The resilience of species to persist in global warming depends on the range of functional responses produced by genotypes and phenotypes within its distribution. Previous research has identified warm-edge populations to be thermally tolerant but genetically impoverished making them more susceptible to environmental change due to the reduced range of functional responses. In marine macroalgae, there is limited research that pairs thermal tipping points with genetic diversity and determining the spatial scale in which global warming will have the greatest effect on physiology. Here, we conducted thermal performance curves using germlings from the habitat-forming, intertidal macroalga *Hormosira banksii*, from warm-edge and non-edge populations and multiple spatial scales (regional, local and individual) to test whether warm-edge populations are susceptible to future warming and whether this was related to population genetic structure. Contrarily, warm-edge populations presented greater thermal sensitivities and narrower thermal breadth to increased temperatures. We propose that thermal history and origin of the individual governs thermal niche, however, warm-edge populations may be facilitating their resilience to short-term environmental stress by actively dissipating a greater proportion of excess energy away from photosystems. Lower genetic diversity and gene flow at warm-edges, however, may constrain responses to environmental change over the longer term.

Funding statement

Australian postgraduate award awarded to JC

Ethics statements

Studies involving animal subjects

Generated Statement: No animal studies are presented in this manuscript.

Studies involving human subjects

Generated Statement: No human studies are presented in this manuscript.

Inclusion of identifiable human data

Generated Statement: No potentially identifiable human images or data is presented in this study.

In review

Data availability statement

Generated Statement: The raw data supporting the conclusions of this article will be made available by the authors, without undue reservation.

In review

Local scale thermal environment and limited gene flow indicate vulnerability at warm-edge population of a habitat-forming macroalga

Jennifer. S. Clark^{1,2*}, Alistair. G.B. Poore³, Melinda A. Coleman⁴, Martina A. Doblin²

¹ University of British Columbia, Department of Botany, Vancouver, BC, Canada

² University of Technology Sydney, Climate Change Cluster, Broadway, NSW, Australia

³ University of New South Wales, School of Biological Earth and Environmental Sciences, Kensington, NSW, Australia

⁴ NSW Fisheries, National Marine Science Center, Coffs Harbour, NSW, Australia

* Correspondence:

Dr Jennifer Clark

Jennifer.Clark@botany.ubc.ca

Keywords: Thermal performance, climate change, photosynthesis, microsatellites, genetic diversity, quantitative breeding designs, seaweed, *Hormosira banksii*

Number of words: 9220

Number of figures: 5

Abstract

Species inhabiting warm-edge populations of their distribution are suggested to be at the forefront of global warming due to reduced fitness, limited gene flow and living close to their physiological thermal limits. Determining the scale that governs thermal niche and the functional responses of habitat-forming species to environmental stressors is critical for successful conservation efforts, particularly as coastal ecosystems are impacted by global change. Here, we examine the susceptibility of warm-edge populations to warming, in the habitat-forming macroalga, *Hormosira banksii*, from south-eastern Australia. We use a quantitative breeding design to quantify intraspecific variation in thermal performance (growth, ontogenic development and photosynthetic efficiency) of different genotypes sourced from sites at the equatorward distributional edge (warm-edge) and those towards the center of its distribution (non-edge). The genetic diversity and structure of *H. banksii* was also examined using microsatellite markers amongst the same sites. Our results found contrasting thermal performance in growth and development which depended on local scale thermal environment rather than distribution origin. Contrarily, warm-edge germlings grew optimally in lower temperatures and had narrower thermal breadth compared to non-edge germlings. Warm-edge germlings however, showed greater plasticity to tolerate high light indicated by a greater proportion of energy being dissipated as regulated nonphotochemical quenching (Y(NPQ)) than nonregulated nonphotochemical quenching (Y(NO)). Overall genetic diversity was lower at the warm-edge sites with evidence of increased structuring and reduced gene flow in comparison to the non-edge location. Evidence of genetic structuring was not found locally between high and low shore within sites. Together, these data suggest that non-edge populations may be “thermally buffered” from increased temperatures associated with ocean warming. Warm-edge populations of *H. banksii*, however, may be vulnerable to warming, due to narrower thermal breadth and sensitivity to higher temperatures, with genetic impoverishment through loss of individuals likely to further reduce population viability.

46 **1. Introduction**

47 Anthropogenic mediated climate change is already having profound impacts on the physiology and
48 distribution of many species worldwide (Pecl et al., 2017; IPCC 2018). By the end of this century,
49 anthropogenic increases in atmospheric greenhouse gases will have increased ocean and air
50 temperatures by 1.5 - 2 °C with global mean surface temperatures already warmed by 0.87 °C during
51 the decade 2006 – 2015 (IPCC 2018). In the last decade, the prevalence of extreme climate events
52 (heatwaves, droughts, floods, cold spells and storms) has caused further loss of populations and
53 poleward shifts in distribution as species are being pushed past their physiological thresholds (Hawkins
54 et al., 2009; Burrows et al., 2014; Poloczanska et al., 2016; Smale et al., 2019). Persistence in the face
55 of a warming climate will require physiological plasticity and adequate genetic diversity for natural
56 selection to act upon (Sgrò and Hoffmann, 2004; Reusch et al., 2005; Hoffmann and Sgrò, 2011;
57 Wernberg et al., 2018; Gurgel et al., 2020).

58
59 Understanding how global warming and future extreme climate events will impact species requires
60 knowledge of species' thermal niche and underlying genetic diversity across its distribution. Thermal
61 niche is developed through acclimation and adaptation to temperatures experienced throughout a
62 species life history which can vary with space and time (for a review see Bennett et al., 2015). For
63 instance, thermal limits will differ for central and marginal populations as thermal regimes vary across
64 a species geographical range (Sunday et al., 2012; Bennett et al., 2019). Thermal breadth is also
65 influenced by the range of temperatures experienced throughout a species life history (Sunday et al.,
66 2012). The climate variability hypothesis suggests that a positive relationship exists between thermal
67 breadth of an organism and climate variability with increasing latitude. Therefore, populations in
68 higher latitudes will have a greater thermal breadth as individuals experience a greater range in
69 temperatures, than those closer to the equator (Stevens, 1989). Thus, cooler populations are suggested
70 to be more resistant to warming as they have a broader thermal breadth compared to warmer
71 populations (however see Bennett et al., 2015). As thermal limits often govern species range
72 boundaries, and individuals are pushed beyond their physiological limits, the fitness of individuals
73 therefore diminish towards distributional limits (Sagarin and Gaines, 2002; Thomas et al., 2004;
74 Hampe and Petit, 2005; Pearson et al., 2009).

75
76 Reduced fitness at range margins often coincides with reduced gene flow and connectivity, habitat
77 fragmentation (local separation of populations) and reduced effective population sizes. These can result
78 in decreased genetic diversity (the range of functional responses provided by genotypes and
79 phenotypes) and increase genetic differentiation between populations (Hampe and Petit, 2005; Eckert
80 et al., 2008; Coleman et al., 2011a; Wernberg et al., 2018). The decrease in genetic diversity towards
81 range limits has been documented extensively in plants and animals (Hampe and Petit, 2005; for a
82 review see Eckert et al., 2008). Such patterns may also exist at the local scale over strong, but small-
83 scale environmental gradients, where individuals live in habitat mosaics or where they may be spatially
84 or temporally segregated. (Helmuth et al., 2006; Harley, 2008). Species inhabiting edge distributions
85 are at the forefront of climate change, as they are already restricted by environmental factors and living
86 close to their physiological limits (Parmesan, 2006; Smale and Wernberg, 2013; Pecl et al., 2017).
87 Without adequate genetic diversity, potential selection for tolerant genotypes may be limited.

88
89 Investigating the spatial scale at which variation in environmental stressors will most strongly influence
90 fitness is needed to determine impacts on biological systems (Helmuth et al., 2014). Species'
91 distributions can often span thousands of kilometres and individuals can be exposed to wide variation
92 in environmental regimes at different spatial scales; regionally among latitudes, and locally, among
93 habitats within a single location. Factors such as local climate and topography within a habitat can
94 translate to mosaics of "hotspots" and "coldspots" (Helmuth et al., 2006). For sessile organisms,

95 morphological differences among individuals add to local habitat topography and daily fluctuations in
96 exposure, all of which can shape differing physiological thresholds (Helmuth and Hofmann, 2001;
97 Harley, 2008; Clark et al., 2018). Studies have also suggested that variation at the scale of an individual
98 can have a greater effect on physiology than broad scale differences observed over kilometres (Helmuth
99 et al., 2002; Helmuth, 2009). Consequently, species declines in response to increasing environmental
100 stress may not occur evenly across their range (Helmuth et al., 2006; Pearson et al., 2009; Miller et al.,
101 2019), yet understanding intraspecific variation in tolerance is required to predict future species
102 distributions.

103
104 The rocky intertidal has been suggested to be a sentinel habitat for global warming, primarily due to
105 resident organisms being already close to their thermal limits (Stillman and Somero, 2000; Somero,
106 2005). Habitat-forming macroalgae are particularly important primary producers in this habitat due to
107 their role as ecosystem engineers through modifying local environmental conditions and providing
108 resources (Dayton, 1972; Jones et al., 1994) that can strongly facilitate associated biodiversity (Schiel,
109 2006; Bishop et al., 2009). As macroalgae are sessile organisms, they cannot move to avoid heat stress
110 so must physiologically tolerate, adapt or perish in the face of climate change. Studies on the effects
111 of temperature in governing species distribution and warm-edge ranges are becoming more apparent
112 in macroalgal dominated communities (Pearson et al., 2009; Martínez et al., 2012; Ferreira et al., 2014;
113 Bennett et al., 2015; Mota et al., 2018; King et al., 2019). Temperature is a fundamental determinant
114 of algal fitness as it regulates photosynthesis as well as enzymes that govern metabolic activity
115 (Allakhverdiev et al., 2008; Falkowski and Raven, 2013). Due to this, photosynthetic health has been
116 widely used to assess thermal tolerance in photosynthetic organisms including many plants and
117 macroalgae (Pearson et al., 2009; Smolina et al., 2016; Wernberg et al., 2016). Photosynthetic
118 performance, in turn, can be evaluated by chlorophyll *a* fluorescence measurements, specifically the
119 maximum quantum yield (F_v/F_m) and investigation of energy dissipation pathways (Genty et al., 1989;
120 Kramer et al., 2004; Schreiber, 2004).

121
122 Variation in thermal response among marine macroalgae has been mostly studied in the context of
123 determining how lethal temperatures set distributional limits both across species ranges and vertically
124 on the shore (Schonbeck and Norton, 1978; Hartnoll and Hawkins, 1985; Davison and Pearson, 1996).
125 On regional scales, individuals inhabiting warm range-edge populations closer to the equator have been
126 shown to have greater thermal tolerances due to exposure to higher temperatures throughout their life
127 history (Mota et al., 2018). Photosynthetic health of macroalgae at warm-range limits also reflect
128 greater thermal tolerances indicating greater ability of warm-edge thalli to maintain maximum quantum
129 yield of PSII (F_v/F_m) in higher temperatures than in cool-edge populations (Mota et al., 2018). Warm-
130 edge populations are often fragmented and reduced in size (Coleman et al., 2011b, 2011a; Zardi et al.,
131 2015), suggesting that these populations are physiologically stressed towards range limits (Araújo et
132 al., 2011) and may be less resilient to prolonged exposure to extreme climate events (Wernberg et al.,
133 2016; Mota et al., 2018). Variation in thermal response on the local scale, among vertical heights on
134 the shore, is well known and is related to daily tidal regimes and topography, and the ability to
135 effectively photosynthesize during periods of increased desiccation and thermal stress (Schonbeck and
136 Norton, 1978; Dring and Brown, 1982; Davison and Pearson, 1996; Williams and Dethier, 2005). Less
137 well understood is the contribution of heritable genetic variation in thermal tolerance, required if there
138 is to be any local adaptation to a given thermal regime or evolution of increased tolerance with
139 increasing temperatures. Relatively few studies have experimentally identified heritable genetic
140 variation in seaweeds with families showing the potential for adaptation to changes in environment
141 associated with climate change (Clark et al., 2013; Al-Janabi et al., 2019; Mabin et al., 2019).

143 Southeastern Australia has been identified as a climate change hotspot (Lough and Hobday, 2011;
144 Hobday and Pecl, 2014) and is therefore an ideal location to study the effects of global warming on the
145 genetic diversity and physiology of marine macroalgae. The east coast of Australia follows a north
146 (equatorward) to south (poleward) thermal gradient with a natural warm to cold water transition, and
147 serves as the main conduit for gene flow for many sessile marine species within coastal ecosystems
148 (Coleman et al., 2011b). *Hormosira banksii*, is a dominant, habitat-forming macroalga found on rocky
149 shores in Australia and New Zealand and spans >3000 km of coastline from the northern distributional
150 edge Skennars Head in New South Wales to Albany in Western Australia (Womersley, 1987; Huisman,
151 2019) as well Tasmania, and the North and South Island of New Zealand (Nelson, 2013). Towards the
152 equatorward distributional limits, percent cover decreases from 80% at Minnie Water to 20% at
153 Angourie and 25% at Skennars Head, with *H. banksii* mostly found in rockpools at the distributional
154 limits (personal observation). It is dioecious and has a monophasic life cycle with oogonium
155 development found in mature conceptacles in every season, producing gametes potentially every low
156 tide (Osborn, 1948). Further, gamete dispersal is less than 10 m (Bellgrove et al., 1997), and has been
157 documented to have limited gene flow (Coleman et al., 2011a, 2019) with local adaptation recently
158 identified in thermally different regions on the south coast of Australia (Miller et al., 2019). Previous
159 research on thermal performance of *H. banksii* is limited, but has identified increased thermal
160 sensitivity through decreased photosynthetic yield of PSII, smaller morphology and lower percent
161 coverage of thalli (~20%) in adult thalli of warm-edge populations of *H. banksii* (Clark et al., 2018).
162 Less is known about how populations of *H. banksii* within the equatorward range edge of this species
163 will respond to elevated temperatures and extreme climate events, given limited gene flow and
164 inhabiting physiological stressful environments.

165
166 In this study, multiple performance traits of the ecologically and functionally important intertidal
167 macroalga, *Hormosira banksii*, were quantified to assess variation in thermal tolerance and genetic
168 structuring at three nested spatial scales: among locations, among heights on the shore and among
169 individual genotypes within a shore, to determine the vulnerability of warm-edge populations to
170 climate warming. We aimed to (1) assess the thermal performance of individuals (2) determine
171 whether genetic variation in traits are heritable, and (3) determine the genetic diversity and connectivity
172 of warm-edge versus non-edge populations of *H. banksii* to assess the role of thermal history and
173 adaptation in providing resilience to global warming.

