1	Influence of Phosphate and Iron Ions in Selective Uptake of
2	Arsenic Species by Water Fern (Salvinia natans L.)
3	
4	
5	
6	M. Azizur Rahman ^{*, 1} ; H. Hasegawa ^{*, 1} ; K. Ueda ¹ ; T. Maki ¹ ; M. Mahfuzur Rahman ²
7	
8	
9	
10	
11	¹ Graduate School of Natural Science & Technology, Kanazawa University, Kakuma, Kanazawa
12	920-1192, Japan; ² Department of Botany, Faculty of Biological Sciences, Jahangirnagar
13	University, Savar, Dhaka-1342, Bangladesh.
14	
15	
16	
17	
18	
19	
20	<u>*Corresponding author</u>
21	E-mail: hhiroshi@t.kanazawa-u.ac.jp (H. Hasegawa)
22	arahman@stu.kanazawa-u.ac.jp (M. A. Rahman)
23	Tel/Fax: 81-76-234-4792
24	
25	
26	

27 Abstract:

2

28 In the present study, the effect of phosphate ion and iron hydroxides (Fe-plaques) on the 29 selective uptake of arsenic species by water fern (Salvinia natans L.) was investigated. The 30 plants were grown for 5 days in aqueous Murashige and Skoog (MS) culture media modified in 31 arsenic and phosphate concentrations. Arsenic accumulations in Salvinia natans L. increased 32 with the increase of arsenate and DMAA concentrations in the culture solutions. Compared to 33 the control treatment, Salvinia natans L. accumulated significantly higher amount of arsenic 34 from phosphate deficient solutions, when the source was arsenate. However, arsenic uptake was 35 not affected significantly by phosphate, when the source was dimethylarsinic acid (DMAA). From solutions modified in 100 µM of phosphate and 4.0 µM of either arsenate or DMAA, the 36 Salvinia natans L. accumulated 0.14±0.02 and 0.02±0.00 µmol (g dry weight)⁻¹ of arsenic, 37 respectively. In contrast, plants accumulated 0.24±0.06 and 0.03±0.00 µmol (g dry weight)⁻¹ of 38 39 arsenic from solutions containing 4.0 µM of either arsenate or DMAA in the absence of 40 phosphate, respectively. Thus, it is reasonable to state that increasing phosphate concentration in 41 culture solutions decreases the arsenic uptake into the water fern significantly, when the source 42 was arsenate. Moreover, arsenic and phosphate content in plant tissue correlated significantly (r 43 = -0.66; p < 0.05), when initial source was arsenate while there were no correlation between arsenic and phosphate, when initial source was DMAA (r = -0.077; p > 0.05). Similarly, 44 significant correlation was observed between arsenic and iron content in plant tissues (r = 0.66; 45 p < 0.05), when initial source was arsenate while the correlation was not significant (r = 0.23; p 46 47 < 0.05), when initial source was DMAA. The results indicate the adsorption of arsenate on Feplaques of aquatic plant surfaces. Further, the study demonstrates that the DMAA uptake 48 49 mechanisms into the water fern are deferent from those of arsenate.

- 50
- 51
- 52

53	Keywords:	Arsenate;	DMAA;	Uptake;	Physico-chemical	Adsorption;	Water	Fern	(Salvinia
54		natans	L.); Phos	phate; Ph	ytofiltration.				

- 55
- 56
- .
- 57

58 Introduction:

3

59 Arsenic is one of the toxic environmental pollutants which have recently attracted attention 60 because of its chronic and epidemic effects to the human health through widespread water and crop contamination. Natural release of arsenic from aquifer rocks has been reported in 61 62 Bangladesh [1-4], West Bengal, India [5, 6]. Geogenic contamination of arsenic in aquifer rocks has also been reported in Thailand [7], Vietnam, inner Mongolia, Greece, Hungary, USA, 63 64 Ghana, Chile, Argentina and Mexico [8, 9]. Beside the large-scale arsenic pollution in soils, water pollution by geogenic arsenic has been a great health problem in many countries [2, 4, 6]. 65 Phytoremediation, a plant based green technology, becomes promising to remediate the 66 67 environmental pollution due to some unavoidable limitations of traditional technologies. It is 68 relatively inexpensive, eco-friendly and proven effective in few cases [10]. Although the arsenic 69 uptake into the plants occurs primarily through the root system, it is not readily translocated to 70 the shoots and the edible parts of all plants. Few terrestrial plant species, such as Agrostis 71 castellana, Agrostis delicatula [11], Bidens cynapiifolia [12], Chinese brake fern (Pteris vittata 72 L.) [13] and silver fern (*Pityrogramma calomelanos* L.) [14] accumulate high concentration of 73 arsenic in their shoots and edible parts even though the background concentration in soil is low 74 [13]. In particular, Chinese brake fern removes a significant amount of arsenic from soil [14, 75 15], and stores in the fronds [14, 16]. Arsenic accumulation in aquatic plants, such as Spirodella 76 polyrhiza L. [5], Lemna gibba L. [17, 18], Hydrilla verticullata [19], Lepidium sativum [20] has 77 also been reported in literatures.

