Photo-physiology and morphology reveal divergent warming responses in northern and southern hemisphere seagrasses

Abstract

A better understanding of species and population responses to thermal stress is critical to predict changes in their

distribution under warming scenarios. Seagrasses are a unique group of marine plants that play fundamental roles in

- marine environments and provide vital ecosystem services. Nevertheless, previous studies on seagrass thermal
- tolerance have focused exclusively on a handful of species, with the majority of these remaining virtually
- unexplored. Moreover, to date, no study has compared the response to thermal stress between northern and southern
- hemisphere seagrasses. Here, we conducted comparative mesocosm experiments using four seagrass species from
- the northern (i.e. Mediterranean: *Posidonia oceanica*, *Cymodocea nodosa*) and southern (i.e. Australia: *Posidonia*
- *australis* and *Zostera muelleri*) hemisphere as representative of two different life strategies, i.e. climax (*P. oceanica,*
- *P. australis*) and pioneer (*C. nodosa, Z. muelleri*). Plants acclimatized to the mesocosm conditions at ambient
- temperature (i.e. 26℃) during a five-week period, were exposed to a simulated marine heatwave (i.e. 32℃) for two
- weeks. Measurements of plant responses, including photo-physiology, morphology, and pigment content, were
- performed at the end of the warming exposure. Results showed that warming had no significant effects on
- photosynthetic performances of northern hemisphere seagrasses while negatively impacted their southern

hemisphere counterparts. Similarly, warming favored the growth of northern hemisphere plants, but strongly

inhibited the development of southern hemisphere species. Furthermore, photo-physiological and pigment content

- results suggested pioneer seagrasses better dealt with warming than climax species. Our study provides more
- insights into the field of seagrass ecology and yields potential implication for future seagrass conservation and restoration activities.
-
-
-
-
-
-
-
-
-
-
-
-

-
-

1. Introduction

- Seagrasses are a unique group of marine angiosperms occupying thousands of square kilometers along the shorelines
- of all continents, except for Antarctica (Short et al. 2007). Being among the most important habitat-forming species
- in the marine environment, seagrasses provide high primary productivity and nursery habitats for fish and
- invertebrates (Unsworth et al. 2019). Seagrasses help to stabilize coastal sediments, thus preventing erosion (Orth et
- al. 2006), and reduce pathogenic threats for humans, fish, and invertebrates (Lamb et al. 2017). Given that a vast
- majority of the human population inhabits coastal areas, seagrasses directly or indirectly influence the livelihoods of
- billions of people worldwide (Bertelli and Unsworth 2014). Economically, seagrass meadows are ranked among the
- most valuable ecosystems on Earth, contributing nearly \$2 trillion annually in ecosystem services (Waycott et al.
- 2009; Costanza et al. 2014). Moreover, seagrasses represent the most significant natural carbon sink on our planet
- (Fourqurean et al. 2012; Macreadie and Hardy 2018), which can potentially be a part of future ocean solutions
- helping us to mitigate the negative effect of greenhouse gas emissions (Gattuso et al. 2018).
- Nonetheless, seagrasses have undergone a global decline mainly due to human-induced environmental changes
- 66 (Orth et al. 2006; Waycott et al. 2009). Across the globe, human activities are wiping out over 100 km^2 of seagrass
- meadows per annum. As a result, nearly 29% of their areal extent has been lost since 1879 (Waycott et al. 2009). For
- instance, the Mediterranean endemic species, *Posidonia oceanica*, already lost approximately 13-50% of its total
- areal extent since the mid-nineteenth century (Telesca et al. 2015). The decline of *P. oceanica* meadows is likely to
- continue as this species' ecological functions have even been predicted to go extinct by the end of this century
- (Marbà and Duarte 2010; Chefaoui et al. 2018).
- Among human-induced stressors to seagrasses, ocean warming appears to be a key player (Nguyen et al. 2021).
- While ocean warming is commonly known as the gradual increment in the mean of seawater temperature, it also
- affects species in the form of extreme climatic events (i.e. marine heatwaves). Marine heatwaves (MHWs) are
- 75 defined as abnormal warming events that last for over 5 days, with temperatures exceeding the $90th$ percentile of a
- three-decade historical baseline database (Hobday et al. 2016). For seagrasses (but also for other benthic organisms),
- the impact of MHWs is generally more detrimental than increases in mean seawater temperatures because seagrasses
- are generally susceptible to sudden thermal changes (Smale et al. 2019). In fact, MHWs have produced devastating
- consequences for seagrasses and associated communities across the globe (Coma et al. 2009; Harley et al. 2012;
- Wernberg et al. 2016; Smale 2020). MHWs were the main cause of massive die-off events of seagrass species
- including *Zostera marina* (Jarvis et al. 2014) and *Amphibolis antarctica* (Arias-Ortiz et al. 2018; Strydom et al.
- 82 2020). MHWs not only cause the decline of seagrass meadows and all their ecological services but also fostered the
- release of greenhouse gases into the atmosphere, consequently contributing to the on-going global warming (Arias-
- Ortiz et al. 2018; Macreadie and Hardy 2018; Salinas et al. 2020).
- With insights from previous studies, we now know that the capacity to cope with warming (or especially MHWs)
- varies among different seagrass species (Marín-Guirao et al. 2016; Collier et al. 2017; Nguyen et al. 2020b), but also

87 among populations of the same species from contrasting thermal environments (e.g. Bergmann et al. 2010; Winters

88 et al. 2011; Marín-Guirao et al. 2018; Marín-Guirao et al. 2019). However, to date, the tolerance to anomalous

thermal events of the majority of seagrasses (especially in the region of southeast Asia and northern Australia, a

hotspot of seagrass diversity) are yet to be investigated (see Nguyen et al. 2021 for a complete review). Indeed, to

- the best of our knowledge, no study has compared the responses of northern versus southern hemisphere seagrasses
- to warming.

Along the ecological succession, plants can be divided in pioneer species (i.e. fast-growing, often with small body

size and annual) and climax species (i.e. slow-growing, long-lived, often with large body size and perennial), with

different biological characteristics and ecological role (Glenn-Lewin et al. 1992). Likewise, some seagrass species

can be classified as climax (e.g. *Posidonia oceanica*, *P. australis*, *Zostera marina*, *Thalassia testudinum*) while

others as pioneer (e.g. *Cymodocea nodosa*, *Z. muelleri*). The contrasting characteristics between the two groups

underpin large variations in the number and type of ecosystem services they provide. In seagrasses, most of their

ecological services (e.g. sediment stabilization, nursery habitat, and blue carbon burial, etc.) depend upon their

- physical structure and primary productivity and, hence, climax seagrasses are considered more ecologically valuable
- 101 than pioneer ones.

Studies from terrestrial plants have documented dissimilarities in response to environmental stressors between

climax versus pioneer plants. For instance, studies from the Brazilian Atlantic Forest showed that pioneer trees were

more tolerant against oxidative stress than climax plants (Favaretto et al. 2011; Brandão et al. 2017; Esposito et al.

2018). In line with these studies from the southern hemisphere, a study from the Mediterranean region

experimentally tested the responses of carbon assimilation under summer stress conditions (water deficits, high

light, and temperature) in four Mediterranean trees, including climax and pioneer species (Faria et al. 1998). This

study indicated that, although both groups of trees suffered a decline in their photosynthetic capacities, the climax

plants exhibited the lowest photosynthetic rates and the highest proportion of carotenoids to chlorophyll (i.e. an

- indicator of photo-protective mechanism activated under stressful conditions) while pioneer species maintained higher photosynthetic rates (Faria et al. 1998). Hence, environmental stressors can impact more strongly climax
- species, favoring the persistence of less complex and stable ecosystems, and providing less valuable ecosystem
- services. This is true also for seagrasses (Johnson et al. 2003; Hyndes et al. 2016; Shields et al. 2019), where it

appears essential to assess the response to stress of both climax and pioneer species, in order to support timely and

- effective conservation and/or restoration actions. Few studies have experimentally compared the response to
- warming of climax and pioneer seagrass species (e.g. see Masini and Manning 1997; Seddon and Cheshire 2001;

Campbell et al. 2006; Collier and Waycott 2014; Marín-Guirao et al. 2016, 2018; Collier et al. 2017; Tutar et al.