174 175 **2. Methods**

176 **2.1. Study locations and collection of model organism**

177 For genetic diversity analysis, *H. banksii* populations were sampled during the austral autumn (April –
178 May 2014) in two eastern Australian regions: a warm-edge region encompassing two locations,
179 Angourie (29°28'41.51" S, 153° 21' 49.53" E) and Minnie Water (29°46'34.23" S, 153°18'07.43" E)
180 at the northern warm-edge of its distribution, and a non-edge region within the center of its range,
181 encompassing two locations, Pearl Beach (33°32'57.70" S, 151°18'32.36" E) and Bilgola Beach
182 (33°38'54.48" S, 151°19'39.59" E), approximately 460 km further south (Fig.1). For thermal response
183 experiments, *H. banksii* was collected from two populations each within the mentioned regions (Minnie
184 Water and Pearl Beach). At the time of sampling, *H. banksii* was the dominant, intertidal macroalgal
185 species (60-80% percent cover) in all locations except Angourie which had 20% percent cover.

186 187 **2.2. Thermal exposure at different spatial scales**

188 Data from multiple sources was used to demonstrate average air and sea surface temperatures (SST)
189 experienced at each warm-edge (Minnie Waters) and non-edge (Pearl Beach) location (Fig S2). SST
190 detected by satellite (MODIS-Aqua), was obtained via GIOVANNI (NASA GES DISC) using a 4 km²
191 area (Minnie Water -29.5 °S, 29.0 °S 153.4 °E 153.9 °E; Pearl Beach -33.844 °S, -33.441°S, 151.296

192 °E, 151.671 °E) for years between 2003-2014. Local weather stations were used to estimate average
193 minimum, maximum air temperature as well as mean number of days above 35 °C at each location
194 (Bureau of Meteorology (BOM), Australian Government, <http://www.bom.gov.au>; Terry Hills station
195 from 1954-2014 and Yamba from 1877-2014)). Data reflect average temperatures recorded from when
196 station became operational. To document the temperature variability locally within *H. banksii* beds at
197 each location, single HOBO® pendant loggers (Onset®, USA) were drilled into the substrate at high
198 and low tidal heights and temperatures were recorded between April and July 2014 (Fig. S3). Data only
199 overlapped for 7 days in June 2014 due to some loggers going missing, therefore were only used to
200 view the temperature variability within each shore height at each location and were not compared
201 between locations.

202 203 **2.3. Assessment of genetic diversity**

204 To assess genetic structure of populations, 32 thalli (approximately 1 m apart) were haphazardly
205 sampled from the low and high shore at each location within each of the regions. High shore thalli were
206 selected based on tidal exposure, local topography and drainage patterns to ensure they contrasted with
207 low shore thalli, which were immediately adjacent to the water's edge. Collection at low and high tidal
208 heights were separated horizontally by ~ 5 – 10 m above the low tide mark. Extraction of genomic
209 DNA was conducted for a total of 235 individuals. Samples comprised of unfouled apical segments
210 that were washed in freshwater to remove salts and epiphytes, snapped frozen in liquid nitrogen before
211 storing in a -80 °C freezer until use. Before DNA extraction, samples were freeze-dried overnight.
212 Genomic DNA was isolated from 20–30 mg of freeze-dried tissue using the Nucleospin® 96 Plant II
213 DNA extraction kit (Machery-Nagel, AGRF). Individuals were genotyped using 10 microsatellite loci
214 as described in Bellgrove et al. (2017). Each 11 µL PCR reaction was set up in which consisted of 5
215 µL 2 x Multiplex Mastermix (Qiagen), 4 µL Primer mastermix and 2 µL of 1 in 20 diluted genomic
216 DNA. Primer mastermix consisted of 10 µM reverse primer, 10 µM forward primer and 10 µM unique
217 fluorophores (FAM, VIC, NED, PET) which tagged the flanking regions of the microsatellites.

218
219 Multiplex PCR reactions were run on a Veriti 96-well thermal cycler (Applied Biosystems) with the
220 PCR conditions of 95°C for 15 min for denaturing, followed by 40 cycles of 94°C for 30 s, 59°C for
221 90s, 72°C 60 s, and a final elongation step at 60°C for 30 mins as per the protocol described in Blackett
222 et al. (2012). PCR products were checked for amplification using 1.5% agarose gel before fragment
223 separation was conducted using ABI Genescan 3730 using the size standard LIZ500 (AGRF).
224 Polymorphisms and allele sizes were visualised and determined manually using GeneMapper (v 4.0,
225 Applied Biosystems).

226 227 **2.4. Effects of temperature on phenotypic traits**

228 To assess how functional thermal responses differed within and across populations, quantitative
229 breeding designs were set up to partition variance amongst different genotypes (families), amongst
230 different vertical heights on the shore and between the warm and non-edge location. Adult thalli were
231 collected two hours before absolute low tide to prevent desiccation-induced spawning (Gunthorpe et
232 al., 1995). Thalli were collected from the low shore, directly adjacent to the seaward edge of the rock
233 platform, and the high shore, in the upper intertidal, 5–10 m vertically distance from low shore region.
234 Thalli were transported on ice and gametes extracted within 48 hours.

235
236 To induce spawning, thalli were gently agitated in tap water (room temperature), blotted dry, placed
237 into individual containers and allowed to desiccate at room temperature (Doblin and Clayton, 1995).
238 After 20 min, gametes were released from conceptacles through osmotic stress and desiccation and the
239 sex of the thallus identified by the colour of its gametes: olive green for females and orange for males
240 (Osborn, 1948). As in Clark et al. (2013), three males and three females from each shore height within

241 each location were used in a North Carolina II breeding design (Lynch and Walsh, 1998), where each
242 male was cross fertilised with each corresponding female in a fully factorial design, yielding nine
243 unique genotypes. Each egg and sperm solution were filtered through nylon mesh (100 μm for egg
244 solution and 40 μm for sperm solution) to filter out debris and larger algal material before being mixed
245 to initiate fertilisation. Aliquots of each egg and sperm solution were distributed amongst multiple petri
246 dishes filled with 0.7 μm (Whatman GFF) filtered seawater containing eight glass coverslips for
247 zygotes to attach to. Petri dishes with settled zygotes were then randomly allocated to each of six
248 temperatures; 22, 24, 26, 28, 30 and 32 $^{\circ}\text{C}$ in a Climatron Plant Growth Chamber (Thermoline
249 Scientific, Australia). These temperatures are representative of temperatures experienced by both
250 populations (from climate weather and HOBO pendants) as well as designed to be physiologically
251 stressful to test thermal performance of intertidal macroalgae in populations at the warmer end of the
252 distribution. To examine thermal responses under more realistic fluctuating environments and hence
253 estimate realized rather than fundamental thermal reaction curves, ± 5 $^{\circ}\text{C}$ diel cycle was implemented
254 (Paaijmans et al., 2013). This temperature regime was determined from examining field data from the
255 HOBO® pendant loggers in a pilot study. Germlings were incubated in a 12:12 hour light cycle at 30
256 ± 5 $\mu\text{mol photons m}^{-2}\text{s}^{-1}$.

257
258 At 120 h after fertilisation, germlings growing on coverslips were removed from temperature
259 treatments, wet mounted on a microscope slide and photographed using a light microscope (Olympus
260 BX50, Japan) with AnalySIS imaging software (v 5.0, Japan). Total germling length (defined as
261 extension along the primary rhizoid axis) was calculated from digital images using Image J (National
262 Institutes of Health, USA V1.6.0_24). To examine stages of development, germlings were scored into
263 five ontogenetic stages: 0 = fertilisation through condensation of the chloroplasts; 1 = protrusion of
264 germling cell wall to create a pear shape which later develops into the rhizoid; 2 = division of the
265 germling germinating cell and elongation of a single rhizoid; 3 = elongation of the rhizoid coupled with
266 secondary and tertiary rhizoid development; and 4 = paraphysis development (apical hairs) on top of
267 the germinating cell. These equate to stages 1, 2, 3-5 and 6 (a-d), respectively, in Clarke and Womersley
268 (1981).

269
270 A pulse amplitude modulated fluorometer (Microscope Imaging-PAM, Walz GmbH, Germany) was
271 used to examine photophysiological traits of multiple germlings in one field of view while returning
272 individual measurements from each individual. The system comprises a modified epi-fluorescence
273 microscope (AxioScope.A1, Zeiss) equipped with a modulated LED light source and a photomultiplier
274 for detection of modulated chlorophyll-*a* fluorescence. Germlings attached to cover slips (120 h post
275 fertilisation) were wet mounted and scanned under green light (non-stimulatory for photosynthesis).
276 Selection of germlings for measurement involved maximising the number of germlings in a field of
277 view, as coverslips were discarded after each fluorescence assay. To ensure we obtained high-quality
278 data across all treatments and limit a potential diel effect in photophysiological assessments, we had to
279 restrict our photophysiological measurements to one assay per coverslip. Therefore, the coverslip was
280 briefly scanned to find a field of view where there were numerous germlings sufficiently close together
281 for simultaneous assessment. Although coverslips had the same zygote density before being randomly
282 allocated to temperature treatments, the germling density differed at the time of measurement (120 h),
283 varying between two and eight. Gain settings were adjusted so that base fluorescence (F_t) was between
284 0.1 and 0.3 for each germling, ensuring fluorescence signals were detectable but not too high to cause
285 saturation during measurements. To use F_t as a proxy for pigment content, values collected at different
286 gain settings were standardised to a single gain setting to make them comparable amongst all
287 germlings. This was done by recording the base fluorescence (F_t) of 120 h old germlings determined
288 using the full range of gain settings. The increase in F_t with gain was then fitted with a log linear

289 regression model ($R^2 > 0.90$) so that F_t could be estimated amongst all germlings, no matter what gain
 290 setting was used to perform the steady-state analysis (Kramer et al., 2004).

291

292 The measurement protocol involved dark adapting germlings for 5 min before a single saturating pulse
 293 of blue light (blue Zeiss LED-Module 470 nm; pulse duration = 0.6 s; pulse intensity $> 3000 \mu\text{mol}$
 294 photons $\text{m}^{-2} \text{s}^{-1}$; F_M determination), followed by a two-step steady-state light curve. The steady state
 295 analysis included consecutive 5 min exposures to actinic blue light of 32 and 113 μmol photons $\text{m}^{-2}\text{s}^{-1}$
 296 (sub-saturating and saturating irradiance, respectively, previously determined in a steady state P vs I
 297 fluorescence curve). Saturating pulses of blue light were spaced 30 s apart to monitor F_M' and F_t . Due
 298 to the length of time needed for measurements, photosynthetic traits were only determined for 24, 28
 299 and 32 °C (3 of the 6 temperature treatments) and represented the highest and stressful temperature for
 300 both populations with other two temperatures equally spanning the thermal gradient.

301

302 Maximum quantum yield of photosystem II (PSII), F_v/F_M , was calculated according to the equation
 303 $(F_M - F_0)/F_M$ (Schreiber, 2004). F_v/F_M is the measure of photosynthetic efficiency of PSII after dark
 304 acclimation of photosystems (Genty et al., 1989) with greater values (0-1) equating to greater number
 305 of photosystems available for light capture. Effective quantum yield of PSII, $\Delta F/F_M'$, was calculated
 306 using $(F_M' - F_t)/F_M'$. The proportions of energy being used in photochemistry Y(PSII), regulated non-
 307 photochemical quenching Y(NPQ), i.e. energy dissipation through the rapid conversion of xanthophyll
 308 pigments) and unregulated non-photochemical quenching of excitation energy (Y(NO; i.e., heat
 309 dissipation) were calculated for each actinic light level assuming $Y(\text{PSII}) + Y(\text{NPQ}) + Y(\text{NO}) = 1$
 310 according to (Kramer et al., 2004). To estimate the capacity of germlings to deal with high light, the
 311 relative NPQ between high light (HL) and low light (LL) steps was calculated:
 312 $(\text{HLY}(\text{NPQ})/\text{LLY}(\text{NPQ}))$. A value less than one means there is less NPQ under high light and the
 313 xanthophyll cycle has exceeded its capacity to deal with excess energy.

314

315

316 2.5. Statistical analyses

317 2.5.1. In situ temperature regimes

318 The difference in temperature variability between high and low shores at each location, temperature
 319 data obtained by HOBO® was tested using Levene's test of variance. SST recorded for each location
 320 (MODIS-Aqua satellite 2003-2014) was analysed with a two-factor ANOVA to test for differences
 321 between location and months.

322

323 2.5.2. Estimates of genetic diversity and structure

324 Prior to analyses, genotyping errors such as null alleles, stuttering, dropped alleles, and typographic
 325 errors were checked using MICROCHECKER (Van Oosterhout et al., 2004). Estimates of allelic
 326 frequencies, observed (H_o) and expected (H_e) heterozygosity and departures from Hardy-Weinberg
 327 equilibrium were conducted in GENETIX (v 4.05.2, Belkhir et al., 2000). For each measure of genetic
 328 variation, univariate analyses of variances (ANOVA) were conducted to test for differences between
 329 locations with between regions (warm-edge and non-edge) as well as between heights on the shore
 330 using PRIMER-E with PERMANOVA (v 6.1.16).

331

332 F_{IS} , the proportion of genetic variance contained in an individual (i) relative to the variance contained
 333 in a subpopulation (s), and F_{ST} , the proportion of genetic variance contained in a subpopulation relative
 334 to total genetic variance (τ) were estimated using the program FSTAT (v 2.9.3.2, Goudet, 1995) where
 335 Weir & Cockerham's estimates of F_{ST} and pairwise comparisons of shore heights were calculated
 336 within each location. F_{ST} estimates the genetic differentiation among populations and ranges from 0 to
 337 1. F_{ST} values of 0-0.05 indicate little differentiation, 0.05-0.25 indicate moderate genetic differentiation

338 and values over 0.25 represent pronounced levels of genetic differentiation (Freeland et al., 2011). F_{IS}
339 estimates the amount of selfing or inbreeding occurring within a population and ranges between -1 and
340 1, where negative values represent an excess of heterozygotes and positive values represent an excess
341 of homozygotes. F_{IS} estimates were tested for significance using GENETIX. Linkage equilibrium was
342 tested using 1000 permutations in FSTAT. F_{ST} estimates among all pairs of populations were
343 calculated in FSTAT and significance levels of pairwise comparisons were corrected using Bonferroni
344 correction (Rice, 1989).

345

346 Analysis of molecular variance (AMOVA) was performed using ARLEQUINN (v 3.5.22, Excoffier et
347 al., 2007) which calculated the percentage of genetic variation attributed among and within each
348 location. This analysis was conducted twice to determine variation amongst regions, among locations
349 within regions and within locations (among individuals). F_{ST} estimates indicated no significant
350 differences between shore heights, therefore shore heights were pooled for the AMOVA analysis.
351 Isolation by distance was tested using Mantel tests in IBD WebService (Jensen et al., 2005,
352 <http://ibdws.sdsu.edu>) which tests the null hypothesis of no correlation between pairwise geographic
353 distance and genetic distance matrices.