Arsenate; As (V) and arsenite; As (III) are the inorganic forms in the oxic aquatic systems. Arsenate predominates and arsenite is oxidized to arsenate in the oxic aquatic systems [21]. The use of aquatic macrophytes or other floating plants in phytoremediation technology is commonly known as phytoextraction. This clean up process involves biosorption and accumulation of pollutants. Recently, aquatic macrophytes and some other small floating plants have been investigated for the remediation of wastewater contaminated with Cu, Cd(II) and Hg(II) [22, 23, 84 24]. The encouraging results of metal uptake capacity by aquatic plants [22-28] gained the
85 attention of researchers and scientists to use them in phytoremediation technology.

Water fern (Salvinia natans L.) is a free floating freshwater macrophyte, which grows rapidly in 86 87 ponds, lakes, ditches, and wastewater bodies mostly in southern Asian countries affected by arsenic especially in Bangladesh, West Bengal, India. Previously, the Salvinia natans L. was 88 89 tested for Hg (II) [24] and Cu (II) [28] removal. In the present study, the authors investigated the 90 effect of phosphate concentrations on arsenate and DMAA uptake and biosorption by Salvinia 91 natans L. from aqueous culture solution. The arsenate was selected because it is the predominant 92 inorganic species in oxic aquatic systems [21]. An organic species (DMAA) was also selected to 93 compare the response of the plant to both organic (DMAA) and inorganic (arsenate) species 94 uptake and biosorption in the plant.

95

96 Materials and Methods:

97 Plant Cultivation

The Salvinia natans L. were collected from rice field of Manikgonj of Dhaka, Bangladesh and 98 99 stock-cultured in a green house for two weeks. The experiment was conducted in an incubator for a 5 days period with the conditions being set as 14/10 h light/dark schedule, $100-125 \ \mu E \ m^{-2}$ 100 s⁻¹ light intensity, 75% humidity, 22 and 20 (\pm 2) °C temperatures for day and night, respectively. 101 102 Plants in the incubator were grown on modified murashige and skoog (MS) culture media where 103 modifications were in phosphorus and arsenic concentrations (Table 1). The modified culture solutions had either 50 or 100 μ M of PO₄³⁻. Either arsenate or DMAA were added to the 104 modified solutions at the rate of 1.0, 2.0 and 4.0 µM prepared from Na₂HAsO₄·7H₂O and 105 $(CH_3)_2AsO_2Na\cdot 3H_2O$, respectively. The control solution contains neither arsenic nor PO_4^{3-} . 106

107

108 Inoculation Procedure

Before inoculation, *Salvinia natans* L. strains from stock-culture were washed three times with DI water. 200-ml polystylene test vessels (118 X 86 X 60 mm) were used for the experiments. About 10 individual plants were inoculated in each of 200-ml test vessels containing 100 ml of test solution. The pH during the experiments was maintained at 5.5 through adjustment with the addition of either 0.1 M HCl or 0.1 M NaOH. Changes in volume of culture solutions during the experiment from evaporation and accumulation were compensated by adding DI water equivalent to the volume difference in every 2 days throughout the experiment.