2017). Most of them suggested that pioneer species are more thermal tolerant than climax ones. These studies

demonstrated that the fast-growing pioneer seagrasses exhibited a better ability to maintain unaltered plant carbon

- balances through improved photosynthetic thermal stability and performance as well as by inhibiting respiratory
- carbon consumption. Moreover, through a higher morphological plasticity, pioneer species can modify their plant
- architecture by increasing the above-ground (photosynthetic)/below-ground (non-photosynthetic) biomass ratio to
- deal with thermal stress (Collier et al. 2017; Marín-Guirao et al. 2018), and have also an increased ability to activate
- antioxidant defense mechanisms to protect themselves from heat-stress induced oxidative damage (Tutar et al.
- 2017). Notwithstanding these evidences, the number of studies on this topic, especially on species with overlapping
- geographical distribution, remains scarce and deserves more effort.
- In the present study, four seagrass species including *P. oceanica* and *C. nodosa* from the Mediterranean (northern
- hemisphere) and *P. australis* and *Z. muelleri* from South East Australia (southern hemisphere) were selected for a
- comparative study of their responses to warming. Plants were collected in the same seasonal conditions (i.e. late
- summer-early autumn: Mar-May in the southern hemisphere and Sept-Nov in the northern hemisphere) from both
- geographic areas and two mesocosm experiments were conducted following the same design. Our study represents a
- unique opportunity to compare (1) two climax species of the genus *Posidonia* (*P. oceanica* and *P. australis*) with
- similar characteristics and ecological functions but distributed in the two hemispheres and (2) two couples of
- climax-pioneer species from both hemispheres (*P. oceanica* vs. *C. nodosa* and *P. australis* vs. *Z. muelleri*). On the
- first hand, we hypothesized that the responses to warming of the two *Posidonia* species (i.e. *P. oceanica* and *P.*
- *australis*) would be different because sampled populations live under a different thermal regime (i.e. 13 28℃ for
- *P. oceanica*; Fig. 1b and 17 26℃ for *P. australis*; Fig. 1c) and because the species thermal ranges are also different
- (i.e. 8 30℃ for *P. oceanica* and 12 28℃ for *P. australis*; Fig. S1). Additionally, we note that the collection sites
- of the Australian seagrasses in this study did not fall into any Mediterranean-climate regions (see Cowling et al.
- 1996 for a map of Mediterranean-climate regions and Fig. 1a for sample collection sites). On the other hand, in both
- hemispheres, the climax seagrasses are expected to suffer more from thermal stress than their pioneer counterparts.
-
-
-
-
-
-
-

Figure 1: Sample collection sites (**a**) and temperature conditions at collection sites (**b**, **c**). (**b**) Monthly average sea surface

- temperature in Ischia, Italy (Mediterranean sites: **1** & **2**). (**c**) Monthly average sea temperature in Port Stephens, NSW, Australia
- (Australian sites: **3** & **4**). Data were taken from World sea temperature of 2020 [\(https://www.seatemperature.org/](https://www.seatemperature.org/) data assessed
- 152 on 29^{th} Nov 2020).
- **2. Materials and methods**
- **2.1. Experiment 1: Northern hemisphere experiment**
- *2.1.1. Targeted species and Plant collection*
- *Posidonia oceanica* is endemic to the Mediterranean Sea (see Fig. S1 for the species distribution) and forms large
- and dense monospecific meadows on rocks and sandy seabed ranging from shallow water (less than 1 m) down to 45
- meter's depth (Procaccini et al. 2003). It ranks as one of the slowest-growing plants and among the longest-living
- plants on Earth with single clones extending over kilometers and living for hundreds to thousands of years (Arnaud-
- Haond et al. 2012). *Cymodocea nodosa* distributes throughout the Mediterranean Sea and extends also in nearby
- subtropical Atlantic areas (see Fig. S1 for *C. nodosa* distribution). *C. nodosa* is a relative fast-growing species,
- commonly found in shallow waters in both sandy and mud substrates where it forms both monospecific and mixed
- meadows with other seagrass species (den Hartog 1970; Guidetti et al. 1998). Hereafter, we use Med-Climax for *P.*
- *oceanica* and Med-Pioneer for *C. nodosa*.
- Plant fragments (i.e. ramets) of Med-Climax (40°44.020'N, 13°58.039'E at 5-6 m depth; Fig. 1a-1) and Med-Pioneer
- (40°47.021'N, 14°04.404'E at 8-10 m depth; Fig. 1a-2) were haphazardly collected by SCUBA diving in the Gulf of
- 167 Naples (Italy) on the $18th$ September 2019. To reduce the likelihood of sampling the same genotype twice, plants
- were collected at a minimum distance of 10 m from each other. Both *P. oceanica* and *C. nodosa* experience a wide
- species thermal range from 8℃ in the winter (especially in the northern Adriatic Sea) to 30℃ in the summer
- (especially in the eastern Mediterranean Sea), see Fig S1 for more details. The two populations used in this study
- came from a similar thermal condition (i.e. 13-28℃, see Fig. 1b) which falls in the middle of the species thermal
- range, therefore excluding the existence of a potential range-edge effects for the selected populations. After
- collection, plants were kept in dark in a cooler filled with seawater at ambient temperature and transported to a
- benthic mesocosm facility at the Stazione Zoologica Anton Dohrn (SZN), Napoli, Italy (see Ruocco et al. 2019b for
- a detailed description of the experimental system). Light intensity, salinity, and seawater temperature were measured
- at the time of plant sampling for setting up the experimental system.

Figure 2: Temperature profile during the two experiments.

2.1.2. Experimental system

 Once at the SZN experimental facility, twelve plant fragments (i.e. ramets) of each species composed by horizontal rhizomes of similar size and a similar number of interconnected vertical shoots (~ 10 shoots) were selected to standardize the experiment. Med-Climax ramets were transplanted in six plastic pots (i.e. two ramets per pot) filled with coarse carbonate sediments as described in Ruocco et al. (2019b), while Med-Pioneer ramets were transplanted in twelve plastic pots (i.e. one ramet per pot) filled with natural sediments from the collection site as described in Marín-Guirao et al. (2018). After transplantation, pots of each species were randomly allocated into six 500L- aquaria with filtered and UV-treated natural seawater from a close area; each aquarium containing two ramets of the Med-Climax species and two ramets of the Med-Pioneer species. Transplant pots were distributed within aquaria to avoid crossed-species shading and their distance from the light source adjusted to reproduce similar light intensities 188 to those measured at their collection sites (i.e. max noon irradiance: 300 and 200 µmol photons $m^{-2} s^{-1}$ above the leaf canopy for Med-Climax and Med-Pioneer, respectively). A 12h:12h light:dark photoperiod was applied, starting from 7:00 a.m., and progressively increasing to the maximum irradiance at 13:00 before a gradual reduction until

- 191 dark at 19:00. Water temperature was measured automatically every 10 min using HOBO Pendant®
- Temperature/Light 64K Data Logger (Onset, USA) and manually checked twice a day with WTW Cond 3310 Set 1
- 193 (Xylem Analytics, Germany). Seawater salinity of 37.5 ± 0.2 was kept constant throughout the experiment through
- regular additions of purified water. Seawater quality was controlled via continuous mechanical filtration, weekly-
- UV sterilizations, and partial renewals. An introductory video was made for this experiment and could be found on
- 196 the website of Dr. Gabriele Procaccini's Laboratory [\(https://gpgroupszn.wixsite.com/website;](https://gpgroupszn.wixsite.com/website) video: EpicSea2019).
- **2.2. Experiment 2: Southern hemisphere experiment**

2.2.1. Targeted species and Plant collection

- *Posidonia australis* is a slow-growing species found on sandy sediment between 1 to 15 m (Trautman and
- Borowitzka 1999). This species is distributed along the southern half of Australia, from Shark Bay in Western
- Australia to Port Macquarie in New South Wales, and along the northern coast of Tasmania (Fig. S1). With this
- distribution, *P. australis* exhibits the narrowest species thermal range (i.e. 12 28℃) of the four seagrass species
- included in this study. *Zostera muelleri* is a fast-growing species, commonly found in shallow water (< 4 m depth)
- on different sediments including fine sand, mud and others (Larkum et al. 2018). *Z. muelleri* is distributed along the
- eastern coast of Australia, Tasmania Kangaroo Island, Lord Howe Island, and New Zealand (Fig. S1; Waycott et al.
- 2004). The species' thermal range varies between 9℃ in the winter season (especially in the southern Tasmania
- Island and New Zealand) and 31℃ during summer (see Fig. S1). Hereafter, we use Aus-Climax for *P. australis* and
- Aus-Pioneer for *Z. muelleri*.
- 209 Ramets of Aus-Climax and Aus-Pioneer were collected, at distances > 25 m one from another to reduce the chance
- 210 of sampling the same genotype twice. Plant fragments (i.e. ramets) were collected during low tides at \sim 70 cm depth
- 211 at Port Stephens (PS), New South Wales (NSW), Australia (32°43'07.4" S 152°10'35.9" E; Fig. 1a) on the 19th of
- 212 March 2019 and at Church Point (CP), NSW, Australia $(33^{\circ}38'46.8''S\ 151^{\circ}17'11.9''E$; Fig. 1a) on the 23^{rd} of March
- 2019, respectively. Temperature ranges at both sampling locations between 17℃ in winter and 26℃ in summer. The
- maximum summer temperature is below the upper limits of species thermal ranges (i.e. 28℃ and 31℃ for *P.*
- *australis* and *Z. muelleri*, respectively; Fig. S1). Neither one of the two populations analyzed were at the edge of the
- species distribution range, but we acknowledged, in the interpretation of results, the difference in the upper limit of
- species thermal ranges. Light intensity and salinity were also measured at the time of sample collection. Plant
- materials were brought to the seagrass mesocosm facility at the University of Technology Sydney (UTS) soon after
- collection. A detailed description of the experimental system can be found in Nguyen et al. (2020b).