354

355 **2.5.3. Effects of temperature on phenotypic traits**

356 Physiological responses (variation in germling length and ontogenic development) were analysed with
357 permutational ANOVA, with location (warm-edge and non-edge), height on the shore (low and high)
358 and temperature (six levels) as fixed factors, and male and female identity as a random factor nested
359 within each combination of height on the shore and location. For significant location x temperature or
360 shore height x temperature interactions, main effects were tested with a reduced two-factor ANOVA
361 to compare each temperature. Tukey's HSD post-hoc comparisons were conducted on significant
362 interactions to determine temperature effects. Thermal breadth was obtained by arbitrarily setting a
363 threshold of 80% of maximum germling length for each height on the shore and location combination.
364 Photophysiological traits were analysed using ANOVA with location, height on the shore and
365 temperature as fixed factors (with the low fecundity of some combinations requiring that males and
366 females were pooled). Univariate ANOVAs were conducted in the PERMANOVA routine of Primer-
367 E (v6) and the proportion of variance explained by each factor calculated by least square estimates of
368 variance components (Anderson et al., 2008). Data was visualised using package 'ggplot2' (Wickham,
369 2009) and post-hoc comparisons were conducted in R Studio (version 1.2.5019) using R (R Core Team,
370 2020, version 3.6.2 (2019-12-12)).

371

372 **3. Results**

373 **3.1. Thermal exposure**

374 Mean maximum monthly air temperatures recorded at local weather stations were similar at both
375 locations, however minimum temperatures were lower at the non-edge location. The warm-edge
376 location experienced a narrower temperature range, 9.7 to 26.7 °C compared to 4.8 to 27.6 °C at the
377 non-edge location (Fig. S1a,b)—a difference between max and min air temperatures of 17.0 °C and
378 22.8 °C, at each location respectively (Fig. S1a, b). The warm-edge location experienced fewer days
379 per year where air temperature exceeded 35 °C than the non-edge location (on average ~1 vs ~8 d y⁻¹,
380 respectively; Fig. S1). Sea surface temperature recorded between 2003-2014 by satellite (MODIS-
381 Aqua; <https://oceancolor.gsfc.nasa.gov/data/aqua/>), ranged from 19.4 to 26.7 °C at the warm-edge
382 location versus 16.3 to 24.4 °C at the non-edge location and were significantly different between
383 locations ($F_1 = 1619$, $P < 0.001$) and months ($F_{11} = 325$, $P < 0.001$; Fig. S2c).

384

385 At the local scale, HOBO® sensors recorded similar mean temperatures on the low and high shore at
386 the warm-edge location (21.9 ± 2.8 °C vs 21.5 ± 3.6 °C, respectively; Fig. S2) as well as the non-edge

387 location (16.0 ± 2.4 °C vs 15.3 ± 3.5 °C, respectively; Fig. S2). High shore temperatures were however
 388 more variable than low shore temperatures at both locations (warm-edge: $F_{1,2074} = 55.89$, $P = 0.001$;
 389 non-edge: $F_{1,1960}$, $P = 0.001$).

390

391 **3.2. Genetic diversity and structure**

392 Null alleles were found in warm-edge populations, Minnie Water (MW) and Angourie (ANG) at low
 393 and high tidal heights for locus HB3 but not at any other locations or loci. All analyses were run with
 394 and without HB3 and results were consistent, therefore the locus HB3 was kept in subsequent analyses.
 395 Linkage disequilibrium was not found for any locus and all loci were in Hardy-Weinberg equilibrium
 396 (Table S1). Amongst the 235 individuals collected, a total of 28 different alleles were genotyped across
 397 10 loci. Total mean (\pm SD) number of alleles across all locations and shore heights sampled was 26.00
 398 ± 1.18 where unique alleles were found at ANG high shore (2 alleles), MW low shore (1 allele) and
 399 BB low shore (2 alleles) populations (Table 1). The total number of alleles in any single population
 400 varied from 24 to 28 and was similar between regions ($F_{1,79} = 0.055$, $P = 0.820$) amongst all locations
 401 ($F_{3,79} = 0.106$, $P = 0.967$), and between heights on the shore (Table 1; $F_{1,79} = 0.053$, $P = 0.814$).

402

403 Genetic diversity was determined by expected heterozygosity (H_E) which was lower in warm-edge
 404 populations (Table 1, Table S1). A trend for positive (but not significant) F_{IS} values indicated some
 405 selfing or inbreeding at both warm-edge and non-edge populations but a significant deviation from
 406 random mating (negative F_{IS} value) was only found in low shore populations at Pearl Beach, indicating
 407 an excess of heterozygotes (Table S2).

408

409 The overall F_{ST} estimate amongst all spatial scales tested was 0.256 indicating high genetic structure,
 410 but genetic structuring diminished from regional to local scales (Table 2). There was pronounced and
 411 significant levels of genetic structuring between pairs of non-edge (Bilgola and Pearl Beach) and warm-
 412 edge populations (Minnie Water and Angourie) with F_{ST} values ranging between 0.274 - 0.413 (Table
 413 2). Among non-edge locations there was moderate and significant genetic structure with pairwise F_{ST}
 414 values ranging from 0.150 – 0.190 (Table 2). Similarly, there was low but significant genetic structure
 415 between locations within the warm-edge region (Minnie Water and Angourie; $F_{ST} = 0.066$ - 0.105
 416 (Table 2)). Pairwise F_{ST} estimates between vertical shore heights were not significantly different at
 417 any location (Table 2).

418

419 There was a large (25.7 %) and significant amount of genetic variation explained by differences in
 420 regions, (warm-edge vs non-edge). Similarly, a large and significant amount of genetic variation
 421 (8.0%) occurred between locations within each region, with the largest amount of genetic variation
 422 (66.3%) amongst individuals within each location ($F_{ST} = 0.337$, $P < 0.001$). When shore heights were
 423 pooled within each location, a separate AMOVA revealed greater variation was explained at the scale
 424 of locations (28.0 %) and amongst individuals within each location (72.5%; $F_{ST} = 0.275$, $P < 0.001$).
 425 There were strong significant relationships between geographic and genetic distance across all
 426 locations (Mantel test: $Z = 2620.41$, $r = 0.924$, $P = 0.002$, Fig. S1) but not within non-edge (Mantel
 427 test: $Z = 7.56$, $r = 0.976$, $P = 0.113$, Fig. S3) or warm-edge regions (Mantel test: $Z = 6.90$, $r = 0.831$, P
 428 $= 0.250$, Fig. S3).

429

430 **3.3. Effects of temperature on germling growth and ontogenesis**

431 Temperature had a strong effect on germling growth, resulting in distinct thermal performance curves
 432 at each location and height on the shore (Fig. 2). Overall, there was a significant *location* \times *temperature*
 433 interaction (15% of the total variation in germling length explained by this interaction) as well as a
 434 significant *shore height* \times *temperature* interaction (3 % of total variation; Table 3) and the interaction
 435 between *male-female* \times *temperature* contribution was 2%. Separate two-factor ANOVAs found

436 significant differences for the main effects of location and height on the shore for each temperature
437 (Table S3) with *location* × *shore height* interactions found for germling length at 22, 24, 26, and 30
438 °C (Table S3, Fig 2). Among these interactions, Tukey's post-hoc comparisons were significant
439 between all *location* × *shore height* combinations at each temperature except for some combinations
440 at 22, 26 and 30 °C (see Table S4 for all pairwise results).

441
442 Thermal breadth, defined as the temperature range in which germling length reached 80% of maximum
443 germling length, varied at each shore height nested in each location. At the warm-edge population
444 (MW) germling lengths from both low and high shore demonstrated greatest growth in the cooler range
445 of temperatures tested (24–26°C). For overall length, warm-edge high shore germlings, were the
446 longest at 24 and 26 °C compared to all other heights on the shore tested (mean length ± SD; 408.73 ±
447 46.03 μm and 399.49 ± 42.61 μm respectively) and declined significantly in temperatures beyond 26
448 °C (22 < 24 = 26 > 28 > 30 > 32 °C; P<0.05, Table S5; Fig 2). Similarly, warm-edge low shore
449 germlings grew optimally at 22- 26°C, (274.03 ± 37.21 μm – 352.52 ± 66.11 μm) with a significant
450 decline in growth after 26 °C (22 = 24 = 26 > 28 > 30 = 32 °C; P<0.01, Table S5; Fig 2). For the non-
451 edge population, germling growth was greater in warmer range of temperatures tested. Non-edge high
452 shore germling growth was sustained across a wider thermal breadth between 22–28°C and low shore
453 between 24–28 °C (Fig. 2). Non-edge high shore germlings were the greatest between 24 - 28°C
454 (356.48 ± 43.84 μm – 377.24 ± 33.14 μm) and were significantly longer than other temperatures tested
455 (22 < 24 = 26 = 28 > 30 > 32 °C; P <0.001, Table S5, Fig 2). Non-edge low shore germlings growth
456 peaked at 28 °C (204.78 ± 41.44 μm) and were significantly greater than germling length tested at other
457 temperatures ((22 < 24 = 26 < 28 > 30 > 32 °C; P <0.05, Table S5, Fig 2). For all germlings despite
458 location or height on the shore, temperatures beyond 28 °C showed significantly reduced growth of
459 germlings with substantial inhibitory effects at 30 °C for warm-edge germlings. At the most extreme
460 temperature, 32 °C, germlings from the non-edge location grew 3-4 fold slower than at their maxima
461 and those from the warm-edge location grew 4-5 fold slower.

462
463 At the level of the individual, the effect of temperature varied significantly with male identity for
464 germling length for shore height and location (*temperature* × *male* interaction; Table 3), which
465 provides evidence of heritable genetic variation in thermal tolerance in different populations of
466 *Hormosira banksii*. The effect of temperature also varied with parental identity (a significant
467 *temperature* × *male* × *female* interaction, Table 3) with ~2% of the variance in growth of all germlings
468 attributed to the variation in temperature effects among male/female combinations.

469
470 The effect of temperature on the ontogenic development of germlings varied among locations, heights
471 on the shore and among genotypes (Fig. 3). Overall, germlings from the warm-edge location developed
472 more rapidly than those from the non-edge location (significant *location* × *temperature* interaction for
473 stages 0, 3 and 4, Table 4), with up to 60% of warm-edge germlings reaching stage 3 or 4 after 5 days,
474 in contrast to only 20% from non-edge (Fig. 3a, c). The proportion of warm-edge germlings with
475 delayed development (i.e. stage 0) increased steadily in temperatures that surpassed the temperatures
476 optimal for growth (24–26 °C) reaching ~85% at 32 °C. At the non-edge location, between 5 and 15%
477 of germlings had delayed development across all temperatures except for high rates of ~40% at 32 °C
478 (Fig. 3b, d). Development at 32 °C was characterised by enlargement of the germinating cell rather
479 than through cell differentiation and rhizoid development.

480
481 The proportion of germlings at any stage, or with delayed development in stage 0, did not vary with
482 temperature and height on the shore in either location (non-significant *temperature* × *height* on shore
483 interactions, Table 4). However, the effect of temperature varied significantly with male identity for

484 the proportion of germlings in stage 2- 4 (genotype by environment interaction; Table 4) indicating
485 that there is heritable genetic variation in the effects of temperature on rates of ontogenic development.
486

487 **3.4. Effects of temperature on photophysiological traits**

489 Temperature had a direct effect on the photochemical efficiency of PSII at a sub-saturating and
490 saturating light intensities (LY(II), HY(II), respectively) amongst all germlings, but there were no
491 significant interactions with location or heights on the shore (Table 5; Fig. 4). Germlings all showed
492 similar photophysiological responses to increasing temperature, with maximum quantum yield and
493 photosynthetic efficiency being relatively constant between 24 °C and 28 °C and decreasing at 32 °C
494 (Fig. 4). Maximum quantum yield (F_v/F_m) was generally greater in warm-edge germlings compared
495 to non-edge germlings but did not differ among heights on the shore or temperature (Fig. 4, Table 5).
496

497 High and low shore germlings used regulated nonphotochemical quenching ($Y(NPQ)$) as a means of
498 photoprotection, with $Y(NPQ)$ remaining similar under ambient and high light intensity (Fig 4, Table
499 5). However, germlings from the warm-edge location diverted more energy proportionally to $Y(NPQ)$
500 under high light intensities than those from the non-edge location (Fig. 4; Table 5). This was also
501 evident with the ratio of regulated nonphotochemical quenching under ambient and high light (HL: LL
502 $Y(NPQ)$, Fig. 4, Table 5) where the germlings from the warm-edge location had ratios above 1
503 (indicating increased regulated quenching of energy at saturating light intensity) compared to
504 germlings from the non-edge location which had ratios below 1 (indicating increased unregulated
505 quenching, or potential photodamage under high light). Baseline fluorescence (F_i), a proxy for
506 photosynthetic pigment content, was significantly higher in the germlings from the non-edge location
507 and those from high on the shore (Table 5). For all photophysiological traits, there were no significant
508 interactions between temperature and location or height on the shore, indicating that germling
509 responses to temperature did not vary at regional and local scales (Table 5).
510

511 **4. Discussion**

512 Populations inhabiting the warm range edge of distributions are suggested to be at the forefront of
513 climate change. The increasing prevalence of extreme climate events such as heatwaves may challenge
514 marine macroalgal dominated populations that have limited physiological plasticity to tolerate
515 prolonged, elevated temperatures or reduced capacity to adapt (Wernberg et al., 2018; Gurgel et al.,
516 2020). In this study, germlings of the dominant intertidal macroalga, *Hormosira banksii*, demonstrated
517 contrasting thermal performance curves which was governed by the thermal environment from where
518 they originated rather than their relative distributional origin. Warm-edge germlings had greater growth
519 rate and development and a narrower thermal breadth, but were sensitive to higher temperatures
520 compared to non-edge germlings, which grew optimally across a wider range of temperatures. Relative
521 position on the vertical shore, had a greater influence on thermal physiology and breadth illustrated by
522 wider thermal breadth for non-edge, high shore germlings. Warm-edge germlings, however, had
523 greater capacity to regulate excess energy as nonphotochemical quenching ($Y(NPQ)$) when exposed to
524 greater temperature and light, suggesting they are less photophysiological sensitive than non-edge
525 population germlings. Evidence of heritable genetic variation (significant genotype by environment
526 interaction) found for growth and development indicate that there is potential for adaptation in thermal
527 tolerance traits. These physiological responses coincided with lower genetic diversity, restricted gene
528 flow and evidence of inbreeding at warm-edge populations. This suggests that warm-edge germlings
529 utilise physiological plasticity to tolerate short-term exposure (hours-days) to environmental stressors
530 but over longer time scales (years) may potentially be less thermally buffered and at greater risk to
531 global warming.
532

4.1. Thermal effects of physiology

Thermal history at study locations played an important role in governing thermal tolerances in *H. banksii* germlings. Germlings grew and developed faster in the warm-edge population but in the cooler, narrower range of temperatures tested, compared to non-edge germlings. Although previous research found that 90% of seaweeds displayed population level variation in upper thermal limits (King et al., 2017), which agrees with our findings, the result of increased thermal sensitivity to high temperatures for warm-edge germling growth is contrary to previous research. Many studies on marine macrophytes and invertebrates in lower latitudes were found to be more tolerant of higher temperatures, as they generally experience greater temperatures throughout their life history (Gerard and Du Bois, 1988; Stillman and Somero, 2000; Kelly et al., 2012; Sunday et al., 2012; Mota et al., 2018). Air temperature data collected from local meteorological stations demonstrated that the warm-edge population (Minnie Water) experience similar annual maximum monthly temperatures as the non-edge population (Pearl Beach), but warmer minimum monthly temperatures, less seasonal variation and fewer days over 35 °C. This would suggest that the warm-edge population should also be more thermally tolerant. However, previous research on the effects of desiccation stress on adult thalli of *H. banksii* from the same warm-edge site (Minnie Water), adults were also more thermally sensitive to higher temperatures (Clark et al 2018). Further, previous research also found a warm-edge macroalgal population was not more thermally tolerant than cooler populations and proposed that the warm-edge population was thermally maladapted (Pearson et al., 2009). The low gene flow and low genetic diversity found in the warm edge populations of this study may support local adaptation to conditions (discussed in section 4.2), however physiological adaptation and local site effects may also play an important role in shaping thermal performance (discussed further in this section).