116

117 Sample Preparation and Chemical Analysis

118 The plants (in whole) were harvested after 5 days of inoculation. After rinsing with DI water for 119 four times, plants were taken on clean absorbent paper to remove water from plant surfaces. The 120 samples were then placed into a drying oven at 65 °C until they reached a constant weight. Dried 121 samples were weighed and 0.10-0.20-g samples were digested in 50-ml polyethylene tubes 122 (DigiTubes, SCP Science, Canada). Five ml of 65% HNO3 were added and the samples were 123 kept under a fume hood for 12 hours. Then the samples were heated to 95 °C for 2 hours on a 124 heating block (DigiPREP, SCP Science, Canada). After cooling to room temperature, 3 ml of 30% hydrogen peroxide were added to the digests and the samples were heated again to 105 °C 125 126 for 20 min and then diluted to 10 ml using DI water and stored in 15-ml polythene bottles (HDPE, NALGENE[®], Nalge Nunc International, Rochester, NY). 127

The concentrations of arsenic and iron were analyzed using a graphite-furnace atomic absorption spectrometer (GF-AAS, Z-8100, Hitachi, Japan). For the determination of arsenic, 5 μ L of 0.05 M nickel nitrate was added to a 10- μ L sample as matrix modifier in the cuvette. The accuracy of the analysis was checked by the analysis of certified standard reference material 1573a tomato leaf (NIST, USA). The arsenic concentration in certified reference material was 0.112±0.004 μ g g⁻¹ while the measured arsenic concentration was 0.123±0.009 μ g g⁻¹. The concentrations 134 detected in all samples were above the instrumental limits of detection ($\geq 0.01 \ \mu M$ in samples in 135 water). Total phosphate was determined spectrophotometrically [29].

136 Chemical reagents used in this experiment were of analytical grade. All glass wares used were

- 137 washed with detergent solution, 3 M HCl and finally with DI water for eight times before use. In
- 138 each analytical batch at least two reagent blanks and three replicate samples were included.
- 139

140 Data Analysis

The experimental data were statistically analyzed for mean separation of different arsenic treatments according to the least significant difference (LSD) at 5% level by IRRI-STAT 4.0 for windows (developed by the Biometrics unit, IRRI, Philippines) and the Pearson correlation coefficient (r) was calculated by SPSS[®] statistical package (version 10.0 for windows).

145

146 **Results and Discussions:**

147 Uptake of Arsenic Species by Salvinia natans L. From Culture Solution

148 The arsenic uptake by water fern (Salvinia natans L.) at different phosphate concentrations are 149 shown in Fig. 1. After 5 days of incubation, the water fern accumulated a maximum of $0.24\pm0.02 \ \mu\text{mol} \ (\text{g dry weight})^{-1}$ of arsenic from phosphate deficient solution (P = 0 μ M) and a 150 minimum of $0.14\pm0.02 \ \mu mol \ (g \ dry \ weight)^{-1}$ from phosphate-rich solution (P = 100 μ M) when 151 152 the MS culture solutions were modified with 4.0 µM of arsenate. The results imply that arsenate 153 uptake into the water fern was significantly higher in phosphate deficient solutions than the phosphate-rich solutions and the increase of phosphate concentration in culture solution 154 155 decreases arsenate uptake. However, arsenic accumulation by the plants was highest (0.03±0.00 μ mol g⁻¹ dry weight) in phosphate sufficient solution (P = 100 μ M) when the initial 156 concentrations of DMAA in growth medium was 4.0 µM. This concentration of arsenic in plant 157 tissue did not differ significantly with the concentration ($0.02\pm00 \mu$ mol g⁻¹ dry weight), when 158 159 the plants were grown in phosphate deficient growth medium ($P = 0 \mu M$). This might be because 160 the DMAA uptake in the aquatic macrophyte was not affected by the initial phosphate161 concentrations in the solution.

162 Phosphate added to the growth medium plays two important roles: i) it enhances arsenate 163 availability in the solution; and, ii) it competes with arsenate for uptake carriers in the 164 plasmalemma due to the similar chemical behavior of arsenate and phosphate [30, 31]. The fact 165 that arsenate and phosphate concentrations in tissues of Salvinia natans L. were significantly 166 negatively correlated (r = -0.662. p < 0.05) (Table 2) suggests that the competition for uptake, 167 indeed, occurred (Fig. 2A). Mkandawire and Dudel [18] also reported that the arsenate uptake in 168 Lemna gibba L. occurs through the phosphate uptake pathway due to similar chemical behavior 169 of arsenate and phosphate.

170 In contrast, DMAA and phosphate concentrations in tissues of *Salvinia natans* L. did not 171 correlate significantly (r = -0.076, p > 0.05) (Fig. 2B). This is because DMAA does not compete 172 with phosphate for plant uptake due to their dissimilar chemical behavior.