2.2.2. Experimental system

- As soon as arrived at UTS, twelve ramets of each species with a similar number of shoots (i.e. 8-10 shoots) were
- selected and transplanted in individual plastic pots (i.e. one ramet per pot) filled with mini pebbles. Subsequently,
- pots were randomly distributed in tanks of the mesocosm facility: six 60-L aquaria for housing Aus-Climax pots and
- six 40-L aquaria for Aus-Pioneer pots (i.e. two ramets per aquarium). For both species, the irradiance level was set
- 225 with a max. noon irradiance of 350 µmol photons $m^{-2} s^{-1}$ at canopy height and a 12 h:12 h light:dark photo-period.
- Light cycle started from 7:30 a.m., with light levels progressively increasing to the maximum irradiance at 12:30
- p.m. and kept for 2 hours, before a progressive reduction until dark at 7:30 p.m. Water temperature was measured
- automatically every 30 min using iButton data logger (iButtonLink, USA) and manually checked twice a day using a
- 229 digital thermometer (FLUKE 52II, USA). During the experiment, purified water was added periodically to maintain
- 230 constant seawater salinity of 34 ± 0.2 similar to those in the fields. Approximately one third of seawater was renewed
- weekly in each aquarium to maintain water quality.

2.3. Experimental design

- Both the northern and the southern hemisphere experiments shared the same experimental design. After
- transplantation and allocation within aquaria, plants of the four studied species were allowed to acclimate at 26℃,
- which is similar to the seawater temperatures recorded during plant collection at the four studied populations (i.e. in
- every case the difference was lower than 1℃). After a 5-week acclimation period, temperature in half of the aquaria
- containing each seagrass species (i.e. 3 aquaria) was progressively increased up to 32^oC at a heating rate of 1^oC day⁻¹
- to simulate a marine heatwave (MHW); whereas the temperature in the rest of aquaria was maintained throughout
- 239 the experiment (Fig 2). Therefore, for each species, three tanks were randomly assigned to heat treatment (TM) and
- 240 other three remained as controls (CT). Seagrass responses were analyzed at the end of the MHW exposure, which
- lasted 12 and 10 days in the northern and southern hemisphere experiments, respectively. The aquarium is the true
- experimental unit for each seagrass species and variable, so that measurements performed on plants of the same
- aquarium (i.e. 'pseudo replicates') were averaged to obtain an independent replicated value. Therefore, the number
- 244 of replicates used in statistical tests was $n = 3$.

2.4. Chlorophyll *a* **fluorescence**

- Identical Diving-PAM fluorometers (WALZ, Germany) were used to determine the photo-physiological responses
- of the four studied seagrass species (Med-Climax, Med-Pioneer, Aus-Climax, and Aus-Pioneer) following the
- methodology described in Marín-Guirao et al. (2013). To standardize the procedure, two chlorophyll *a* fluorescence
- 249 measurements were conducted on the same middle portion of the second youngest leaf of each plant (Ruocco et al.,
- 2019a). Measurements included (*a*) maximum quantum yield (*Fv*/*Fm*) of photosystem II (PSII) measured on night
- dark-adapted plants (around 6:00 7:00 am before the light cycle started), (*b*) Effective quantum yield (∆*F*/*Fm'*)
- measured on light-adapted plants (around 12:30 13:30 while the irradiances were highest) and (*c*) Non-
- photochemical quenching (*NPQ*) calculated by using the method described elsewhere (Maxwell and Johnson 2000)
- to estimate the amount of photosynthetic energy lost as heat (i.e. the photo-protective mechanisms associated to the
- xanthophyll-cycle pigments; Marín-Guirao et al. 2013).

2.5. Plant growth

- For both experiments, plant growth measurements were performed by adopting the leaf marking method (Zieman
- 1974). Two plants from each aquarium and species were marked at the same position above the ligule at the end of
- the acclimation period and subsequently collected at the end of the heatwave to measure leaf elongation (mm). Then,

 the newly developed leaf segments were cleaned of epiphytes and dehydrated at 70℃ for 24h before being weighted 261 to assess biomass production (mg Dry weight).

2.6. Pigment content

263 At the end of the experiments, two plants of each species and from each aquarium were collected for the analysis of leaf pigment content. Approximately 50 mm of leaf tissue from the middle portion of the second youngest leaf of climax species (Med-Climax and Aus-Climax) and the whole second youngest leaf of pioneer species (Med-Pioneer and Aus-Pioneer) was used for the analysis. Epiphytes were immediately removed from the collected material, which was then kept on ice in darkness until further processing. Pigment extractions were done on the same day of sample collection. After weight measurements, samples were homogenized in liquid nitrogen by using pestles and mortars before being transferred into 1.5 mL tubes filled with 1 mL of 100% methanol. Thenceforward, samples 270 were kept in complete darkness at 4°C for 8 hours before centrifugation. 200 μ L of the extracted solution was used to determine the absorbance at 4 different wavelengths (i.e. 470, 652, 665, and 750 nm) by the mean of microplate 272 readers (TECAN Infinite[®] M1000PRO, Switzerland) to calculate chlorophyll *a*, chlorophyll *b*, chlorophyll $a+b$, chlorophyll *b*/*a* molar ratio and total carotenoids. Pigments were calculated using equations from Wellburn (1994) 274 after converting microplate readings into 1cm cuvette readings following Warren (2008). Finally, results were normalized to milligrams of fresh weight.

2.7. Statistical analyses

 The response of seagrasses to experimental conditions was assessed using a three-way *ANOVA* (*n* = 3), including the following factors: Hemisphere (2 levels: northern and southern, fixed), Life-strategy (2 levels: climax and pioneer, fixed), and Treatment (2 levels: control and treatment, fixed). Cochran's C test was used to test homogeneity of variances and data were square root transformed when necessary. Data were analyzed even when homogeneity of variances could not be achieved, as *ANOVA* is robust for this kind of assumption when the sizes of samples are equal 282 (Underwood et al. 1997). However, in this case, the significance was judged more conservatively $(p < 0.01)$ when interpreting results to reduce the livelihood of Type I error (which is inflated by heterogeneous variances). For each measurement, Student-Newman-Keuls (SNK) *post-hoc* tests were used to identify significant differences between (*1*) control versus treatment plants of each Hemisphere, each Life-strategy, (*2*) northern versus southern plants of each Life-strategy, each Treatment, and (*3*) climax versus pioneer plants of each Hemisphere, each Treatment. All statistical analyses were conducted in R-studio v.1.2.5033 (R Core Team 2018) using package *GAD* (Sandrini-Neto and Camargo 2014).

Graphs were made with R-studio using package *ggplot2* (Wickham 2009).

3. Results

- **3.1. Photo-physiological responses**
-

294 Figure 3: Boxplot graphs present photo-physiological results at the end of the experiment $(n = 3)$. (a) Maximum quantum yield (*Fv*/*Fm*), (**b**) Effective quantum yield (∆*F*/*Fm'*), and (**c**) Non-Photochemical quenching (*NPQ*). Asterisks indicate statistical

differences between control and treatment within each species (Student-Newman-Keuls *post-hoc* test, ** *p* < 0.01, *** *p* < 0.001,

more details can be found in Supplementary data, Table S1).