Our result of narrower thermal breadth of warm-edge population is consistent with the climate variability hypothesis of narrower thermal breadth towards lower latitudes (Stevens 1989). However, a recent study of a non-edge population of a subtidal macroalgal species (*Scythothalia dorycarpa*) had similar thermal safety margins (defined as the the temperature buffer between an organisms upper thermal-tolerance limit and the maximum ambient temperatures it experiences') to warm-edge populations but different absolute temperature tolerances, which demonstrated that not all species at distributional limits have a narrower thermal breadth (Bennett et al., 2015). One explanation for our result of different thermal breadths for both populations is that intertidal species are exposed to dynamic environmental stress imposed by the terrestrial and marine environment, opposed to constantly being submerged, therefore differences such as emersion and air temperature variation may be more important in shaping thermal niche. The significant difference in thermal breadth in non-edge high and low shore germlings illustrates how local scale effects in the intertidal can influence thermal performance.

Despite locational differences in thermal regimes among locations, germlings of *H. banksii* demonstrated similar photophysiological responses to elevated temperatures. Divergence among locations in the ability to tolerate greater light intensity, however, was found for dissipation of excess energy. The lack of any interactions between temperature and location for photophysiological parameters in *H. banksii* suggests that germlings have a high degree of plasticity, and can adjust their photosystems to tolerate differences in light and temperature regimes. Despite significant reductions of growth at 28 °C for warm-edge germlings and 30 °C for non-edge germlings, *H. banksii* was still able to maintain a high level of PSII efficiency in ambient and high light intensities across 24 and 28 °C, suggesting acclimation of photosystems (Major and Davison 1998). This result is consistent with previous studies which also found no significant temperature interactions in photosynthetic response of macroalgae (Clark et al., 2013; McCoy and Widdicombe, 2019). The adjustment of photosystems to different temperatures and light intensities to optimise photosynthesis may be an important trait for

582 intertidal macroalgae as temperature and light gradients can change rapidly with wave action and tidal
583 cycles. Furthermore, in locations closer to the equator, light intensity is greater seasonally, therefore
584 warm-edge germlings may be able to tolerate higher light intensity through phenotypic plasticity
585 indicated by more energy being dissipated via Y(NPQ) rather than Y(NO), whereas non-edge
586 germlings are more light sensitive indicated by the greater proportion of Y(NO). This may also explain
587 the greater growth reduction at higher temperatures for germlings in the warm-edge location as more
588 energy is being diverted towards photoprotection rather than to photochemistry. In sporophytes of the
589 subtidal kelp *Ecklonia radiata*, physiological performance was maintained in higher temperatures
590 through an increase in critical light demand (E_c) (Staehr and Wernberg, 2009). This reduction allowed
591 for similar levels of light limited photosynthesis to be achieved in warm and cool adapted populations
592 found at different latitudes, consistent with this study.

593 Growth of germlings from low and high on the shore was also affected by differences in local
594 temperatures, indicated by significant interactions between height on the shore and temperature.
595 Temperatures recorded by HOBO pendants in the high shore at both locations were significantly more
596 variable than low shore temperatures. These results are consistent with a growing body of research
597 that suggests that local scale topography and environmental conditions may be more important in
598 driving physiology and species' distributions than larger regional effects of climate (Helmuth et al.,
599 2002, 2006; Helmuth, 2009). For example, local scale topography and environmental conditions
600 experienced by individuals of the intertidal mussel *Mytilus californianus* can result in body
601 temperatures varying between 6 to 13° C within a population at a given time (Helmuth and Hofmann,
602 2001; Harley, 2008). Consequently, temperatures experienced by individuals may not be easily
603 predicted by larger scale variation in temperatures (e.g., among latitudes), but instead be a mosaic of
604 smaller scale hot and coldspots. In this study, the warm-edge location is characterised by large boulders
605 that can shade *H. banksii* and trap small pools of water, potentially reducing the stress experienced by
606 individual thalli in contrast with temperatures experienced on flatter rock platforms such as at the non-
607 edge location. The shore topography at the warm-edge location could thus modify the thermal exposure
608 of individuals and lead to similar growth rates of germlings from low and high on the shore as found
609 in this study.

610 Maintaining thermal tolerance across broader temperatures can be physiologically costly, therefore
611 germlings may not grow optimally across all temperatures (Huey et al., 2012). This is demonstrated
612 by differences in optimal temperatures for germling growth and may reflect increased energy
613 dissipation (i.e., non-photochemical quenching) and decreased photochemistry (YII) with increased
614 temperatures and light intensities likely experienced for longer periods during low tide high on the
615 shore (Davison and Pearson, 1996). In addition, the reduced growth and narrow thermal optima
616 amongst low shore germlings in both populations may reflect light limited photosynthesis of adults as
617 they experience longer periods spent submerged compared to those on the high shore, while optimising
618 growth within a narrow range of temperature that they most commonly experience (Huey et al., 2012).
619 There were no significant interactions between temperature and height on the shore for photosynthetic
620 parameters, suggesting phenotypic plasticity for these photophysiological traits. Given that intertidal
621 macroalgae at different heights on the shore must contend with dynamic variation in light and
622 temperature during daily tidal cycles, it suggests that photosystems need to be able to rapidly
623 acclimatise to different light and temperature regimes (Hanelt et al., 1993). Over longer time scales,
624 adaptation of the population at the local scale involving genotypes tolerant to the prevailing thermal
625 and light regime may also be important (Hanelt et al., 1993; Al-Janabi et al., 2019).

626 The potential for adaptation in temperature tolerance traits amongst *H. banksii* germlings is indicated
627 by a significant male x temperature interaction for germling length and ontogenic development. This

628 is consistent with earlier investigations of this species (Clark et al., 2013) and suggests that as
629 temperatures increase with global warming, genotypes that are better able to tolerate higher
630 temperatures will be favoured (Deutsch et al., 2008; Sunday et al., 2012; Fusi et al., 2015). This will
631 be particularly important for populations that have limited gene flow such as the warm-edge
632 population. A significant interaction between female identity and temperature was also found for the
633 proportion of germlings that did not develop (stage 0), suggesting a role for either female genotype or
634 non-genetic maternal effects in thermal responses. Maternal effects have been identified previously in
635 different organisms (e.g. bryozoans, Marshall, 2008; terrestrial plants, Galloway et al., 2009; sea
636 urchins, Foo et al., 2012; fish, Chambers and Leggett, 2015) and are potentially relevant in *H. banksii*
637 where egg size differs among different females (Clark, 2016). Maternal environment may impact the
638 resources available for reproduction which can affect egg size and growth trajectory of offspring (Wolf
639 and Wade, 2009). The significant interaction between temperature and parental identity (i.e., *male* ×
640 *female* × *temperature*) suggests that different genotypes are more susceptible to different temperatures.
641 There were no interactions between temperature and male or female identity for any of the
642 photosynthetic parameters, suggesting that photosynthesis is highly regulated amongst individuals.
643 This agrees with previous studies in which no heritable genetic variation was found in *H. banksii*
644 photosynthetic traits (Clark et al., 2013).

645 4.2. Genetic diversity and structure

646 Consistent with previous studies (Coleman et al., 2011a, 2019; Miller et al., 2019) we found strong
647 genetic structure between the warm-edge and non-edge regions (~ 500 km apart) as well as isolation
648 by distance suggesting that dispersal capacity is limited across long distances as well as between
649 neighbouring populations (> 50 km). Moreover, trends for lower estimates of genetic diversity towards
650 distributional edges found in this study is in accordance with previous studies of *H. banksii* across a
651 longitudinal gradient (Miller et al., 2019) and other macroalgal species (Faugeron et al., 2004; Teixeira
652 et al., 2016; King et al., 2017; Wernberg et al., 2018). The observed patterns of lower genetic diversity
653 at warm-edge populations is suggested to be the result of reduced gene flow and connectivity which
654 can create isolation among populations and reduce within population genetic diversity (Hampe and
655 Petit, 2005). In addition, as distributional limits often represent the physiological limits of a species,
656 environmental conditions can impose strong selection pressure resulting in decreased diversity as
657 environmental conditions and habitat become suboptimal with only tolerant genotypes and phenotypes
658 persisting at range edges. While we cannot tease apart these mechanisms with the neutral markers used
659 here, the early life stages of *H. banksii* from warm-edge populations had a narrower range of thermal
660 performance compared to populations found within the center of its distribution suggesting that
661 reduced genetic diversity may constrain responses. With lower genetic diversity, the warm-edge
662 population may not have the range of the functional responses such as greater tolerance for higher
663 temperature, however, greater regulated nonphotochemical protection in warm-edge germlings,
664 suggests that this population may have greater phenotypic plasticity to tolerate dynamic light
665 conditions.

666 Moderate gene flow is evident between neighbouring *H. banksii* populations within each warm-edge
667 and non-edge region separated by < 50 km. Dispersal of gametes or zygotes is not a likely method of
668 long-distance dispersal as fertilised zygotes sink to the substrate and adhere within hours of fertilisation
669 (Dimartino et al., 2015). Rather, rafting of buoyant dislodged adult thalli which drift with ocean
670 currents with the aid of air bladders or vesicles has been suggested as the most likely method of long-
671 distance dispersal and has been evident amongst different macroalgal species (Muhlin et al., 2008;
672 Valero et al., 2011; Bussolini and Waters, 2015; Coleman et al., 2019). There is limited empirical
673 evidence that supports whether floating thalli contribute to long distance gene flow in *H. banksii*,

674 however a recent study suggests that floating thalli can end up in estuaries where they can grow and
675 survive (Coleman et al., 2019). The moderate but significant levels of genetic structure between
676 neighbouring populations within each non-edge and warm-edge region may show restriction in gene
677 flow possibly due to the existence of physical barriers such as sandy beaches, and mouths of estuaries
678 which may serve as barriers to gene flow in other macroalgae (Billot et al., 2003; Coleman, 2013).

679
680 Within smaller scales (within 5-10 m), gene flow of *H. banksii* was not restricted between vertical
681 heights on the intertidal shore which agrees with other studies on macroalgae (Engel et al., 2004;
682 Tatarenkov et al., 2005; Teixeira et al., 2016; Bellgrove et al., 2017). The intertidal is characterised by
683 steep environmental gradients suggesting selection for stress-tolerant genotypes on high shores may be
684 an important driver of genetic structure. This has been demonstrated amongst barnacles (Schmidt and
685 Rand, 2001), gastropods (Johannesson et al., 1995) as well as amongst hybrids of the macroalgae *Fucus*
686 *vesiculosus* and *Fucus spiralis* (Billard et al., 2010; Zardi et al., 2011). Nonetheless, the lack of small-
687 scale genetic structure between shore heights found in this study, suggests that gene flow is
688 unobstructed and that *H. banksii* zygotes and gametes may be readily dispersed across these smaller
689 distances (Dudgeon et al., 2001). Studies on the attachment strength of *H. banksii* zygotes have found
690 that adhesion to the substrate is not at maximum strength until 24 h after fertilisation suggesting that
691 zygotes could potentially be dislodged and recruit elsewhere (Dimartino et al., 2015). Specific habitat
692 types related to strong environmental gradients within the intertidal have been found to influence
693 phenotypic divergence independently of genetic structure (Engel et al., 2004; Zardi et al., 2013) Zardi
694 et al. 2013). Lack of differences at small scales suggest that *H. banksii* may survive living in different
695 environmental gradients through phenotypic plasticity rather than genetic differentiation as
696 documented in *F. vesiculosus* (Zardi et al., 2013). This suggests that thermal exposure within the
697 intertidal may not necessarily select for different genotypes but perhaps genotypes that are highly
698 plastic. An alternative explanation is that adaptive genetic differentiation between tidal heights may
699 exist, but is not apparent in our neutral markers (which only show variation due to dispersal and
700 connectivity, not selection). Testing this idea would require use of markers such as SNPs which
701 examine portions of the genome under selection.

702
703 Genetic diversity was found to be lower at warm-edge populations. This is not surprising as these
704 populations are at the edge of their equatorward distribution, where populations are more fragmented,
705 conditions are not optimal and macroalgal populations are therefore at their physiological threshold.
706 The lower genetic diversity found at these populations suggests that these populations may lack the
707 potential to adapt to future warming and be particularly vulnerable to extreme climate events (i.e. heat
708 waves). Previous studies have already shown local extinction in warm-edge populations of macroalgal
709 populations with extreme climate events which may be a consequence of a smaller gene pool (Araújo
710 and Williams, 2001; Smale and Wernberg, 2013; Wernberg et al., 2018).

711

712

713 **5. Conclusions**

714 The results of this study provide evidence that germlings of *H. banksii* inhabiting populations within
715 the warm-edge of its distribution may at risk to increases in temperatures associated with global
716 warming (Fig. 5). The sensitivity to higher temperatures, narrower thermal breadth as well as relatively
717 low genetic diversity and limited gene flow are all indications of populations that are vulnerable to
718 warming (Pearson et al., 2009; Mota et al., 2018; King et al., 2019). Significant genotype by
719 environment interactions found for growth and ontogenic development suggests that there is heritable
720 genetic variation in growth and development under different temperatures, which could be important
721 particularly for the warm-edge populations with lower genetic diversity and gene flow. Our

722 experimental data show that these warm-edge populations may also be surviving through phenotypic
723 plasticity by obtaining similar levels of photochemistry through greater levels of regulated non-
724 photochemical quenching (photoprotection) at higher light and temperature than non-edge population,
725 however over the long-term the genetic impoverishment and reduced gene flow may be problematic as
726 global warming and extreme climate events continue to push species past their physiological limits.
727 The prevalence of greater number of hot days (days over 35 °C) in non-edge populations, suggests that
728 the non-edge population may be at risk to habitat fragmentation. Greater tolerance to higher
729 temperatures as well as the significant genotype x environment interactions suggest that non-edge
730 populations may be locally adapted to local environmental conditions and have heritable genetic
731 variation in thermal tolerance traits. Further greater genetic diversity and gene flow suggests that the
732 non-edge population have greater connectivity and therefore available for genetic rescue from
733 surrounding populations.

734
735 Contrasting the relative magnitude of within-population variation to variation in thermal responses on
736 larger spatial scales, this study shows that the interaction between temperature and location comprised
737 an effect size of 15% of the total variation in growth, the interaction between temperature and heights
738 on the shore had an effect size of 3%, and the interaction between temperature and male-female
739 combination had an effect size of 2%. This within-population variation in thermal tolerance will be
740 particularly important under a changing climate as populations with greater diversity will have a
741 broader suite of tolerant genotypes for selection to act upon (Reusch, 2014; Wernberg et al., 2018).
742 The results of this study and previous research on genetic diversity of *H. banksii* across its species
743 distribution (Miller et al 2019) has helped improve predictions of how this species will respond to
744 ongoing warming and identified potentially sensitive populations. Conservation efforts such as
745 transplanting tolerant individuals or reseeded to increase genetic diversity in genetically impoverished
746 populations may aid in providing greater functional resilience to warming climates (Campbell et al.,
747 2014; Wood et al., 2019; Fredriksen et al., 2020).