173

174 Effect of Arsenic Species on Phosphate Uptake by Salvinia natans L.

Arsenate in the culture solutions significantly (p < 0.05) reduced phosphate uptake in tissues of 175 176 Salvinia natans L. However, the DMAA did not affect phosphate uptake into the plant significantly (p > 0.05). The Pearson correlation analysis (Table 2) revealed a significant 177 178 negative relationship between arsenate and phosphate concentrations in tissues of Salvinia 179 natans L. (Fig. 2A). No significant correlation was observed between DMAA and phosphate 180 concentrations in tissues of Salvinia natans L. (Fig. 2B). Reduction of phosphate uptake in 181 plants exposed to arsenate has also been reported in literatures [31, 32]. This is because the 182 arsenate uptake occurs through the phosphate uptake pathway even replacing the phosphate from sorption site [33]. The DMAA may be accumulated in Salvinia natans L. through different 183 184 mechanisms.

186 Arsenic Removal Efficiency of Salvinia natans L.

187 After 5 days of exposure to culture solutions containing different concentrations of arsenate, the 188 Salvinia natans L. removed a significant amount of arsenic (Fig. 3). Regardless of phosphate 189 concentrations in solution, between 32-65% arsenate was removed from the solution by Salvinia 190 natans L. within the five days for a plant dry biomass of 0.15 g. On the other hand, DMAA 191 removal was negligible (about 0.7-3.2%). The results indicate that removal of arsenic were 192 increased with the increase of arsenate concentrations and decreased with the increase of 193 phosphate concentrations in the solution. Mukherjee et al. [34] reported a 74.8% removal of 194 arsenic by the same plant within 120 hrs of exposure when the initial source of arsenic was 195 arsenate (As(V)).

196

197 Influence of Phosphate and Iron on Arsenic Uptake in Salvinia natans L.

198 Fig. 4 shows the correlation between arsenic and iron concentrations in Salvinia natans L. 199 Arsenate was found to be significantly positively correlated (r = 0.662; p < 0.05) with iron while 200 DMAA was independent of iron concentration (r = 0.233; p > 0.05) (Table 2). Robinson et al. 201 [33] also found a positive correlation between arsenic and iron in native aquatic ferns 202 (Asplenium bulbiferum, Blechum discolor, Histiopteris incisa, Pneumatopteris penningera and 203 Polystichum vestitum) as well as watercress (Rorippa nasturium-aquaticum). This might be due 204 to the physico-chemical adsorption of arsenate on iron oxides on plant surfaces. Robinson et al. 205 [33] discussed the physico-chemical as an alternative mechanism of arsenic accumulation in 206 aquatic plants. In this mechanism, iron oxides (iron plaques) on the plant surfaces adsorb and 207 accumulate arsenic. Although arsenic adsorption on iron oxide plaques on the surface of aquatic 208 plants has been reported by Robinson et al. [33], which species of arsenic predominated in such 209 adsorption was not clear from their studies. However, Blute et al. [35] reported arsenate to be 210 positively correlated with iron plaques on roots of Typha latifolia (cattail) grown in arseniccontaminated wetland sediments. According to Blute et al. [35], the ferric plaques were 211

predominantly Fe(III) oxyhydroxide and 80% of the arsenic in it were arsenate. The present study demonstrates that arsenic adsorbed on the iron plaques of aquatic plant surfaces is mainly arsenate, as it was adsorbed on iron plaques of wetland plant *Typha latifolia* (cattail).

215 Arsenate and iron concentrations in *Salvinia natans* L. were highly positively correlated (p < p216 0.01) when the plants were grown in phosphate-deficient solution while their correlation was not 217 significant (p > 0.05), when the plants were grown in phosphate-sufficient solution. The result 218 suggests that phosphate is adsorbed on iron oxides (Fe-plaques) of aquatic plant surfaces and 219 displace arsenate from the sorption sites on iron oxides. It is well established that iron 220 (hydr)oxides are important phosphate adsorbents in soils [36-39] oxic sediments [40]. The use of 221 Fe oxides to adsorb phosphate on-site and reduce its concentrations in runoff and leachates is a 222 proven approach to potentially lowering phosphate loadings of water bodies [41-43]. Numerous 223 laboratory studies have also been directed at the sorption of phosphate on Fe oxides [44-47]. 224 Some studies have attempted to quantify differences in phosphate adsorption associated with 225 variations in mineral properties such as surface area, morphology, and chemical composition 226 [47, 48]. Ferrihydrite is perhaps the most effective of these minerals in terms of phosphate 227 adsorption in soils due to its small particle size, high surface area, and gel-like form. In nature, 228 ferrihydrite is formed by the rapid oxidation of Fe(II) in Fe-rich waters [49]. Thus, the phosphate 229 provably not only compete with arsenate for uptake carriers in plasmalemma [17] but also 230 compete for adsorption on iron oxides of roots or plant surfaces as the phosphate and arsenate 231 are analogous in chemical properties. The competition between arsenate and phosphate for the 232 adsorption on iron oxides of plant surfaces results in the reduction of physico-chemical 233 adsorption of arsenate in aquatic plants.