- 3.1.1. Northern versus southern hemisphere seagrasses
- Warming had strong impacts on the southern hemisphere seagrasses while did not result in any significant changes
- 300 for the northern hemisphere plants (ANOVA: H×T, $F(1,16) = 12.030, p < 0.01$) (Fig. 3). Warming significantly
- reduced *Fv/Fm* of the Aus-Climax plants (SNK test for 'H:L:T' among 'T' within 'H:L': *p* < 0.001; Fig. 3, Table
- S1), while slightly impacted the Med-Climax plants. Similarly, the Aus-Climax plants enhanced their *NPQ* with
- warming (SNK test for 'H:L:T' among 'T' within 'H:L': *p* < 0.001; Fig. 3, Table S1) while Med-Climax's *NPQ*
- remained relatively unchanged (Fig. 3). Both Aus-species dramatically lowered their ∆*F*/*Fm'* as a result of thermal
- stress (Fig. 3, Table 1), while the ∆*F*/*Fm'* values of Med-plants were not negatively affected but, rather, slightly
- increased under warming in the case of the Med-Pioneer species (Fig. 3, Table 1). As a consequence, we detected a
- significant interaction in H×L×T for ∆*F*/*Fm'* measurements (ANOVA: *F* (1,16) = 14.267, *p* < 0.01). It is important
- to highlight that while the control plants exhibited a similar level of performance, heated Climax plants from the two
- hemispheres responded differently and significant differences were detected from all photo-physiological
- measurements (SNK test for 'H:L:T' among 'H' within 'L:T': *p* < 0.001; Table S1).
- 3.1.2. Climax versus pioneer seagrasses
- The simulated MHW strongly impacted the photosynthetic performances (both *Fv/Fm* and ∆*F*/*Fm'*) of Aus-Climax
- plants, however, the level of warming impacts were much lower in the Aus-Pioneer plants (Fig. 3). Climax-pioneer
- dissimilarities in response to warming were also found in the activation of NPQ machinery. While Aus-Climax
- plants significantly activated their NPQ machinery (Fig. 3) as mentioned above, on the other hand, Aus-Pioneer
- plants did not alter their NPQ even at the same warming condition (Fig. 3). This is also evidenced from the SNK
- results for 'H:L:T' among 'L' within 'H:T' when no significant difference detected for Aus-control plants but Aus-
- 318 treatment plants (SNK test: $p < 0.001$; Table S1).

3.2. Plant growth responses

 Figure 4: Boxplot graphs present plant growth response results at the end of the experiments (*n* = 3). Asterisks indicate statistical differences between control and treatment within each species (Student-Newman-Keuls *post-hoc* test, * *p* < 0.05, ** *p* < 0.01, more details can be found in Supplementary data, Table S1).

- 3.2.1. Northern versus southern hemisphere seagrasses
- There were differences in response to warming between northern versus southern hemisphere seagrasses in both
- 326 biomass production and leaf elongation measurements (Fig. 4), as shown by the significant $H \times T$ interactions
- 327 (ANOVA: $F(1,16) = 14.532$, $p < 0.01$ and $F(1,16) = 10.151$, $p < 0.01$, respectively). Among climax plants,
- warming significantly reduced biomass production (SNK test: *p* < 0.05; Fig. 4a) as well as leaf elongation (SNK
- 329 test: $p < 0.01$; Fig. 4b) of the southern plants. On the other hand, warming favored the developments of the northern
- ones in terms of productivity (Fig. 4a,b). Differently, warming increased the growth of northern pioneer plants (e.g.
- a significant difference between control versus treatment detected for biomass production, SNK test: *p* < 0.05; Fig.
- 4a). In contrast, the southern pioneer plants suffered a reduction in growth as a result of their exposure to a
- simulated MHW (Fig. 4a,b).
- 3.2.2. Climax versus pioneer seagrasses
- Even if we did not detect any significant difference between climax versus pioneer species within each hemisphere
- 336 (ANOVA: L×T, $p > 0.05$ for both plant growth response measurements), it is interesting to note that there were
- significant interactions of H×L for both biomass production and leaf elongation (ANOVA: *F* (1,16) = 13.540, *p* <
- 338 0.01 and $F(1,16) = 16.271$, $p < 0.001$, respectively). For Med-seagrasses, even if warming generally enhanced the
- developments of both Med-Climax plants and Med-Pioneer plants, the levels of increments were significantly higher
- in Med-Pioneer plants in comparison with its climax counterpart (Fig. 4a,b). Differently, Aus-Climax plants
- exhibited greater impact of warming when compared with their pioneer counterparts (Fig. 4a,b) with significant
- differences between control versus treatment detected for both plant growth response measurements only for Aus-
- 343 Climax plants (SNK test: $p < 0.05$ and $p < 0.01$; Fig. 4a, b).

3.3. Pigment content responses

- Figure 5: Boxplot graphs of pigment results at the end HW exposure (*n* = 3). Asterisks indicate statistical differences between
- control and treatment within each species (Student-Newman-Keuls *post-hoc* test, * *p* < 0.05, ** *p* < 0.01, *** *p* < 0.001, more details can be found in Supplementary data, Table S1).
- Results from pigment content measurements showed complex interactions between northern versus southern as well
- 351 as climax versus pioneer seagrasses in response to warming. Significant interactions were detected in $H \times L \times T$ for all
- pigment measuring parameters (ANOVA, *p* < 0.05, Table 1). Details are presented below.
- 3.3.1. Northern versus southern hemisphere seagrasses
- Warming significantly reduced all pigments content of Med-Climax plants such as Chl *a* (SNK test: *p* < 0.01, Fig.
- 355 5a), Chl *b* (SNK test: $p < 0.01$, Fig. 5b), Chl $a+b$ (SNK test: $p < 0.01$, Fig. 5c), and total carotenoids (SNK test: $p <$
- 0.01, Fig. 5d) but did not result in any significant reduction in pigment content for Aus-Climax plants (except for the
- case of Chl *a*, although the level of reduction was greater in Med-Climax plants; Fig. 5a,b,c,d). Interestingly, while
- Med-Climax plants maintained their Chl *b/a* molar ratio during HW, Aus-Climax plants significantly increased the
- ratio (SNK test: *p* < 0.01, Fig. 5e). It is worth mentioning that while Med-Pioneer plants accumulated more pigment
- content under the increased temperature, Aus-Pioneer plants reduced the accumulation of these pigments (see Fig.
- 5a,b,c,d). Likewise, Aus-Pioneer plants increased their Chl *b/a* molar ratio as a result of warming, while Med-
- Pioneer plants exposed to warming showed values similar to control plants (Fig. 5e).
- 3.3.2. Climax versus pioneer seagrasses
- Warming greatly impacted the Med-Climax plants in terms of pigment contents including Chl *a*, Chl *b*, Chl *a*+*b* as
- well as total carotenoids with significant differences detected between control versus heated plants across all these
- measurements (Fig. 5a,b,c,d). On the contrary, warmed Med-Pioneer plants significantly improved pigment contents
- as a result of warming (Fig. 5a,b,c,d). Furthermore, we assessed a statistical difference between heated Med-Climax
- plants versus heated Med-Pioneer plants in terms of total carotenoids' response (SNK test for 'H:L:T' among 'L'
- within 'H:T': *p* < 0.001; Table S1). For the southern hemisphere plants, warming negatively affected both climax
- and pioneer plants in terms of pigments (Fig. 5).
-
-

373

374

Table 1: Results of three-way ANOVA analyses. Significant codes: *** *p* < 0.001; ** *p* < 0.01; * *p* < 0.05; ns *p* > 0.05; φ means 0.01 < *p* < 0.05 but not interpreted as significant because of variance heterogeneity 376 because of variance heterogeneity in Cochran's C test; *Sqrt*: Square root.

4. Discussion

4.1. Difference between northern versus southern hemisphere seagrasses in response to warming

When two sister species of the genus *Posidonia* were exposed to a similar simulated MHW (i.e. 32℃), their photo-