748

749 **Data availability**

750 Data is available upon request

751 **Conflict of Interest Statement**

752 The authors declare that the research was conducted in the absence of any commercial or financial
753 relationships that could be construed as a potential conflict of interest.

754 **Author Contributions**

755 JC conceived the idea for the manuscript, collected the data, curated and formally analysed the data
756 and wrote the original manuscript. MD, AP, MC supervised the project, reviewed, wrote and edited
757 the manuscript, MD collected some of the data and provided resources for the project.

758 **Acknowledgements**

759 The authors would like to thank Daniel Lewis for assistance in the field and sample collections and
760 Kevin Davies for providing Figure 1. This research was supported by the University of Technology
761 Sydney Climate Change Cluster (C3), and an Australian postgraduate award to JC. Samples were
762 collected under the Department of the Environment and Heritage permit number P10/0057-2.0.

763

764 **References**

- 765 Al-Janabi, B., Wahl, M., Karsten, U., Graiff, A., and Kruse, I. (2019). Sensitivities to global change
766 drivers may correlate positively or negatively in a foundational marine macroalga. *Scientific*
767 *Reports* 9. doi:10.1038/s41598-019-51099-8.
- 768 Allakhverdiev, S. I., Kreslavski, V. D., Klimov, V. V., Los, D. A., Carpentier, R., and Mohanty, P.
769 (2008). Heat stress: an overview of molecular responses in photosynthesis. *Photosynthesis*
770 *Research* 98, 541–550. doi:10.1007/s11120-008-9331-0.
- 771 Anderson, M., Gorley, R., Clarke, K., Anderson, M., Gorley, R., Clarke, K., et al. (2008).
772 PERMANOVA+ for PRIMER. Guide to software and statistical methods.
- 773 Araújo, M. B., and Williams, P. H. (2001). The bias of complementarity hotspots toward marginal
774 populations. *Conservation Biology* 15, 1710–1720.
- 775 Araújo, R., Serrão, E. A., Sousa-Pinto, I., and Åberg, P. (2011). Phenotypic differentiation at
776 southern limit borders: the case study of two furoid macroalgal species with different life-
777 history traits: phenotypic differentiation at southern limit borders. *Journal of Phycology* 47,
778 451–462. doi:10.1111/j.1529-8817.2011.00986.x.
- 779 Belkhir, K., Borsa, P., Goudet, J., Chikhi, L., and Bonhomme, F. (2000). GENETIX, version 4.05.
780 Laboratoire Genome, Populations. *Interactions, CNRS UPR 9060*.
- 781 Bellgrove, A., Clayton, M. N., and Quinn, G. P. (1997). Effects of secondarily treated sewage
782 effluent on intertidal macroalgal recruitment processes. *Marine and Freshwater Research* 48,
783 137. doi:10.1071/MF96011.
- 784 Bellgrove, A., van Rooyen, A., Weeks, A. R., Clark, J. S., Doblin, M. A., and Miller, A. D. (2017).
785 New resource for population genetics studies on the Australasian intertidal brown alga,
786 *Hormosira banksii*: isolation and characterization of 15 polymorphic microsatellite loci
787 through next generation DNA sequencing. *Journal of Applied Phycology* 29, 1721–1727.
788 doi:10.1007/s10811-016-1015-0.
- 789 Bennett, S., Duarte, C. M., Marbà, N., and Wernberg, T. (2019). Integrating within-species variation
790 in thermal physiology into climate change ecology. *Philosophical Transactions of the Royal*
791 *Society B: Biological Sciences* 374, 20180550. doi:10.1098/rstb.2018.0550.
- 792 Bennett, S., Wernberg, T., Arackal Joy, B., de Bettignies, T., and Campbell, A. H. (2015). Central
793 and rear-edge populations can be equally vulnerable to warming. *Nature Communications* 6.
794 doi:10.1038/ncomms10280.
- 795 Billard, E., Serrão, E., Pearson, G., Destombe, C., and Valero, M. (2010). *Fucus vesiculosus* and
796 *spiralis* species complex: a nested model of local adaptation at the shore level. *Marine*
797 *Ecology Progress Series* 405, 163–174. doi:10.3354/meps08517.
- 798 Billot, C., Engel, C., Rousvoal, S., Kloareg, B., and Valero, M. (2003). Current patterns, habitat
799 discontinuities and population genetic structure: the case of the kelp *Laminaria digitata* in the
800 English Channel. *Marine Ecology Progress Series* 253, 111–121. doi:10.3354/meps253111.

- 801 Bishop, M. J., Morgan, T., Coleman, M. A., Kelaher, B. P., Hardstaff, L. K., and Evenden, R. W.
802 (2009). Facilitation of molluscan assemblages in mangroves by the fucalean alga *Hormosira*
803 *banksii*. *Marine Ecology Progress Series* 392, 111–122. doi:10.3354/meps08247.
- 804 Blacket, M. J., Robin, C., Good, R. T., Lee, S. F., and Miller, A. D. (2012). Universal primers for
805 fluorescent labelling of PCR fragments-an efficient and cost-effective approach to genotyping
806 by fluorescence: UNIVERSAL PRIMERS FOR FLUORESCENT GENOTYPING.
807 *Molecular Ecology Resources* 12, 456–463. doi:10.1111/j.1755-0998.2011.03104.x.
- 808 Burrows, M. T., Schoeman, D. S., Richardson, A. J., Molinos, J. G., Hoffmann, A., Buckley, L. B., et
809 al. (2014). Geographical limits to species-range shifts are suggested by climate velocity.
810 *Nature* 507, 492–495. doi:10.1038/nature12976.
- 811 Bussolini, L. T., and Waters, J. M. (2015). Genetic analyses of rafted macroalgae reveal regional
812 oceanographic connectivity patterns. *Journal of Biogeography* 42, 1319–1326.
- 813 Campbell, A. H., Marzinelli, E. M., Vergés, A., Coleman, M. A., and Steinberg, P. D. (2014).
814 Towards restoration of missing underwater forests. *PloS one* 9.
- 815 Chambers, R. C., and Leggett, W. C. (2015). Maternal influences on variation in egg sizes in
816 temperate marine fishes. *American Zoologist* 36, 180–196. doi:10.1093/icb/36.2.180.
- 817 Clark, J. S. (2016). Assessing the vulnerability of a habitat forming macroalga to climate warming:
818 roles of physiology, ecology and evolutionary processes in determining resilience. PhD
819 Dissertation. [Sydney (NSW)]: University of Technology Sydney.
- 820 Clark, J. S., Poore, A. G. B., and Doblin, M. A. (2018). Shaping up for stress: Physiological
821 flexibility is key to survivorship in a habitat-forming macroalga. *Journal of Plant Physiology*
822 231, 346–355. doi:10.1016/j.jplph.2018.10.005.
- 823 Clark, J. S., Poore, A. G. B., Ralph, P. J., and Doblin, M. A. (2013). Potential for adaptation in
824 response to thermal stress in an intertidal macroalga. *Journal of Phycology* 49, 630–639.
825 doi:10.1111/jpy.12067.
- 826 Clarke, S., and Womersley, H. (1981). Cross-fertilization and hybrid development of forms of the
827 brown alga *Hormosira banksii* (Turner) Decaisne. *Australian Journal of Botany* 29, 497–505.
- 828 Coleman, M. A. (2013). Connectivity of the habitat-forming kelp, *Ecklonia radiata* within and
829 among estuaries and open coast. *PloS one* 8.
- 830 Coleman, M. A., Chambers, J., Knott, N. A., Malcolm, H. A., Harasti, D., Jordan, A., et al. (2011a).
831 Connectivity within and among a network of temperate marine reserves. *PLoS ONE* 6,
832 e20168. doi:10.1371/journal.pone.0020168.
- 833 Coleman, M. A., Clark, J. S., Doblin, M. A., Bishop, M. J., and Kelaher, B. P. (2019). Genetic
834 differentiation between estuarine and open coast ecotypes of a dominant ecosystem engineer.
835 *Marine and Freshwater Research* 70, 977. doi:10.1071/MF17392.
- 836 Coleman, M. A., Roughan, M., Macdonald, H. S., Connell, S. D., Gillanders, B. M., Kelaher, B. P.,
837 et al. (2011b). Variation in the strength of continental boundary currents determines

- 838 continent-wide connectivity in kelp: Boundary currents determine connectivity of kelp.
839 *Journal of Ecology* 99, 1026–1032. doi:10.1111/j.1365-2745.2011.01822.x.
- 840 Davison, I. R., and Pearson, G. A. (1996). Stress tolerance in intertidal seaweeds. *Journal of*
841 *Phycology* 32, 197–211. doi:10.1111/j.0022-3646.1996.00197.x.
- 842 Dayton, P. K. (1972). Toward an understanding of community resilience and the potential effects of
843 enrichments to the benthos at McMurdo Sound, Antarctica. in (Allen Press Lawrence, KS),
844 81–96.
- 845 Deutsch, C. A., Tewksbury, J. J., Huey, R. B., Sheldon, K. S., Ghalambor, C. K., Haak, D. C., et al.
846 (2008). Impacts of climate warming on terrestrial ectotherms across latitude. *Proceedings of*
847 *the National Academy of Sciences* 105, 6668. doi:10.1073/pnas.0709472105.
- 848 Dimartino, S., Mather, A. V., Alestra, T., Nawada, S., and Haber, M. (2015). Experimental and
849 computational analysis of a novel flow channel to assess the adhesion strength of sessile
850 marine organisms. *Interface Focus* 5, 20140059. doi:10.1098/rsfs.2014.0059.
- 851 Doblin, M., and Clayton, M. (1995). Effects of secondarily-treated sewage effluent on the early life-
852 history stages of two species of brown macroalgae: *Hormosira banksii* and *Durvillaea*
853 *potatorum*. *Marine Biology* 122, 689–698.
- 854 Dring, M. J., and Brown, F. A. (1982). Photosynthesis of intertidal brown algae during and after
855 periods of emersion: a renewed search for physiological causes of zonation. *Marine Ecology*
856 *Progress Series* 8, 301–308.
- 857 Dudgeon, S., Kübler, J., Wright, W., Vadas Sr, R., and Petraitis, P. S. (2001). Natural variability in
858 zygote dispersal of *Ascophyllum nodosum* at small spatial scales. *Functional Ecology* 15,
859 595–604.
- 860 Eckert, C. G., Samis, K. E., and Loughheed, S. C. (2008). Genetic variation across species'
861 geographical ranges: the central–marginal hypothesis and beyond. *Molecular Ecology* 17,
862 1170–1188. doi:10.1111/j.1365-294X.2007.03659.x.
- 863 Engel, C., Destombe, C., and Valero, M. (2004). Mating system and gene flow in the red seaweed
864 *Gracilaria gracilis*: effect of haploid–diploid life history and intertidal rocky shore landscape
865 on fine-scale genetic structure. *Heredity* 92, 289–298.
- 866 Excoffier, L., Laval, G., and Schneider, S. (2007). Arlequin (version 3.0): an integrated software
867 package for population genetics data analysis. *Evolutionary bioinformatics online* 1, 47–50.
- 868 Falkowski, P. G., and Raven, J. A. (2013). *Aquatic Photosynthesis*. Princeton University Press.
- 869 Faugeron, S., Martínez, E. A., Correa, J. A., Cardenas, L., Destombe, C., and Valero, M. (2004).
870 Reduced genetic diversity and increased population differentiation in peripheral and
871 overharvested populations of *Gigartina skottsbergii* (Rhodophyta, Gigartinales) in southern
872 Chile. *Journal of Phycology* 40, 454–462. doi:10.1111/j.1529-8817.2004.03114.x.

- 873 Ferreira, J. G., Arenas, F., Martínez, B., Hawkins, S. J., and Jenkins, S. R. (2014). Physiological
874 response of furoid algae to environmental stress: comparing range centre and southern
875 populations. *New Phytologist* 202, 1157–1172. doi:10.1111/nph.12749.
- 876 Foo, S. A., Dworjanyn, S. A., Poore, A. G. B., and Byrne, M. (2012). Adaptive capacity of the
877 habitat modifying sea urchin *Centrostephanus rodgersii* to ocean warming and ocean
878 acidification: performance of early embryos. *PLoS One* 7, e42497–e42497.
879 doi:10.1371/journal.pone.0042497.
- 880 Fredriksen, S., Filbee-Dexter, K., Norderhaug, K. M., Steen, H., Bodvin, T., Coleman, M. A., et al.
881 (2020). Green gravel: a novel restoration tool to combat kelp forest decline. *Scientific Reports*
882 10. doi:10.1038/s41598-020-60553-x.
- 883 Freeland, J. R., Petersen, S. D., and Kirk, H. (2011). *Molecular Ecology*. 2nd ed. Wiley-Blackwell.
- 884 Fusi, M., Giomi, F., Babbini, S., Daffonchio, D., McQuaid, C. D., Porri, F., et al. (2015). Thermal
885 specialization across large geographical scales predicts the resilience of mangrove crab
886 populations to global warming. *Oikos* 124, 784–795.
- 887 Galloway, L. F., Etterson, J. R., and McGlothlin, J. W. (2009). Contribution of direct and maternal
888 genetic effects to life-history evolution. *New Phytologist* 183, 826–838.
- 889 Genty, B., Briantais, J.-M., and Baker, N. R. (1989). The relationship between the quantum yield of
890 photosynthetic electron transport and quenching of chlorophyll fluorescence. *Biochimica et*
891 *Biophysica Acta (BBA) - General Subjects* 990, 87–92. doi:10.1016/S0304-4165(89)80016-9.
- 892 Gerard, V. A., and Du Bois, K. R. (1988). Temperature ecotypes near the southern boundary of the
893 kelp *Laminaria saccharina*. *Marine Biology* 97, 575–580. doi:10.1007/BF00391054.
- 894 Goudet, J. (1995). FSTAT (version 1.2): a computer program to calculate F-statistics. *Journal of*
895 *heredity* 86, 485–486.
- 896 Gunthorpe, L., Nottage, M., Palmer, D., and Wu, R. (1995). The development of a fertilisation
897 inhibition assay using gametes of the brown alga *Hormosira banksii*. *Australasian Journal of*
898 *Ecotoxicology* 1, 25–31.
- 899 Gurgel, C. F. D., Camacho, O., Minne, A. J. P., Wernberg, T., and Coleman, M. A. (2020). Marine
900 heatwave drives cryptic loss of genetic diversity in underwater forests. *Current Biology*.
901 doi:10.1016/j.cub.2020.01.051.
- 902 Hampe, A., and Petit, R. J. (2005). Conserving biodiversity under climate change: the rear edge
903 matters. *Ecology Letters* 8, 461–467. doi:10.1111/j.1461-0248.2005.00739.x.
- 904 Hanelt, D., Huppertz, K., and Nultsch, W. (1993). Daily course of photosynthesis and photoinhibition
905 in marine macroalgae investigated in the laboratory and field. *Marine ecology progress*
906 *series. Oldendorf* 97, 31–37.
- 907 Harley, C. (2008). Tidal dynamics, topographic orientation, and temperature-mediated mass
908 mortalities on rocky shores. *Marine Ecology Progress Series* 371, 37–46.
909 doi:10.3354/meps07711.