234

235 **Conclusion:**

Phosphate and iron are two important nutrient elements affecting the arsenic uptake in water fern *Salvinia natans* L. The *Salvinia natans* L. uptake arsenate probably through symplastic or

apoplastic pathway and compete with phosphate for uptake carriers in plasmalemma. But stronger binding affinity of phosphate with the uptake carriers inhibits arsenate uptake in aquatic plants. However, physicochemical adsorption would be an alternative and potential mechanism for arsenic uptake in aquatic plants. In this mechanism, arsenate is adsorbed by iron oxides on plant surfaces.

243 Although the present study reveals the physicochemical uptake of arsenate in water fern, the 244 individual concentrations of arsenic in plant tissue and iron plaques were not measured. 245 Therefore, it is difficult to interpret how much arsenic and iron was taken up in the plant tissues. 246 It needs microanalysis of the tissues to make the fact clear. But as iron (hydr)oxides are 247 important phosphate adsorbents and the phosphate has stronger binding affinity to the uptake 248 carriers in plasmalemma, low correlation coefficient between arsenate and iron in plants of 249 phosphate-sufficient solution suggest that most of the arsenate might be bound to the outer cell wall rather then entering into the plant tissues. Nevertheless, this does not decrease the 250 251 importance of aquatic macrophytes in arsenic phytoremediation.

252

253 Acknowledgements:

This research was supported partly by Grants-in-Aid for Scientific Research (18510071) from the Japan Society for the Promotion of Science, and the Steel Industry Foundation for the Advancement of Environmental Protection Technology, Japan.

257

258

259 **References:**

[1] M. A. Rahman, H. Hasegawa, K. Ueda, T. Maki, C. Okumura, M. M. Rahman, Arsenic
accumulation in duckweed (*Spirodela polyrhiza* L.): A good option for
phytoremediation. Chemosphere. 69 (2007) 493–499.

- [2] M. A. Fazal, T. Kawachi, E. Ichio, Validity of the latest research findings on causes of
 groundwater arsenic contamination in Bangladesh. Water International. 26 (2001) 380 389.
- [3] A. H. Smith, E.O. Lingas, M. Rahman, Contamination of drinking water by arsenic in
 Bangladesh: a public health emergency. Bull. Of the World Health Organization, 78
 (2000) 1093-1103.
- [4] K. M. Ahmed, Groundwater arsenic contamination in Bangladesh: An overview. In:
 Bhattacharya, P. and Welch, A. H. (Eds.). Arsenic in groundwater of sedimentary
 aquifers. 31st International geological congress, Rio de Janerio, Brazil. (2000) 3-11.
- [5] A. K. Chakraborti, D. K. Das, Arsenic pollution and its environmental significance. J.
 Interacad 1 (1997) 262-276.
- [6] D. M. Banerjee, Some comments on the source of arsenic in the Bengal Deltaic sediments.
 In: Bhattacharya, P. and Welch, A. H. (Eds.). Arsenic in groundwater of sedimentary aquifers. 31st International geological congress, Rio de Janerio, Brazil. (2000) 15-17.
- [7] P. Visoottiviseth, K. Francesconi, W. Sridokchan, The potential of Thai indigenous plant
 species for the phytoremediation of arsenic contaminated land. Environ. Poll. 118 (2002)
 453-461.
- 280 [8] P. O'Neill, Arsenic, In: Heavy metals in soils. B. J. Alloway (Ed.). (1995) 105-121.
- [9] P. L. Smedley, D. G. Kinniburgh, A review of the source, behaviour and distribution of
 arsenic in natural waters. Appl. Geochem. 17 (2002) 517-568.
- [10] I.Raskin, P. B. A. Nanda-Kumar, S. Dushenkov, D. E. Salt, B. D. Ensley, Removal of
 radionuclides and heavy metals from water and soil by plants. OECD Document,
 Bioremediation. (1994) 345-354.
- [11] T. De Koe, *Agrostic castellana* and *Agrostis delicatula* on heavy metal and arsenic enriched
 sites in NE Portugal. Sci. Total. Envi. 145 (1994) 103-109.