- physiological and plant growth responses clearly demonstrated that the southern hemisphere species *P. australis* (i.e.
- Aus-Climax) is more sensitive to anomalous thermal events than the northern hemisphere species *P. oceanica* (i.e.
- Med-Climax). Warming dramatically affected the photosynthetic performance of *P. australis,* while the
- photosynthetic functioning of *P. oceanica* was unaffected. The impairment of the photosynthetic apparatus, reflected
- as a reduction in the maximum (i.e. *Fv*/*Fm*) and effective photochemical efficiency (i.e. ∆*F*/*Fm'*), is a common
- response in seagrasses subjected to thermal stress (e.g. see Winters et al. 2011; Marín-Guirao et al. 2016; Ruocco et
- al. 2019a; Nguyen et al. 2020a for some recent studies) and evidenced a higher photosynthetic thermal sensitivity in
- *P. australis* with regard to *P. oceanica*. This was further supported by the fact that only *P. australis* activated the
- NPQ machinery, a well-known photo-protective mechanism in plants (including seagrasses) that mitigates the
- damaging effects of a heat-induced photosynthetic malfunction by dissipating excess energy as heat (e.g. see Ashraf
- and Harris 2013 for a review in plants and Marín-Guirao et al. 2016; Ontoria et al. 2019 for some recent studies in
- seagrasses). Moreover, only *P. australis* experienced a significant growth inhibition during the warming exposure.
- Reduction in plant growth is a major consequence of growing under stress conditions and is commonly associated
- with photosynthetic constrains under high temperatures and with the diversion of resources from growth to sustain a
- heat-stress response and to repair heat-induced damage (e.g. Wahid et al. 2007; Bita and Gerats 2013; York et al.
- 2013; Collier et al. 2017; Marín-Guirao et al. 2018). Interestingly, while warming reduced the overall pigment
- content (i.e. Chl *a,* Chl *b* and carotenoids content) of *P. oceanica* plants, the same level of warming only reduced
- Chl *a* content in the southern hemisphere plants. This resulted in a significant Chl *b/a* molar ratio increment (i.e. a
- proxy of PSII antenna size), suggesting that *P. australis* attempted to counterbalance their heat-impaired
- photosynthetic performance by enhancing their light harvesting efficiency.
- Both *P. oceanica* and *P. australis* together with 7 other species including *P. sinuosa*, *P. angustifolia*, *P. coriacea*, *P.*
- *denhartogii*, *P. kirkmanii*, *P. ostenfeldii*, and *P. robertsoniae* belong to the genus *Posidonia* which is among the
- most primitive marine angiosperm genus (den Hartog 1970; Kuo and Cambridge 1984). Interestingly, while *P.*
- *oceanica* is endemic to the Mediterranean, the other 8 species (including *P. australis*) occur exclusively in the
- subtropical and temperate Australian seas (Kuo and Cambridge 1984). It is still unclear when the single
- Mediterranean species and the Australian congeneric counterparts diverged. Phillips and Meinez (1988) suggested it
- could have happened during the late Eocene, about 40 million years ago (Mya), while Les et al. (2003) estimated a
- 407 more recent separation of 16.7 ± 12.3 Mya. A more recent study from Aires et al. (2011) predicted this divergence
- would have taken place much earlier in the ancient Tethys Sea (i.e. over 60 Mya). In any case, the disconnection of
- Mediterranean *Posidonia* with the Australian ones has allowed the two groups to evolve in two contrasting
- environmental and evolutionary conditions (i.e. Mediterranean Sea versus Australian Seas). Compared to the
- Australian Seas, the Mediterranean has undergone massive changes during its history (Bianchi, Carlo and Morri
- 2017). Especially, due to anthropogenic climate change, the Mediterranean Sea waters have warmed up at a faster
- pace (Bianchi, Carlo and Morri 2017; Ozer et al. 2017; Nguyen et al. 2020a), become saltier (Borghini et al. 2014),
- and exhibited more frequent and intense extreme oceanic events (e.g. MHWs, see Darmaraki et al. 2019). In
- addition, not only the species but also the studied *P. oceanica* population thrives in a broader thermal regime (i.e. 13
- 28℃, Fig. 1B; see Fig. S1 for more details about species distribution and species thermal range) than *P. australis*
- 417 (i.e. $17 26\degree$ C, Fig. 1c; see Fig. S1 for more details about species distribution and species thermal regime); and this,
- together with the evolutionary differences among both *Posidonia* species stated above, may explain why the
- northern hemisphere *Posidonia* was less affected by warming than its southern hemisphere counterpart.
- Regarding the pioneer seagrass species, this study also pinpoints some dissimilarities in the response to warming
- between *C. nodosa* (i.e. Med-Pioneer) and *Z. muelleri* (i.e. Aus-Pioneer). For example, warming significantly
- impacted the photosynthetic functioning of *Z. muelleri* (i.e. reduced ∆*F*/*Fm'* values) while no significant changes
- were detected for *C. nodosa*. Likewise, warming favored the growth and biomass production of *C. nodosa* but not
- for *Z. muelleri*, and similar divergences were also found in their responses at the level photosynthetic pigments.
- These divergent responses to seawater warming manifested that the *C. nodosa* species, which is indeed benefited by
- increased temperatures, is more tolerant to anomalous heat events than the *Z. muelleri*. This finding suggests that the
- differences in response to warming among northern and southern hemisphere seagrasses may not be limited to the
- genus *Posidonia*, but extended to other seagrass species across hemispheres. However, since both pioneer species
- belong to a different family with contrasting origins and estimated ages (*Cymodoceaceae*: 67 Mya vs. *Zosteraceae*:
- 47Mya; Janssen and Bremer 2004; Waycott et al. 2007), the comparison is not as direct as in the two studied
- *Posidonia* species. Hence, further studies to compare the responses to warming of other seagrass species across
- hemisphere are warranted.

4.2. Difference between climax versus pioneer seagrasses in response to warming

- The northern hemisphere climax and pioneer species reacted almost in the same way to warming in terms of photo-
- physiology (i.e. no significant changes along with warming) and growth (i.e. greatly enhancements along with
- warming), whereas their responses differed in regard to pigment content modifications. The climax plants reduced
- all pigments (i.e. Chl *a*, Chl *b* and total carotenoids) during the warming exposure, while on the contrary, the pioneer
- plants increased the overall pigment content as a result of warming. These results indicated that, although both
- species came from the same thermal regime (both population and species), the climax seagrass was slightly
- impacted by the simulated MHW while the pioneer species even benefited from the warming exposure. The
- differences between the southern hemisphere species in response to warming was stronger for the photo-
- physiological parameters. *P. australis* plants experienced greater reductions in both *Fv*/*Fm* and ∆*F*/*Fm'* values
- compared with *Z. muelleri* plants (Fig. 3a,b). In addition, only the climax plants significantly increased their NPQ as
- a result of thermal stress. Additionally, the Australian species did not show differences between climax and pioneer
- species in regard to pigment content response. These results are in line with findings from the northern hemisphere
- experiment on demonstrating that climax seagrasses are more prone to be adversely affected by warming than
- pioneer species. The fact that we observed the 'climax-pioneer' pattern in different traits for the two hemispheres
- suggests that species also differed in their mechanisms of response to warming.
- In order to correctly interpret this result, we should take into consideration that the difference in the species thermal
- range can affect the species response to the temperature imposed during the experiment. *P. australis* was exposed to
- a warming treatment that goes above its theoretical thermal tolerance, possibly masking the effect due to its life
- strategy. Nevertheless, species temperature range was here assessed looking at species distribution, but maximum
- temperature tolerance limits can be higher. This is the case of *P. oceanica*, where the species occurs in a coastal
- enclosed lagoon (i.e. the Stagnone di Marsala, Italy) exhibiting temperature and salinity tolerant levels above its
- theoretical tolerance limit (Tomasello et al. 2009).

 Differences between the response of climax versus pioneer species to environmental stressors have been previously documented in other seagrasses. For instance, Masini and Manning (1997) showed the pioneer seagrasses (i.e. *Amphibolis griffithii* and *A. antarctica*) were more resilient to changes in light and temperature when compared to two other climax seagrasses (i.e. *P. sinuosa* and *P. australis*) inhabiting in the same region of Western Australia. Similarly, the Mediterranean pioneer *C. nodosa* was also shown to be more thermal tolerant than the Mediterranean climax *P. oceanica* (Marín-Guirao et al. 2018), but also to other abiotic stress factors including light (Olesen et al. 2002) and salinity (Sandoval-Gil et al. 2014), which seems to be related to their different levels of phenotypic plasticity (Pazzaglia et al. 2021). Seddon and Cheshire (2001) also suggested that the climax *P. australis* is more vulnerable to desiccation in high-temperature conditions than the pioneer *A. antarctica*. All these evidences imply that warming can reshape the seagrass landscape by reducing the presence of climax species while enhancing the distribution of pioneer seagrasses. For instance, in Mission Bay, San Diego Bay, and Chesapeake Bay (USA), the climax seagrass *Z. marina* was replaced by the pioneer *Ruppia maritima* following extreme climatic events (Johnson et al. 2003; Shields et al. 2019). The same phenomenon is predicted to occur also in the Mediterranean, where ocean warming is expected to cause a decline of *P. oceanica* (Marbà and Duarte 2010; Chefaoui et al. 2018) while favoring the expansion of some pioneer species (e.g. *C. nodosa*, *Halophila stipulacea*) (Savva et al. 2018; Winters et al. 2020). Changes in seagrass meadow composition at the landscape scale would ultimately reduce their ecological value (Orth et al. 2006; Lamb et al. 2017; Unsworth et al. 2019) and, hence, affect the livelihoods of billions of people living in coastal areas (Bertelli and Unsworth 2014). The replacement of climax seagrass species, generally

- characterized by high biomass and productivity, by pioneer species will also decrease the capacity of seagrass
- meadows to mitigate the effects of carbon emissions (Gattuso et al. 2018). Under some warming scenarios,
- seagrasses ecosystem may even switch metabolism from autotrophic to heterotrophic (Burkholz et al. 2019), and
- 477 enhance CO₂ and methane fluxes from the meadows into the atmosphere (Burkholz et al. 2020).