- 910 Hartnoll, R. G., and Hawkins, S. J. (1985). Patchiness and fluctuations on moderately exposed rocky
911 shores. *Ophelia* 24, 53–63.
- 912 Hawkins, S., Sugden, H., Mieszkowska, N., Moore, P., Poloczanska, E., Leaper, R., et al. (2009).
913 Consequences of climate-driven biodiversity changes for ecosystem functioning of North
914 European rocky shores. *Marine Ecology Progress Series* 396, 245–259.
915 doi:10.3354/meps08378.
- 916 Helmuth, B. (2009). From cells to coastlines: how can we use physiology to forecast the impacts of
917 climate change? *Journal of Experimental Biology* 212, 753–760. doi:10.1242/jeb.023861.
- 918 Helmuth, B., Broitman, B. R., Blanchette, C. A., Gilman, S., Halpin, P., Harley, C. D. G., et al.
919 (2006). Mosaic patterns of thermal stress in the rocky intertidal zone: implications for climate
920 change. *Ecological Monographs* 76, 461–479.
- 921 Helmuth, B., Harley, C. D. G., Halpin, P. M., O'Donnell, M., Hofmann, G. E., and Blanchette, C. A.
922 (2002). Climate change and latitudinal patterns of intertidal thermal stress. *Science* 298,
923 1015–1017. doi:10.1126/science.1076814.
- 924 Helmuth, B., Russell, B. D., Connell, S. D., Dong, Y., Harley, C. D., Lima, F. P., et al. (2014).
925 Beyond long-term averages: making biological sense of a rapidly changing world. *Climate*
926 *Change Responses* 1. doi:10.1186/s40665-014-0006-0.
- 927 Helmuth, B. S. T., and Hofmann, G. E. (2001). Microhabitats, thermal heterogeneity, and patterns of
928 physiological stress in the rocky intertidal zone. *The Biological Bulletin* 201, 374–384.
929 doi:10.2307/1543615.
- 930 Hobday, A. J., and Pecl, G. T. (2014). Identification of global marine hotspots: sentinels for change
931 and vanguards for adaptation action. *Reviews in Fish Biology and Fisheries* 24, 415–425.
932 doi:10.1007/s11160-013-9326-6.
- 933 Hoffmann, A. A., and Sgrò, C. M. (2011). Climate change and evolutionary adaptation. *Nature* 470,
934 479–485. doi:10.1038/nature09670.
- 935 Huey, R. B., Kearney, M. R., Krockenberger, A., Holtum, J. A., Jess, M., and Williams, S. E. (2012).
936 Predicting organismal vulnerability to climate warming: roles of behaviour, physiology and
937 adaptation. *Philosophical Transactions of the Royal Society B: Biological Sciences* 367,
938 1665–1679.
- 939 Huisman, J. M. (2019). *Marine Plants of Australia*. 2nd ed. Crawley Western Australia: UWA
940 Publishing.
- 941 Jensen, J. L., Bohonak, A. J., and Kelley, S. T. (2005). Isolation by distance, web service. *BMC*
942 *genetics* 6, 13.
- 943 Johannesson, K., Johannesson, B., and Lundgren, U. (1995). Strong natural selection causes
944 microscale allozyme variation in a marine snail. *Proc Natl Acad Sci USA* 92, 2602.
945 doi:10.1073/pnas.92.7.2602.

- 946 Jones, C. G., Lawton, J. H., and Shachak, M. (1994). Organisms as ecosystem engineers. *Oikos* 69,
947 373. doi:10.2307/3545850.
- 948 Kelly, M. W., Sanford, E., and Grosberg, R. K. (2012). Limited potential for adaptation to climate
949 change in a broadly distributed marine crustacean. *Proceedings of the Royal Society B:*
950 *Biological Sciences* 279, 349–356. doi:10.1098/rspb.2011.0542.
- 951 King, N. G., McKeown, N. J., Smale, D. A., and Moore, P. J. (2017). The importance of phenotypic
952 plasticity and local adaptation in driving intraspecific variability in thermal niches of marine
953 macrophytes. *Ecography* 41, 1469–1484. doi:10.1111/ecog.03186.
- 954 King, N. G., McKeown, N. J., Smale, D. A., Wilcockson, D. C., Hoelters, L., Groves, E. A., et al.
955 (2019). Evidence for different thermal ecotypes in range centre and trailing edge kelp
956 populations. *Journal of Experimental Marine Biology and Ecology* 514–515, 10–17.
957 doi:10.1016/j.jembe.2019.03.004.
- 958 Kramer, D. M., Johnson, G., Kiirats, O., and Edwards, G. E. (2004). New fluorescence parameters
959 for the determination of Q A redox state and excitation energy fluxes. *Photosynthesis*
960 *research* 79, 209.
- 961 Lough, J. M., and Hobday, A. J. (2011). Observed climate change in Australian marine and
962 freshwater environments. *Mar. Freshwater Res.* 62, 984–999. doi:10.1071/MF10272.
- 963 Lynch, M., and Walsh, B. (1998). *Genetics and analysis of quantitative traits*. Sinauer Sunderland,
964 MA.
- 965 Mabin, C., Johnson, C., and Wright, J. (2019). Physiological response to temperature, light, and
966 nitrates in the giant kelp *Macrocystis pyrifera*, from Tasmania, Australia. *Marine Ecology*
967 *Progress Series* 614, 1–19. doi:10.3354/meps12900.
- 968 Major, K. (Machalek), and Davison, I. (1998). Influence of temperature and light on growth and
969 photosynthetic physiology of *Fucus evanescens* (Phaeophyta) embryos. *European Journal of*
970 *Phycology* 33, 129–138. doi:10.1080/09670269810001736623.
- 971 Marshall, D. J. (2008). Transgenerational plasticity in the sea: Context-dependent maternal effects
972 across the life history. *Ecology* 89, 418–427.
- 973 Martínez, B., Arenas, F., Rubal, M., Burgués, S., Esteban, R., García-Plazaola, I., et al. (2012).
974 Physical factors driving intertidal macroalgae distribution: physiological stress of a dominant
975 furoid at its southern limit. *Oecologia* 170, 341–353. doi:10.1007/s00442-012-2324-x.
- 976 IPCC, Masson-Delmotte, V., Zhai, P., Portner, H.-O., Roberts, D., Skea, J., Shukla, P.R., Pirani, A.,
977 Moufouma-Okia, W., Pean, C., Pidock, R., Connors, S., Matthews, J.B.R., Chen, Y., Zhou,
978 X., Gomis, M.I., Lonnoy, E., Maycock, T., Tignor, M., Waterfield, T. (Eds.), 2018. Summary
979 for Policymakers, in: *Global Warming of 1.5°C. An IPCC Special Report on the Impacts of*
980 *Global Warming of 1.5°C above Pre-Industrial Levels and Related Global Greenhouse Gas*
981 *Emission Pathways, in the Context of Strengthening the Global Response to the Threat of*
982 *Climate Change, Sustainable Development, and Efforts to Eradicate Poverty*. World
983 Meteorological Organization, Geneva, Switzerland, p. 32.

- 984 McCoy, S. J., and Widdicombe, S. (2019). Thermal plasticity is independent of environmental
985 history in an intertidal seaweed. *Ecology and Evolution*. doi:10.1002/ece3.5796.
- 986 Miller, A. D., Coleman, M. A., Clark, J., Cook, R., Naga, Z., Doblin, M. A., et al. (2019). Local
987 thermal adaptation and limited gene flow constrain future climate responses of a marine
988 ecosystem engineer. *Evolutionary Applications*. doi:10.1111/eva.12909.
- 989 Mota, C. F., Engelen, A. H., Serrao, E. A., Coelho, M. A. G., Marbà, N., Krause-Jensen, D., et al.
990 (2018). Differentiation in fitness-related traits in response to elevated temperatures between
991 leading and trailing edge populations of marine macrophytes. *PLOS ONE* 13, e0203666.
992 doi:10.1371/journal.pone.0203666.
- 993 Muhlin, J. F., Engel, C. R., Stessel, R., Weatherbee, R. A., and Brawley, S. H. (2008). The influence
994 of coastal topography, circulation patterns, and rafting in structuring populations of an
995 intertidal alga. *Molecular Ecology* 17, 1198–1210. doi:10.1111/j.1365-294X.2007.03624.x.
- 996 Nelson, W. (2013). *New Zealand Seaweeds: An Illustrated Guide*. Te Papa Press.
- 997 Osborn, J. E. (1948). The structure and life history of *Hormosira banksii* (Turner) Decaisne. in (J.
998 Hughes, Printer), 47–71.
- 999 Paaijmans, K. P., Heinig, R. L., Seliga, R. A., Blanford, J. I., Blanford, S., Murdock, C. C., et al.
1000 (2013). Temperature variation makes ectotherms more sensitive to climate change. *Global
1001 change biology* 19, 2373–2380.
- 1002 Parmesan, C. (2006). Ecological and evolutionary responses to recent climate change. *Annual Review
1003 of Ecology, Evolution, and Systematics* 37, 637–669.
1004 doi:10.1146/annurev.ecolsys.37.091305.110100.
- 1005 Pearson, G. A., Lago-Leston, A., and Mota, C. (2009). Frayed at the edges: selective pressure and
1006 adaptive response to abiotic stressors are mismatched in low diversity edge populations.
1007 *Journal of Ecology* 97, 450–462. doi:10.1111/j.1365-2745.2009.01481.x.
- 1008 Pecl, G. T., Araújo, M. B., Bell, J. D., Blanchard, J., Bonebrake, T. C., Chen, I.-C., et al. (2017).
1009 Biodiversity redistribution under climate change: Impacts on ecosystems and human well-
1010 being. *Science* 355, eaai9214. doi:10.1126/science.aai9214.
- 1011 Poloczanska, E. S., Burrows, M. T., Brown, C. J., García Molinos, J., Halpern, B. S., Hoegh-
1012 Guldberg, O., et al. (2016). Responses of marine organisms to climate change across oceans.
1013 *Front. Mar. Sci.* 3. doi:10.3389/fmars.2016.00062.
- 1014 R Core Team (2020). *R: A language and environment for statistical computing*. Vienna, Austria: R
1015 Foundation for Statistical Computing Available at: <https://www.R-project.org/>.
- 1016 Reusch, T. B. H. (2014). Climate change in the oceans: evolutionary versus phenotypically plastic
1017 responses of marine animals and plants. *Evolutionary Applications* 7, 104–122.
1018 doi:10.1111/eva.12109.

- 1019 Reusch, T. B. H., Ehlers, A., Hämmerli, A., and Worm, B. (2005). Ecosystem recovery after climatic
1020 extremes enhanced by genotypic diversity. *PNAS* 102, 2826–2831.
1021 doi:10.1073/pnas.0500008102.
- 1022 Rice, W. R. (1989). Analyzing tables of statistical tests. *Evolution* 43, 223–225. doi:10.1111/j.1558-
1023 5646.1989.tb04220.x.
- 1024 Sagarin, R. D., and Gaines, S. D. (2002). The “abundant centre” distribution: to what extent is it a
1025 biogeographical rule? *Ecology Letters* 5, 137–147. doi:10.1046/j.1461-0248.2002.00297.x.
- 1026 Schiel, D. R. (2006). Rivets or bolts? When single species count in the function of temperate rocky
1027 reef communities. *Journal of Experimental Marine Biology and Ecology* 338, 233–252.
1028 doi:10.1016/j.jembe.2006.06.023.
- 1029 Schmidt, P. S., and Rand, D. M. (2001). Adaptive maintenance of genetic polymorphism in an
1030 intertidal barnacle: habitat-and life-stage-specific survivorship of *Mpi* genotypes. *Evolution*
1031 55, 1336–1344.
- 1032 Schonbeck, M., and Norton, T. A. (1978). Factors controlling the upper limits of fucoid algae on the
1033 shore. *Journal of Experimental Marine Biology and Ecology* 31, 303–313. doi:10.1016/0022-
1034 0981(78)90065-5.
- 1035 Schreiber, U. (2004). “Pulse-amplitude-modulation (PAM) fluorometry and saturation pulse method:
1036 an overview,” in *Chlorophyll a fluorescence* (Springer), 279–319.
- 1037 Sgrò, C. M., and Hoffmann, A. A. (2004). Genetic correlations, tradeoffs and environmental
1038 variation. *Heredity* 93, 241–248. doi:10.1038/sj.hdy.6800532.
- 1039 Smale, D. A., and Wernberg, T. (2013). Extreme climatic event drives range contraction of a habitat-
1040 forming species. *Proceedings of the Royal Society B: Biological Sciences* 280, 20122829.
1041 doi:10.1098/rspb.2012.2829.
- 1042 Smale, D. A., Wernberg, T., Oliver, E. C. J., Thomsen, M., Harvey, B. P., Straub, S. C., et al. (2019).
1043 Marine heatwaves threaten global biodiversity and the provision of ecosystem services.
1044 *Nature Climate Change* 9, 306–312. doi:10.1038/s41558-019-0412-1.
- 1045 Smolina, I., Kollias, S., Jueterbock, A., Coyer, J. A., and Hoarau, G. (2016). Variation in thermal
1046 stress response in two populations of the brown seaweed, *Fucus distichus*, from the Arctic
1047 and subarctic intertidal. *Royal Society Open Science* 3, 150429. doi:10.1098/rsos.150429.
- 1048 Somero, G. N. (2005). Linking biogeography to physiology: Evolutionary and acclimatory
1049 adjustments of thermal limits. *Frontiers in Zoology*, 9.
- 1050 Staehr, P. A., and Wernberg, T. (2009). Physiological responses of *Ecklonia radiata* (Laminariales)
1051 to a latitudinal gradient in ocean temperature. *Journal of Phycology* 45, 91–99.
1052 doi:10.1111/j.1529-8817.2008.00635.x.
- 1053 Stevens, G. C. (1989). The latitudinal gradient in geographical range: how so many species coexist in
1054 the tropics. *The American Naturalist* 133, 240–256. doi:10.1086/284913.