4.3. Future perspectives

Our study brings evidence of differences in the response to warming between climax and pioneer seagrasses and

suggests that these differences are consistent between the northern and southern hemispheres. Ocean warming is

- happening fast and accelerating the tropicalization of temperate seagrass meadows (see review by Hyndes et al.
- 2016). Our results emphasize the need to protect climax seagrasses because they are more ecologically valuable but
- also more susceptible to warming than pioneer species. Optimistically, we can enhance the resilience of climax
- seagrasses to warming through genotype selection, synthetic biology, assisted evolution, and the use of microbiome
- (Bulleri et al. 2018). Such approaches require, however, a deeper understanding of the relationship between
- seagrasses and the surrounding environments.

Compliance with ethical standards

Conflicts of interest: The authors declare that there is no conflict of interest.

Ethical approval: This article does not contain any studies with animals performed by any of the authors.

Authors' contributions: Conceptualization: Hung Manh Nguyen, Fabio Bulleri, Lázaro Marín-Guirao, Mathieu

Pernice, Gabriele Procaccini; **Methodology**: Hung Manh Nguyen, Lázaro Marín-Guirao, Mathieu Pernice, Gabriele

Procaccini; **Formal analysis and investigation**: Hung Manh Nguyen; **Writing - original draft preparation**: Hung

Manh Nguyen; **Writing - review and editing**: Hung Manh Nguyen, Fabio Bulleri, Lázaro Marín-Guirao, Mathieu

Pernice, Gabriele Procaccini; **Funding acquisition**: Fabio Bulleri, Lázaro Marín-Guirao, Mathieu Pernice, Gabriele

- Procaccini; **Resources**: Mathieu Pernice, Gabriele Procaccini; **Supervision**: Lázaro Marín-Guirao, Mathieu Pernice,
- Gabriele Procaccini.

Acknowledgment

- HN was supported by an SZN Ph.D. fellowship via the Open University. This work was funded by EPIC-SEA
- Project, Extra-Eu Scientific Research and Cooperation Grant of the SZN, by the project Marine Hazard,
- PON03PE_00203_1, Italian Ministry of Education, University and Research (MIUR) and by the project Assemble
- Plus EU-FP7. The authors are grateful for precious support from Peter J. Ralph, Mikael Kim, Nasim Shah
- Mohammadi, Paul Brooks, Scott Allchin, Susan Fenech, and Kun Xiao (UTS) for the Aus-mesocosm experiment at
- UTS, Australia as well as Emanuela Dattolo, Alex Santillán-Sarmiento, Ludovica Pedicini and Jessica Pazzaglia
- (SZN) for the Med-mesocosm experiment at SZN, Italy. Moreover, the authors thank three anonymous reviewers for
- many constructive criticisms and suggestions.

References

- Aires T, Marbà N, Cunha RL, Kendrick GA, Walker DI, Serrão EA, Duarte CM, Arnaud-Haond S (2011) Evolutionary history of the seagrass genus *Posidonia*. Mar Ecol Prog Ser 421:117–130.
- Arias-Ortiz A, Serrano O, Masqué P, Lavery PS, Mueller U, Kendrick GA, Rozaimi M, Esteban A, Fourqurean JW,
- Marbà N, Mateo MA, Murray K, Rule MJ, Duarte CM (2018) A marine heatwave drives massive losses from
- the world's largest seagrass carbon stocks. Nat Clim Chang 8:338–344. doi: 10.1038/s41558-018-0096-y
- Arnaud-Haond S, Duarte CM, Diaz-Almela E, Marbà N, Sintes T, Serrão EA (2012) Implications of extreme life
- span in clonal organisms: Millenary clones in meadows of the threatened seagrass *Posidonia oceanica*. PLoS
- One 7:e30454. doi: 10.1371/journal.pone.0030454
- Ashraf M, Harris PJC (2013) Photosynthesis under stressful environments: an overview. Photosynthetica 51:163–

190.

- Bergmann N, Winters G, Rauch G, Eizaguirre C, Gu J, Nelle P, Fricke B, Reusch TBH (2010) Population- specificity of heat stress gene induction in northern and southern eelgrass *Zostera marina* populations under simulated global warming. Mol Ecol 19:2870–2883. doi: 10.1111/j.1365-294X.2010.04731.x
- Bertelli CM, Unsworth RKF (2014) Protecting the hand that feeds us: Seagrass (*Zostera marina*) serves as
- commercial juvenile fish habitat. Mar Pollut Bull 83:425–429. doi: 10.1016/j.marpolbul.2013.08.011
- Bianchi, Carlo N, Morri C (2017) Global sea warming and "tropicalization" of the Mediterranean Sea: biogeographic and ecological aspects. Biogeogr – J Integr Biogeogr 24:319–327. doi: 10.21426/b6110129
- Bita CE, Gerats T (2013) Plant tolerance to high temperature in a changing environment: scientific fundamentals and production of heat stress-tolerant crops. Front Plant Sci 4:273.
- Borghini M, Bryden H, Schroeder K, Sparnocchia S, Vetrano A (2014) The Mediterranean is getting saltier. Ocean Sci 10: 693–700.
- Brandão SE, Bulbovas P, Lima MEL, Domingos M (2017) Biochemical leaf traits as indicators of tolerance potential in tree species from the Brazilian Atlantic Forest against oxidative environmental stressors. Sci Total Environ 575:406–417.
- Bulleri F, Eriksson BK, Queirós A, Airoldi L, Arenas F, Arvanitidis C, Bouma TJ, Crowe TP, Davoult D, Guizien
- K, Iveša L, Jenkins SR, Michalet R, Olabarria C, Procaccini G, Serrão EA, Wahl M, Benedetti-Cecchi L
- (2018) Harnessing positive species interactions as a tool against climate-driven loss of coastal biodiversity.

PLoS Biol 16:e2006852. doi: 10.1371/journal.pbio.2006852

- Burkholz C, Duarte CM, Garcias-Bonet N (2019) Thermal dependence of seagrass ecosystem metabolism in the Red Sea. Mar Ecol Prog Ser 614:79–90. doi: 10.3354/meps12912
- Burkholz C, Garcias-Bonet N, Duarte CM (2020) Warming enhances carbon dioxide and methane fluxes from Red Sea seagrass sediments. Biogeosciences 17:1717–1730. doi: 10.5194/bg-2019-93
- Campbell SJ, McKenzie LJ, Kerville SP (2006) Photosynthetic responses of seven tropical seagrasses to elevated seawater temperature. J Exp Mar Bio Ecol 330:455–468.
- Chefaoui RM, Duarte CM, Serrão EA (2018) Dramatic loss of seagrass habitat under projected climate change in the Mediterranean Sea. Glob Chang Biol 24:4919–4928.
- Collier CJ, Waycott M (2014) Temperature extremes reduce seagrass growth and induce mortality. Mar Pollut Bull 83:483–490. doi: 10.1016/j.marpolbul.2014.03.050
- Collier CJ, Ow YX, Langlois L, Uthicke S, Johansson CL, O'Brien KR, Hrebien V, Adams MP (2017) Optimum temperatures for net primary productivity of three tropical seagrass species. Front Plant Sci 8:1446.
- Coma R, Ribes M, Serrano E, Jimé Nez E, Salat J, Pascual J (2009) Global warming-enhanced stratification and mass mortality events in the Mediterranean. Proc Natl Acad Sci U S A 106:6176–6181.
- Costanza R, de Groot R, Sutton P, van der Ploeg S, Anderson SJ, Kubiszewski I, Farber S, Turner RK (2014) Changes in the global value of ecosystem services. Glob Environ Chang 26:152–158. doi:
- http://dx.doi.org/10.1016/j.gloenvcha.2014.04.002
- Cowling RM, Rundel PW, Lamont BB, Arroyo MK, Arianoutsou M (1996) Plant diversity in Mediterranean-climate regions. Trends Ecol Evol 11:362–366.
- Darmaraki S, Somot S, Sevault F, Nabat P, Cabos Narvaez WD, Cavicchia L, Djurdjevic V, Li L, Sannino G, Sein D V. (2019) Future evolution of Marine Heatwaves in the Mediterranean Sea. Clim Dyn 53:1371–1392. doi: 10.1007/s00382-019-04661-z
- den Hartog C (1970) The sea-grasses of the World. North-Holland Publication Co., Amsterdam
- Esposito MP, Nakazato RK, Pedroso ANV, Lima MEL, Figueiredo MA, Diniz AP, Kozovits AR, Domingos M
- (2018) Oxidant-antioxidant balance and tolerance against oxidative stress in pioneer and non-pioneer tree species from the remaining Atlantic Forest. Sci Total Environ 625:382–393.
- Faria T, Silvério D, Breia E, Cabral R, Abadia A, Abadia J, Pereira JS, Chaves MM (1998) Differences in the response of carbon assimilation to summer stress (water deficits, high light and temperature) in four Mediterranean tree species. Physiol Plant 102:419–428.
- Favaretto VF, Martinez CA, Soriani HH, Furriel RPM (2011) Differential responses of antioxidant enzymes in pioneer and late-successional tropical tree species grown under sun and shade conditions. Environ Exp Bot 70:20–28.
- Fourqurean JW, Duarte CM, Kennedy H, Marbà N, Holmer M, Mateo MA, Apostolaki ET, Kendrick GA, Krause- Jensen D, McGlathery KJ (2012) Seagrass ecosystems as a globally significant carbon stock. Nat Geosci 5:505–509.
- Gattuso J-P, Magnan AK, Bopp L, Cheung WWL, Duarte CM, Hinkel J, Mcleod E, Micheli F, Oschlies A, Williamson P, Billé R, Chalastani VI, Gates RD, Irisson J-O, Middelburg JJ, Pörtner H-O, Rau GH (2018) Ocean Solutions to Address Climate Change and Its Effects on Marine Ecosystems. Front Mar Sci 5:337. doi: 10.3389/fmars.2018.00337
- Glenn-Lewin DC, Peet RK, Veblen TT (1992) Plant succession: theory and prediction. Springer Science & Business Media, Berlin
- Guidetti P, Cristina Buia M, Zupo V, Beatrice M, Guidetti P, Lorenti M, Buia M, Zupo V, Scipione M, Rismondo A, Curiel D (1998) Biomass partitioning in Adriatic seagrass ecosystems (*Posidonia oceanica*, *Cymodocea nodosa*, *Zostera marina*). Rapp Comm int Mer Médit 35:562–563.
- Harley CDG, Anderson KM, Demes KW, Jorve JP, Kordas RL, Coyle TA, Graham MH (2012) Effects Of Climate Change On Global Seaweed Communities. J Phycol 48:1064–1078. doi: 10.1111/j.1529-8817.2012.01224.x
- Hobday AJ, Alexander L V, Perkins SE, Smale DA, Straub SC, Oliver ECJ, Benthuysen JA, Burrows MT, Donat MG, Feng M (2016) A hierarchical approach to defining marine heatwaves. Prog Oceanogr 141:227–238.
- Hyndes GA, Heck KL, Vergés A, Harvey ES, Kendrick GA, Lavery PS, McMahon K, Orth RJ, Pearce A,
- Vanderklift M, Wernberg T, Whiting S, Wilson S (2016) Accelerating tropicalization and the transformation of temperate seagrass meadows. Bioscience 66:938–945. doi: 10.1093/biosci/biw111
- 587 Janssen T, Bremer K (2004) The age of major monocot groups inferred from 800+ rbcL sequences. Bot J Linn Soc 146:385–398.
- Jarvis JC, Brush MJ, Moore KA (2014) Modeling loss and recovery of *Zostera marina* beds in the Chesapeake Bay: The role of seedlings and seed-bank viability. Aquat Bot 113:32–45.

 Johnson MR, Williams SL, Lieberman CH, Solbak A (2003) Changes in the abundance of the seagrasses *Zostera marina* L. (eelgrass) and *Ruppia maritima* L.(widgeongrass) in San Diego, California, following and El Niño Event. Estuaries 26:106.

- Kuo J, Cambridge ML (1984) A taxonomic study of the *Posidonia ostenfeldii* complex (*Posidoniaceae*) with description of four new Australian seagrasses. Aquat Bot 20:267–295.
- Lamb JB, van de Water JAJMJM, Bourne DG, Altier C, Hein MY, Fiorenza EA, Abu N, Jompa J, Harvell CD,
- Crispr H, States U (2017) Seagrass ecosystems reduce exposure to bacterial pathogens of humans, fishes, and invertebrates. Science 355:731–733. doi: 10.1126/science.aal1956
- Larkum AWD, Kendrick GA, Ralph PJ (2018) Seagrasses of Australia: Structure, Ecology and Conservation. Springer, New York
- Les DH, Crawford DJ, Kimball RT, Moody ML, Landolt E (2003) Biogeography of discontinuously distributed hydrophytes: a molecular appraisal of intercontinental disjunctions. Int J Plant Sci 164:917–932.
- Macreadie PI, Hardy SSS (2018) Response of seagrass "Blue Carbon" stocks to increased water temperatures. Diversity 10:115. doi: 10.3390/d10040115
- Marbà N, Duarte CM (2010) Mediterranean warming triggers seagrass (*Posidonia oceanica*) shoot mortality. Glob Chang Biol 16:2366–2375. doi: 10.1111/j.1365-2486.2009.02130.x
- Marín-Guirao L, Ruiz JM, Sandoval-Gil JM, Bernardeau-Esteller J, Stinco CM, Meléndez-Martínez A (2013)
- Xanthophyll cycle-related photoprotective mechanism in the Mediterranean seagrasses *Posidonia oceanica* and *Cymodocea nodosa* under normal and stressful hypersaline conditions. Aquat Bot 109:14–24.
-
- Marín-Guirao L, Ruiz JM, Dattolo E, Garcia-Munoz R, Procaccini G (2016) Physiological and molecular evidence
- of differential short-term heat tolerance in Mediterranean seagrasses. Sci Rep 6: 28615.
- Marín-Guirao L, Bernardeau-Esteller J, García-Muñoz R, Ramos A, Ontoria Y, Romero J, Pérez M, Ruiz JM, Procaccini G (2018) Carbon economy of Mediterranean seagrasses in response to thermal stress. Mar Pollut Bull 135:617–629.
- 615 Marín-Guirao L, Entrambasaguas L, Ruiz JM, Procaccini G (2019) Heat-stress induced flowering can be a potential adaptive response to ocean warming for the iconic seagrass *Posidonia oceanica*. Mol Ecol 28:2486–2501. doi:
- 10.1111/mec.15089
- Masini RJ, Manning CR (1997) The photosynthetic responses to irradiance and temperature of four meadow-forming seagrasses. Aquat Bot 58:21–36.
- Maxwell K, Johnson GN (2000) Chlorophyll fluorescence A practical guide. J Exp Bot 51:659–668. doi: 10.1093/jxb/51.345.659
- Nelson WG (2017) Patterns of shading tolerance determined from experimental light reduction studies of seagrasses. Aquat Bot 141:39–46.
- Nguyen HM, Ralph PJ, Marín-Guirao L, Pernice M, Procaccini G (2021) Seagrasses in an era of ocean warming: a review. Biol Rev *in press* https://doi.org/10.1111/brv.12736
- Nguyen HM, Yadav NS, Barak S, Lima FP, Sapir Y, Winters G (2020a) Responses of invasive and native
- populations of the seagrass *Halophila stipulacea* to simulated climate change. Front Mar Sci 6:812. doi: 10.3389/FMARS.2019.00812
- Nguyen HM, Kim M, Ralph PJ, Marín-Guirao L, Pernice M, Procaccini G (2020b) Stress memory in seagrasses: first insight into the effects of thermal priming and the role of epigenetic modifications. Front Plant Sci 11:494. doi: 10.3389/FPLS.2020.00494
- Olesen B, Enríquez S, Duarte CM, Sand-Jensen K (2002) Depth-acclimation of photosynthesis, morphology and demography of *Posidonia oceanica* and *Cymodocea nodosa* in the Spanish Mediterranean Sea. Mar Ecol Prog Ser 236:89–97.
- Ontoria Y, Cuesta-Gracia A, Ruiz JM, Romero J, Pérez M (2019) The negative effects of short-term extreme thermal events on the seagrass *Posidonia oceanica* are exacerbated by ammonium additions. PLoS One 14:e0222798. doi: 10.1371/journal.pone.0222798
- Orth RJ, Carruthers TJB, Dennison WC, Duarte CM, Fourqurean JW, Heck KL, Hughes AR, Kendrick GA, Kenworthy WJ, Olyarnik S (2006) A global crisis for seagrass ecosystems. Bioscience 56:987–996.
- Ozer T, Gertman I, Kress N, Silverman J, Herut B (2017) Interannual thermohaline (1979–2014) and nutrient
- (2002–2014) dynamics in the Levantine surface and intermediate water masses, SE Mediterranean Sea. Glob
- Planet Change 151:60–67. doi: 10.1016/j.gloplacha.2016.04.001
- Pazzaglia J, Reusch TBH, Terlizzi A, Marin Guirao L, Procaccini G (2021) Prompt phenotypic plasticity under rapid global changes: the intrinsic force for future seagrasses survival. Evol Appl 14:1181–1201. doi: https://doi.org/10.1111/eva.13212
- Phillips RC, Menez EG (1988) Smithsonian contributions to the marine sciences. Smithsonian Institution Press, Washington, D.C
- Procaccini G, Buia MC, Gambi MC, Perez M, Pergent G, Pergent-Martini C, Romero J (2003) Seagrass status and
- extent along the Mediterranean coasts of Italy, France and Spain. In: EP G, FT S, Spalding M (eds) World Atlas of Seagrass: Present status and future conservation. University of California Press, Berkeley
- R Core Team (2018) R: A Language and Environment for Statistical Computing, R Foundation for Statistical Computing, Austria, 2015.
- Ralph PJ, Durako MJ, Enriquez S, Collier CJ, Doblin MA (2007) Impact of light limitation on seagrasses. J Exp Mar Bio Ecol 350:176–193.
- Ruocco M, De Luca P, Marín-Guirao L, Procaccini G (2019a) Differential Leaf Age- Dependent Thermal Plasticity in the Keystone Seagrass *Posidonia oceanica*. Front Plant Sci 10:1556. doi: 10.3389/fpls.2019.01556
- Ruocco M, Marín-Guirao L, Procaccini G (2019b) Within- and among-leaf variations in photo-physiological functions, gene expression and DNA methylation patterns in the large-sized seagrass *Posidonia oceanica*. Mar Biol 166:24. doi: 10.1007/s00227-019-3482-8
- Salinas C, Duarte CM, Lavery PS, Masque P, Arias‐Ortiz A, Leon JX, Callaghan D, Kendrick GA, Serrano O
- (2020) Seagrass losses since mid‐20th century fuelled CO2 emissions from soil carbon stocks. Glob Chang Biol 26:4772–4784.
- Sandoval-Gil JM, Ruiz JM, Marín-Guirao L, Bernardeau-Esteller J, Sánchez-Lizaso JL (2014) Ecophysiological plasticity of shallow and deep populations of the Mediterranean seagrasses *Posidonia oceanica* and *Cymodocea nodosa* in response to hypersaline stress. Mar Environ Res 95:39–61.
- Sandrini-Neto L, Camargo M (2014) GAD: an R package for ANOVA designs from general principles. software
- Savva I, Bennett S, Roca G, Jordà G, Marbà N (2018) Thermal tolerance of Mediterranean marine macrophytes: Vulnerability to global warming. Ecol Evol 8:12032–12043.
- Seddon S, Cheshire AC (2001) Photosynthetic response of *Amphibolis antarctica* and *Posidonia australis* to temperature and desiccation using chlorophyll fluorescence. Mar Ecol Prog Ser 220:119–130.
- Shields EC, Parrish D, Moore K (2019) Short-Term Temperature Stress Results in Seagrass Community Shift in a Temperate Estuary. Estuaries and Coasts 42:755–764. doi: 10.1007/s12237-019-00517-1
- Short FT, Carruthers T, Dennison W, Waycott M (2007) Global seagrass distribution and diversity: a bioregional
- model. J Exp Mar Bio Ecol 350:3–20.
- Smale DA (2020) Impacts of ocean warming on kelp forest ecosystems. New Phytol 225:1447–1454.