- 1055 Stillman, J. H., and Somero, G. N. (2000). A comparative analysis of the upper thermal tolerance
1056 limits of eastern pacific porcelain crabs, genus *Petrolisthes*: influences of latitude, vertical
1057 zonation, acclimation, and phylogeny. *Physiological and Biochemical Zoology* 73, 200–208.
1058 doi:10.1086/316738.
- 1059 Sunday, J. M., Bates, A. E., and Dulvy, N. K. (2012). Thermal tolerance and the global redistribution
1060 of animals. *Nature Climate Change* 2, 686–690. doi:10.1038/nclimate1539.
- 1061 Tatarenkov, A., Bergström, L., Jönsson, R. B., Serrão, E. A., Kautsky, L., and Johannesson, K.
1062 (2005). Intriguing asexual life in marginal populations of the brown seaweed *Fucus*
1063 *vesiculosus*: clonality in marginal fucoid populations. *Molecular Ecology* 14, 647–651.
1064 doi:10.1111/j.1365-294X.2005.02425.x.
- 1065 Teixeira, S., Pearson, G., Candeias, R., Madeira, C., Valero, M., and Serrão, E. (2016). Lack of fine-
1066 scale genetic structure and distant mating in natural populations of *Fucus vesiculosus*. *Marine*
1067 *Ecology Progress Series* 544, 131–142. doi:10.3354/meps11607.
- 1068 Thomas, C. D., Cameron, A., Green, R. E., Bakkenes, M., Beaumont, L. J., Collingham, Y. C., et al.
1069 (2004). Extinction risk from climate change. *Nature* 427, 145–148. doi:10.1038/nature02121.
- 1070 Valero, M., Destombe, C., Mauger, S., Ribout, C., Engel, C. R., Daguin-Thiebaut, C., et al. (2011).
1071 Using genetic tools for sustainable management of kelps: a literature review and the example
1072 of *Laminaria digitata*. 18.
- 1073 Van Oosterhout, C., Hutchinson, W. F., Wills, D. P. M., and Shipley, P. (2004). Micro-checker:
1074 software for identifying and correcting genotyping errors in microsatellite data. *Molecular*
1075 *Ecology Notes* 4, 535–538. doi:10.1111/j.1471-8286.2004.00684.x.
- 1076 Wernberg, T., Coleman, M. A., Bennett, S., Thomsen, M. S., Tuya, F., and Kelaher, B. P. (2018).
1077 Genetic diversity and kelp forest vulnerability to climatic stress. *Scientific Reports* 8, 1–8.
1078 doi:10.1038/s41598-018-20009-9.
- 1079 Wernberg, T., de Bettignies, T., Joy, B. A., and Finnegan, P. M. (2016). Physiological responses of
1080 habitat-forming seaweeds to increasing temperatures: seaweed responses to increasing
1081 temperature. *Limnology and Oceanography* 61, 2180–2190. doi:10.1002/lno.10362.
- 1082 Wickham, H. (2009). Elegant graphics for data analysis (ggplot2).
- 1083 Williams, S. L., and Dethier, M. N. (2005). High and dry: variation in net photosynthesis of the
1084 intertidal seaweed *Fucus gardneri*. *Ecology* 86, 2373–2379. doi:10.1890/04-1569.
- 1085 Wolf, J. B., and Wade, M. J. (2009). What are maternal effects (and what are they not)?
1086 *Philosophical Transactions of the Royal Society B: Biological Sciences* 364, 1107–1115.
1087 doi:10.1098/rstb.2008.0238.
- 1088 Womersley, H. B. S. (1987). *The marine benthic flora of southern Australia. Part II*. Adelaide: South
1089 Australian Government Printing Division.

- 1090 Wood, G., Marzinelli, E. M., Coleman, M. A., Campbell, A. H., Santini, N. S., Kajlich, L., et al.
1091 (2019). Restoring subtidal marine macrophytes in the Anthropocene: trajectories and future-
1092 proofing. *Mar. Freshwater Res.* doi:10.1071/MF18226.
- 1093 Zardi, G. I., Nicastro, K., Costa, J. F., Serrão, E., and Pearson, G. (2013). Broad scale agreement
1094 between intertidal habitats and adaptive traits on a basis of contrasting population genetic
1095 structure. *Estuarine, Coastal and Shelf Science* 131, 140–148.
- 1096 Zardi, G. I., Nicastro, K. R., Canovas, F., Costa, J. F., Serrão, E. A., and Pearson, G. A. (2011).
1097 Adaptive traits are maintained on steep selective gradients despite gene flow and
1098 hybridization in the intertidal zone. *PLoS One* 6, e19402–e19402.
1099 doi:10.1371/journal.pone.0019402.
- 1100 Zardi, G. I., Nicastro, K. R., Serrão, E. A., Jacinto, R., Monteiro, C. A., and Pearson, G. A. (2015).
1101 Closer to the rear edge: ecology and genetic diversity down the core-edge gradient of a
1102 marine macroalga. *Ecosphere* 6, art23. doi:10.1890/ES14-00460.1.

1103

In review

Figure legends

Figure 1: Map of locations from which *H. banksii* populations were collected from southeastern coast of Australia. Sea surface Temperature (SST) data reflect average SST taken from 2003 - 2015 from the Integrated Marine Observing System (IMOS) and used to illustrate coastal temperatures.

Figure 2: Mean (\pm SD) germling length after 5 days at 6 different temperatures (22, 24, 26, 28, 30 and 32 ° C) from the warm-edge and non-edge populations. Data from all crosses are pooled. (n = 54).

Figure 3: Percentage of *H. banksii* germlings reaching stage 1, 2, 3 and 4 or remaining in stage 0 after incubation at 6 different temperatures (22, 24, 26, 28, 30 and 32 ° C) for 5 days following fertilisation. Germlings are from warm-edge (A, C) and non-edge (B, D) populations. White columns represent ontogenic stage 0 (fertilisation through condensation of the chloroplasts); dark grey is stage 1 (protrusion of germling cell wall to create a pear shape which later develops into the rhizoid); medium grey is stage 2 (division of the germling germinating cell and elongation of a single rhizoid); light grey is stage 3 (elongation of the rhizoid coupled with secondary and tertiary rhizoid development); and black is stage 4 = paraphysis development on top of the germinating cell). Data represent pooled crosses (n= 40) amongst high shore (HS; A, B). low shore (LS; C, D).

Figure 4: Mean proportion of light energy dissipated by *H. banksii* germlings amongst three complementary PSII pathways: photochemistry (white bars, Y(II)); unregulated nonphotochemical quenching (grey bars, Y(NO) or regulated nonphotochemical quenching (black bars, Y(NPQ)) and potential photodamage, in germlings from warm-edge (A, C, E, G) and non-edge (B, D, F, H) populations from high on the shore (A, B, E, F) and low on the shore (C, D, G, H) after incubation for 120 h at three temperatures: 24, 28 and 32 °C. Photophysiological measurements were made under two irradiances - low light: 32 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ (A-D) and high light 113 $\mu\text{mol photons m}^{-2} \text{s}^{-1}$ (E-H).

Figure 5: Conceptual model of summarized data indicating how physiological traits differed amongst high and low on the shore as well as between warm-edge and non-edge locations. Thermal breadth was determined through arbitrarily setting a threshold of 80% of maximum germling length. Optimal temperature for growth was determined through pairwise comparisons. Ontogenic staged reached was the furthest developmental stage reached within 120 h. Photosynthetic efficiency stated are the overall changes in the proportion of energy dissipated in to unregulated nonphotochemical quenching (Y(NO) or regulated nonphotochemical quenching (Y(NPQ)) at both light intensities. Temperature data (air and SST) are presented for each location only.

Table 1: The number of individuals sampled (n), total number of alleles (a), mean number of alleles (\pm SD), unique alleles, observed heterozygosity (H_o) and expected heterozygosity (H_E) for *Hormosira banksii* for each height on the shore within each of the four locations.

Location	Height on the shore	n	Total number of alleles (a)	Mean number of alleles	Unique alleles	H_o	H_E
Angourie	High	27	26	2.60 ± 1.17	2	0.189	0.256
	Low	32	25	2.50 ± 0.97	0	0.166	0.250
Minnie Water	High	32	26	2.60 ± 1.07	0	0.201	0.269
	Low	32	26	2.60 ± 1.26	1	0.197	0.278
Pearl Beach	High	32	27	2.70 ± 0.67	0	0.299	0.310
	Low	32	24	2.40 ± 0.97	0	0.375	0.315
Bilgola Beach	High	27	26	2.60 ± 0.70	0	0.400	0.396
	Low	25	28	2.80 ± 0.79	2	0.403	0.399

Table 2: Pairwise F_{ST} estimates between all pairs of heights on the shore (L and H) and locations Bilgola Beach (BB), Pearl Beach (PB), Minnie Water (MW), and Angourie (ANG). Significant values are highlighted in bold and the adjusted p-value using Bonferroni Correction ($P < 0.002$). Grey shading shows comparisons between non-edge and warm-edge regions.

		Warm-edge				Non-edge			
		ANGL	ANGH	MWL	MWH	PBL	PBH	BBL	BBH
Warm-edge	ANGL								
	ANGH	0.002							
	MWL	0.066	0.105						
	MWH	0.032	0.075	0.000					
Non-edge	PBL	0.374	0.413	0.276	0.302				
	PBH	0.358	0.392	0.286	0.303	0.020			
	BBL	0.375	0.387	0.318	0.347	0.150	0.188		
	BBH	0.338	0.355	0.274	0.309	0.150	0.190	0.000	

Table 3: Results of analysis of variance of *Hormosira banksii* germling length 5 days (120 h) post fertilisation. Germlings from non-edge and warm-edge populations at low and high shore heights in 9 different genotypes were grown at 6 temperatures (22, 24, 26, 28, 30 and 32 °C). Location, height on the shore, and temperature are fixed factors, with male and female nested in shore height and location as a random factor. Results were achieved using 999 permutations and tested at a significance level of 0.05.

ANOVA Source	Germling Length		
	df	<i>F</i>	<i>P</i>
Location	1	0.556	0.749
Height	1	25.701	0.001
Temperature	5	159.160	0.001
Location x Height	1	3.871	0.029
Location x Temperature	5	22.746	0.001
Height x Temperature	5	5.501	0.001
Male (Location x Height)	8	6.524	0.004
Female (Location x Height)	8	1.840	0.163
Location x Height x Temperature	5	1.976	0.024
Temperature x Male (Location x Height)	40	1.680	0.027
Temperature x Female (Location x Height)	40	1.168	0.281
Male (Location x Height) x Female (Location x Height)	16	3.148	0.001
Temperature x Male (Location x Height) x Female (Location x Height)	76	2.927	0.001
Residuals	1080		

Bold denotes significance at $P < 0.05$

Table 4: Results of analysis of variance of the percent of *Hormosira banksii* germlings in each developmental stages, (stage 0-4) at 120 h after fertilisation. Germlings from non-edge and warm-edge populations, and from low and high on the shore in 9 different crosses were grown at 6 temperatures (22, 24, 26, 28, 30 and 32 °C). Location, height on the shore, and temperature are fixed factors, with male and female nested in shore height and location as a random factor. Probabilities were calculated using 999 permutations and tested at a significance level of 0.05.

ANOVA Source	Ontogenic Development										
	Stage 0			Stage 1		Stage 2		Stage 3		Stage 4	
	df	F	P	F	P	F	P	F	P	F	P
Location	1	0.712	0.596	3.838	0.026	1.754	0.178	8.523	0.003	8.079	0.002
Height	1	0.320	0.950	0.321	0.944	0.338	0.912	0.286	0.971	0.382	0.895
Temperature	5	10.595	0.001	10.354	0.001	22.597	0.001	10.430	0.001	9.302	0.001
Location x Height	1	0.502	0.784	3.528	0.031	3.276	0.046	0.584	0.714	0.605	0.73
Location x Temperature	5	2.232	0.009	1.222	0.226	0.829	0.653	4.438	0.001	4.055	0.001
Height x Temperature	5	1.091	0.341	1.336	0.189	0.484	0.986	1.064	0.388	1.110	0.338
Male (Location x Height)	8	2.803	0.042	1.638	0.192	0.667	0.730	3.932	0.012	4.512	0.004
Female (Location x Height)	8	2.301	0.060	2.395	0.055	2.709	0.048	1.470	0.239	1.537	0.218
Location x Height x Temperature	5	0.766	0.720	0.500	0.974	1.475	0.102	1.003	0.475	1.117	0.337
Temperature x Male (Location x Height)	40	0.850	0.718	0.977	0.537	1.567	0.050	1.799	0.010	1.952	0.006
Temperature x Female (Location x Height)	40	1.836	0.013	1.326	0.127	1.394	0.104	1.072	0.385	1.284	0.162
Male (Location x Height) x Female (Location x Height)	16	1.382	0.148	1.001	0.452	1.791	0.042	1.105	0.358	1.019	0.439
Residuals	80										

Bold denotes significance $P < 0.05$

Table 5: Results of analysis of variance of the temperature effects on chlorophyll-a fluorescence F_v/F_m , maximum quantum yield (F_v/F_m), complementary photosynthetic pathways of photosynthesis $Y(II)$, nonregulated nonphotochemical quenching $Y(NO)$, regulated nonphotochemical quenching $Y(NPQ)$ and high light (HL) to low light (LL) ratio of $Y(NPQ)$. Location, height on the shore and temperature are fixed factors. Probabilities were calculated using 9999 permutations and tested at a significance level of 0.05.

Source	F			F_v/F_m		HL:LL YNPQ	
	df	<i>F</i>	<i>P</i>	<i>F</i>	<i>P</i>	<i>F</i>	<i>P</i>
Location	1	12.17	0.003	4.39	0.050	0.09	0.790
Height	1	4.83	0.038	1.91	0.178	0.74	0.430
Temperature	2	1.61	0.222	5.40	0.011	9.35	< 0.001
Location x Height	1	0.69	0.417	0.09	0.771	0.84	0.397
Location x Temperature	2	0.82	0.446	0.89	0.433	2.18	0.123
Height x Temperature	2	0.06	0.938	0.02	0.981	0.46	0.671
Location x Height x Temperature	2	0.10	0.902	0.86	0.4373	0.18	0.872
Residuals	23						
Source	LY(II)			LY(NO)		LY(NPQ)	
	df	<i>F</i>	<i>P</i>	<i>F</i>	<i>P</i>	<i>F</i>	<i>P</i>
Location	1	0.46	0.500	3.20	0.091	5.44	0.028
Height	1	0.56	0.468	0.20	0.666	0.01	0.922
Temperature	2	8.76	0.001	6.50	0.006	0.89	0.433
Location x Height	1	1.37	0.255	0.73	0.405	0.01	0.927
Location x Temperature	2	0.62	0.562	0.40	0.681	0.09	0.918
Height x Temperature	2	0.26	0.783	0.05	0.947	0.04	0.964
Location x Height x Temperature	2	0.55	0.578	0.10	0.913	0.29	0.755
Residuals	23						
Source	HY(II)			HY(NO)		HY(NPQ)	
	df	<i>F</i>	<i>P</i>	<i>F</i>	<i>P</i>	<i>F</i>	<i>P</i>
Location	1	0.52	0.477	6.66	0.016	8.03	0.009
Height	1	1.47	0.252	0.001	0.982	0.58	0.444
Temperature	2	25.29	< 0.001	1.70	0.207	3.68	0.044
Location x Height	1	1.81	0.190	2.68	0.112	0.54	0.467
Location x Temperature	2	0.44	0.659	0.24	0.783	0.28	0.757
Height x Temperature	2	1.18	0.327	0.19	0.828	0.74	0.482
Location x Height x Temperature	2	1.11	0.346	0.21	0.809	0.19	0.831
Residuals	23						