Smale DA, Wernberg T, Oliver ECJ, Thomsen M, Harvey BP, Straub SC, Burrows MT, Alexander L V, Benthuysen

- JA, Donat MG (2019) Marine heatwaves threaten global biodiversity and the provision of ecosystem services.
- Nat Clim Chang 9:306–312.
- Strydom S, Murray K, Wilson S, Huntley B, Rule M, Heithaus M, Bessey C, Kendrick GA, Burkholder D, Fraser
- MW, Zdunic K (2020) Too hot to handle: Unprecedented seagrass death driven by marine heatwave in a World Heritage Area. Glob Chang Biol 26:3525–3538. doi: 10.1111/gcb.15065
- Telesca L, Belluscio A, Criscoli A, Ardizzone G, Apostolaki ET, Fraschetti S, Gristina M, Knittweis L, Martin CS, Pergent G (2015) Seagrass meadows (*Posidonia oceanica*) distribution and trajectories of change. Sci Rep 5:12505.
- Tomasello A, Di Maida G, Calvo S, Pirrotta M, Borra M, Procaccini G (2009) Seagrass meadows at the extreme of environmental tolerance: the case of *Posidonia oceanica* in a semi‐enclosed coastal lagoon. Mar Ecol 30: 288– 300.
- Touchette BW (2007) Seagrass-salinity interactions: Physiological mechanisms used by submersed marine angiosperms for a life at sea. J Exp Mar Bio Ecol 350:194–215. doi: 10.1016/j.jembe.2007.05.037
- Trautman DA, Borowitzka MA (1999) Distribution of the epiphytic organisms on *Posidonia australis* and *P. sinuosa*, two seagrasses with differing morphology. Mar Ecol Prog Ser 179:215–229. doi: 10.3354/meps179215
- Tutar O, Marín-Guirao L, Ruiz JM, Procaccini G (2017) Antioxidant response to heat stress in seagrasses. A gene expression study. Mar Environ Res 132:94–102. doi: 10.1016/j.marenvres.2017.10.011
- Underwood AJ, Underwood AL, Underwood AJ, Wnderwood AJ (1997) Experiments in ecology: their logical design and interpretation using analysis of variance. Cambridge university press, Cambridge
- Unsworth RKF, Nordlund LM, Cullen-Unsworth LC (2019) Seagrass meadows support global fisheries production. Conserv Lett 12:e12566. doi: 10.1111/conl.12566
- Wahid A, Gelani S, Ashraf M, Foolad MR (2007) Heat tolerance in plants: an overview. Environ Exp Bot 61:199– 223.
- Warren CR (2008) Rapid measurement of chlorophylls with a microplate reader. J Plant Nutr 31:1321–1332
- Waycott M, McMahon K, Mellors J, Calladine A, Kleine D (2004) A guide to tropical seagrasses of the Indo-West Pacific. James Cook University, Douglas
- Waycott M, Procaccini G, Les DH, Reusch TBH (2007) Seagrass evolution, ecology and conservation: a genetic
- perspective. In: Seagrasses: biology, ecologyand conservation. Springer, Heidelberg pp 25–50
- Waycott M, Duarte CM, Carruthers TJB, Orth RJ, Dennison WC, Olyarnik S, Calladine A, Fourqurean JW, Heck KL, Hughes AR (2009) Accelerating loss of seagrasses across the globe threatens coastal ecosystems. Proc Natl Acad Sci 106:12377–12381.
- Wellburn AR (1994) The spectral determination of chlorophylls a and b, as well as total carotenoids, using various solvents with spectrophotometers of different resolution. J Plant Physiol 144:307–313.
- Wernberg T, Bennett S, Babcock RC, De Bettignies T, Cure K, Depczynski M, Dufois F, Fromont J, Fulton CJ,
- Hovey RK, Harvey ES, Holmes TH, Kendrick GA, Radford B, Santana-Garcon J, Saunders BJ, Smale DA,
- Thomsen MS, Tuckett CA, Tuya F, Vanderklift MA, Wilson S (2016) Climate-driven regime shift of a
- temperate marine ecosystem. Science (80) 353:169–172. doi: 10.1126/science.aad8745
- Wickham H (2009) Elegant graphics for data analysis (ggplot2).
- Winters G, Nelle P, Fricke B, Rauch G, Reusch TBH (2011) Effects of a simulated heat wave on photophysiology and gene expression of high-and low-latitude populations of *Zostera marina*. Mar Ecol Prog Ser 435:83–95.
- Winters G, Beer S, Willette DA, Viana I, Chiquillo KL, Beca-Carretero P, Betty Villamayor B, Azcárate-García T, Shem-Tov R, Mwabvu B (2020) The tropical seagrass *Halophila stipulacea*: reviewing what we know from its native and invasive habitats, alongside identifying knowledge gaps. Front Mar Sci 7:300.
- York PH, Gruber RK, Hill R, Ralph PJ, Booth DJ, Macreadie PI (2013) Physiological and morphological responses of the temperate seagrass *Zostera muelleri* to multiple stressors: investigating the interactive effects of light and temperature. PLoS One 8:e76377.
- Zieman JC (1974) Methods for the study of the growth and production of turtle grass, *Thalassia testudinum* König. Aquaculture 4:139–143.