Bold denotes significance at $P < 0.05$

Figure 1.JPEG

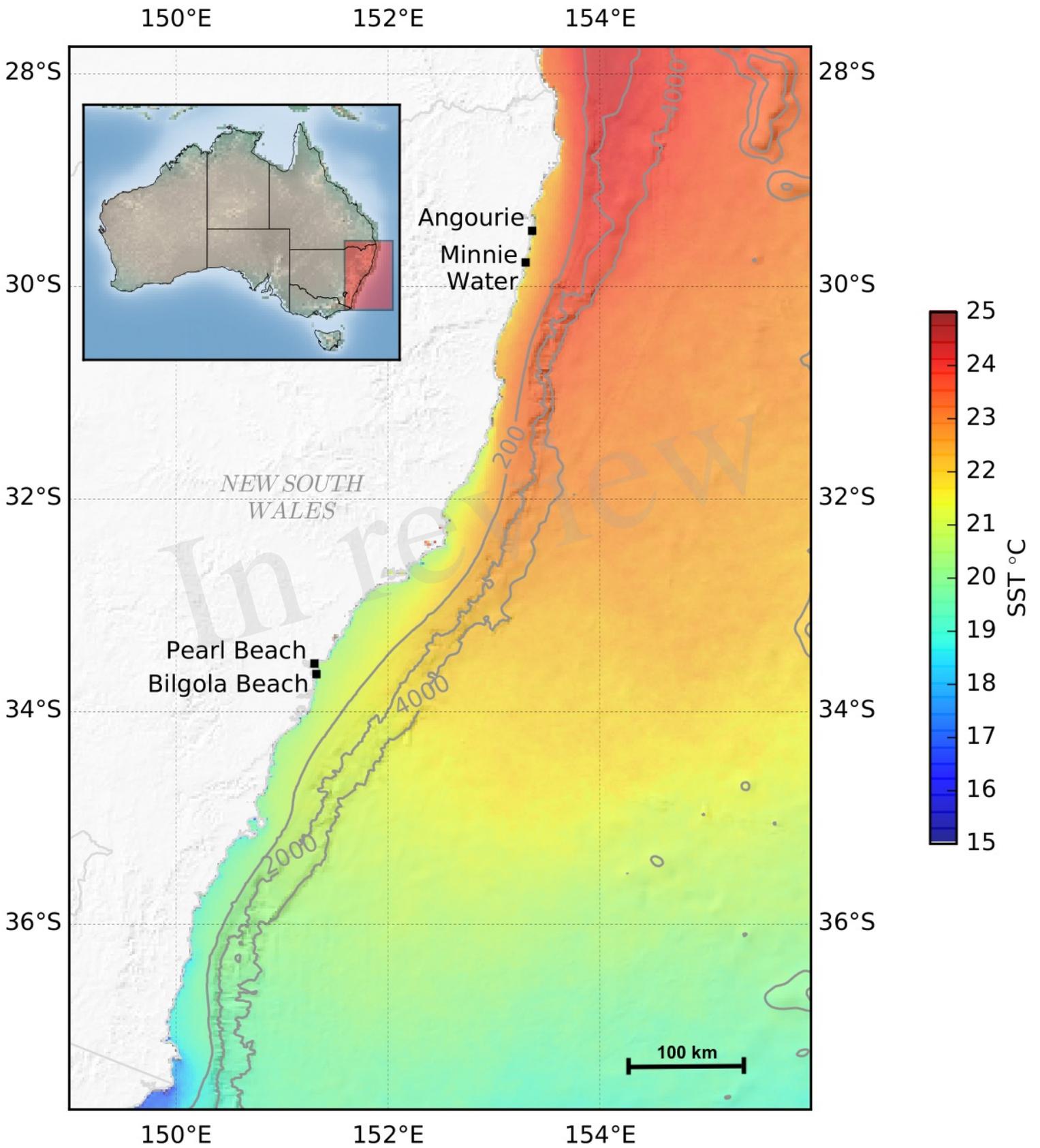


Figure 2.JPEG

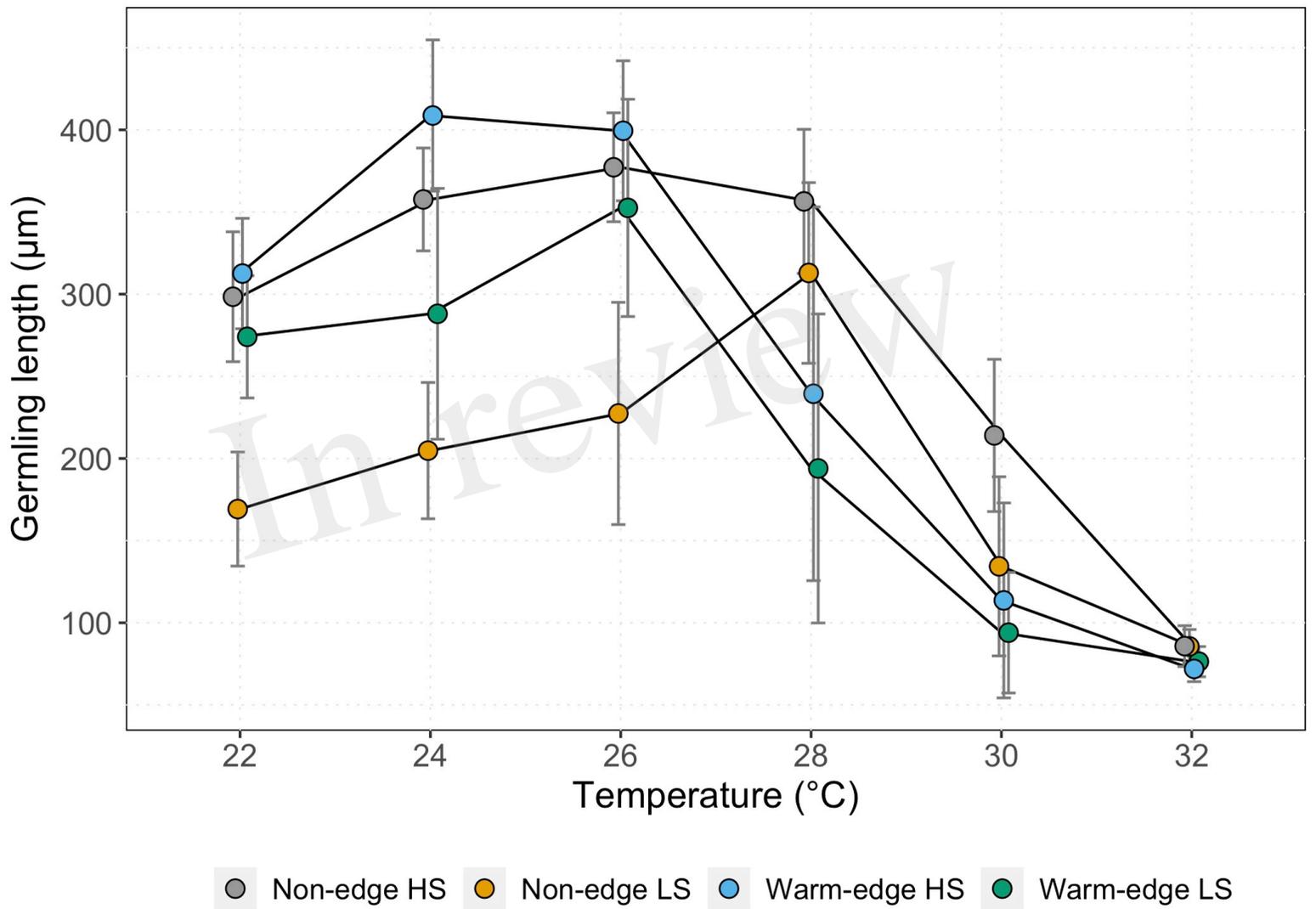


Figure 3.JPEG

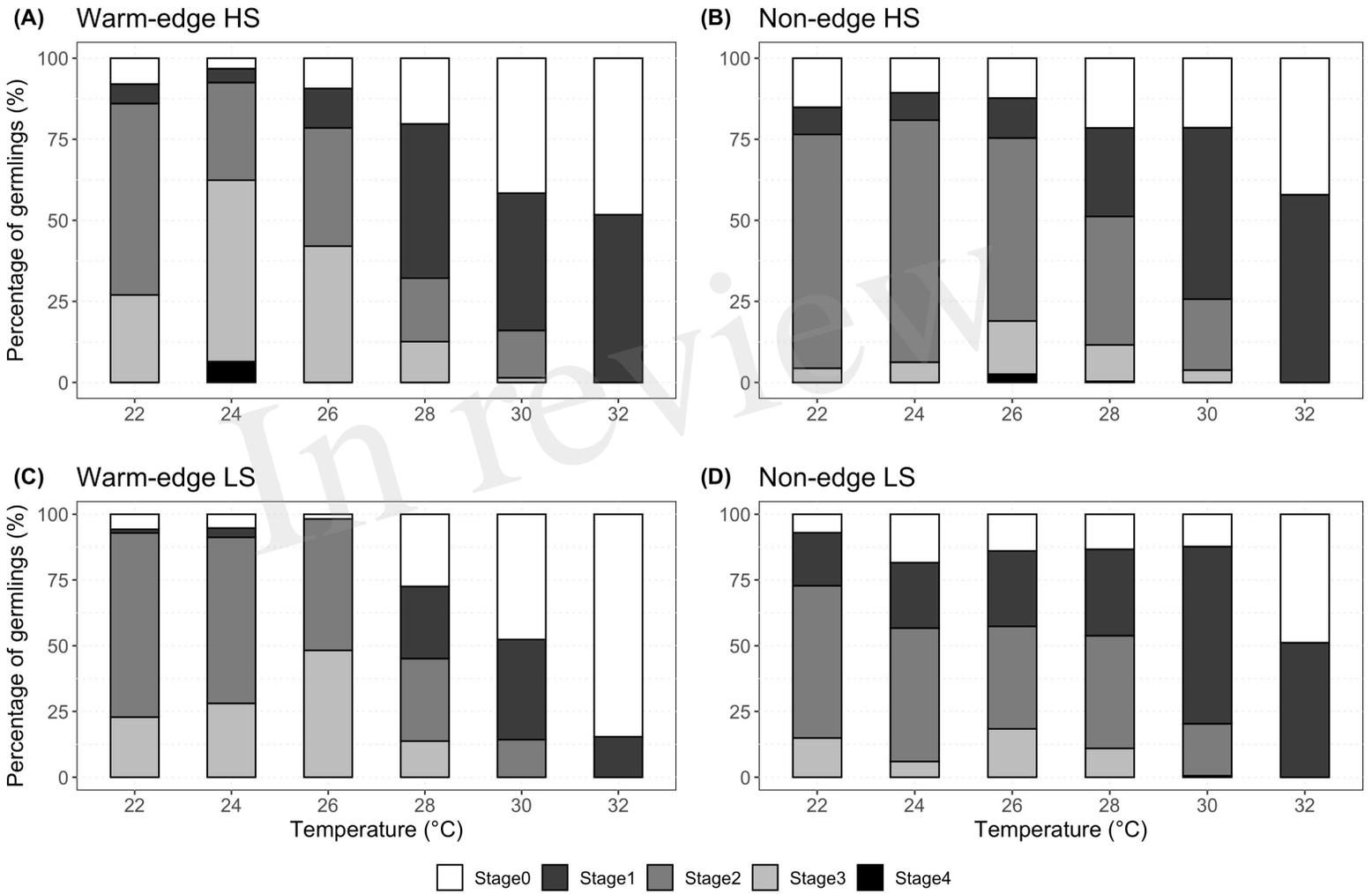


Figure 4.JPEG

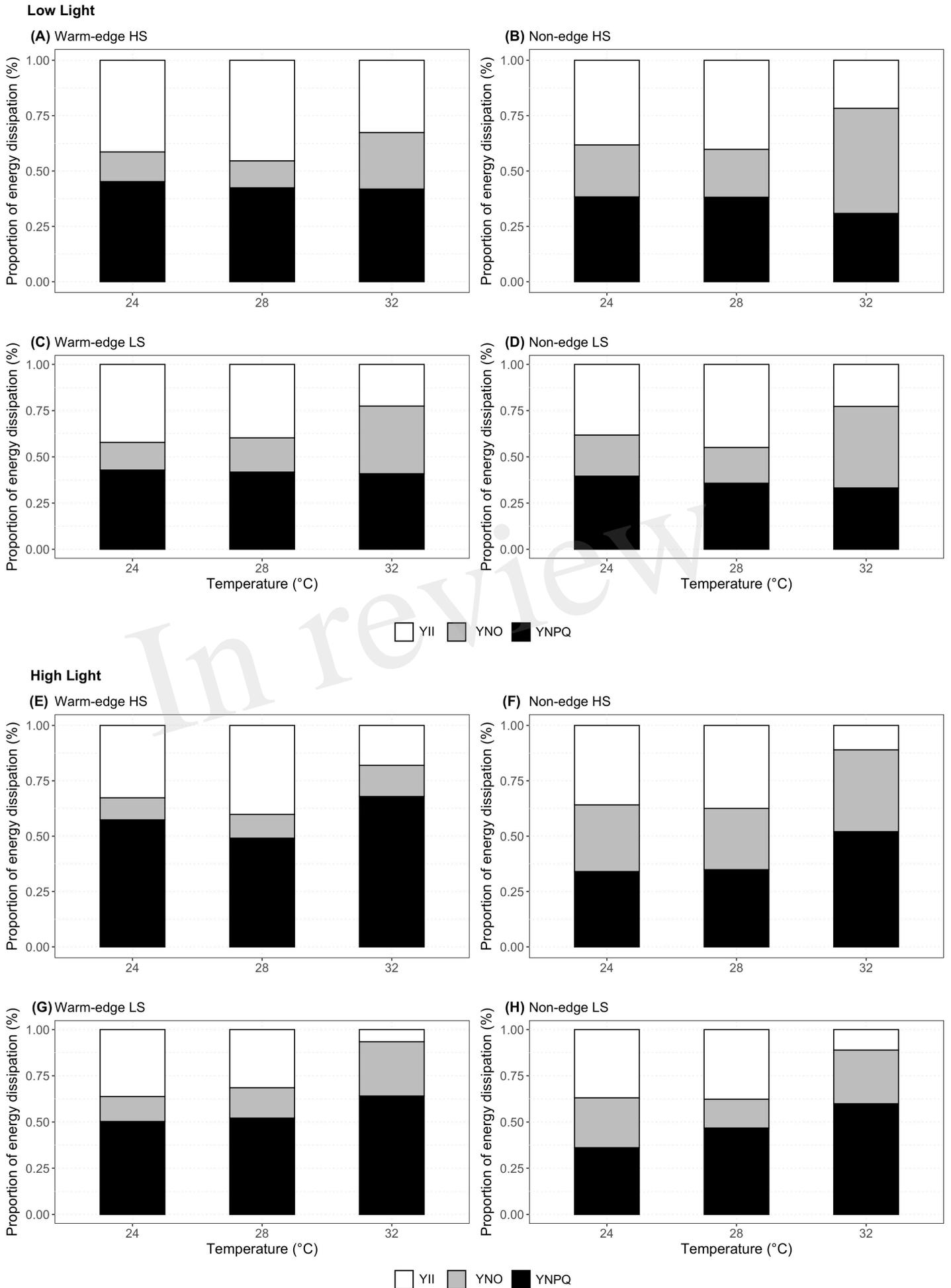


Figure 5.JPEG