- [12] J. Bech, C. Poschenrieder, M. Llugany, J. Barcelo, P. Tume, F.J. Toloias, As and heavy
 metal contamination of soil and vegetation around a copper mine in Northern Peru. Sci.
 Total Envi. 203 (1997) 83-91.
- [13] L. Q. Ma, K. M. Komar, C. Tu, W. Zhang, Y. Cai, E. D. Kennelley, A fern that
 hyperaccumulates arsenic. Nature. 409 (2001) 579.
- [14] P. A. Gulz, S. K. Gupta, R. Schulin, Arsenic accumulation of common plants from
 contaminated soils. Plant and soil. 272 (2005) 337-347.
- [15] K. Komar, L. Q. Ma, D. Rockwood, A. Syed, Identification of arsenic tolerant and
 hyperaccumulating plants from arsenic contaminated soils in Florida. Agron. Abstr.
 (1998) 343.
- [16] C. Tu, L. Q. Ma, B. Bondada, Arsenic accumulation in the hyperaccumulator Chinese brake
 and its utilization potential for phytoremediation. J. Environ. Qual. 31 (2002) 1671-1675.
- 300 [17] M. Mkandawire, Y. V. Lyubun, P. V. Kosterin, E. G. Dudel, Toxicity of arsenic species to
 301 *Lemna gibba* L. and the influence of phosphate on arsenic bioavailability. Environ.
 302 Toxicol. 19 (2004) 26-35.
- 303 [18] M. Mkandawire, E. G. Dudel, Accumulation of arsenic in Lemna gibba L. (ducweed) in
- tailing waters of two abandoned uranium mining sites in Saxony, Germany. Sci. Total
 Environ. 336 (2005) 81-89.
- 306 [19] C. K. Lee, K. S. Low, N. S. Hew, Accumulation of arsenic by aquatic plants. Sci. Total
 307 Environ. 103 (1991) 215-227.
- 308 [20] B. Robinson, C. Duwing, N. Bolan, M. Kannathasan, A. Saravanan, Uptake of arsenic by
 309 New Zeland watercress (*Lepidium sativum* L.). Sci. Total Environ. 301 (2003) 67-73.
- 310 [21] O. I. Sizova, V. V. Kochetkov, S. Z. Validov, A. M. Boronin, P. V. Kosterin, Y. V.
- 311 Lyubun, Arsenic-contaminated soils: genetically modified *Pseudomonas* spp. and their
- arsenic-phytoremediation potential. J. Soils and Sediments. 2 (2002) 19-23.

- 313 [22] P. Selvapathy, P. Sreedhar, Heavy metals removal by water hyacinth. J. Indian Publ. Health
 314 Engr. 3 (1991) 11-17.
- 315 [23] B. Alam, A. K. Chatterjee, S. Duttagupta, Bioaccumulation of Cd (II) by water lettuce.
 316 Pollut. Res. 14 (1995) 59-64.
- 317 [24] A. K. Sen, N. G. Mondal, *Salvinia natans* as the scavenger of Hg (II). Wat. Air Soil Pollut.
 318 34 (1987) 439-446.
- 319 [25] A. K. Sen, M. Bhattacharyya, Studies on uptake and toxic effects of lead on *Salvinia* 320 *natans*. Indian J. Environ. Health. 35 (1993) 308-320.
- [26] K. S. Low, C. K. Lee, C. H. Tai, Biosorption of copper by water hyacinth roots. J. Environ.
 Sci. Health A. 29(1) (1994) 171-188.
- 323 [27] N. W. Ingole, J. P. Ting, Study on nutrient removal potential of selected aquatic
 324 macrophytes. J. Inst. Engr (India) Environ. Engng Div. 83 (2002) 1-6.
- 325 [28] A. K. Sen, N. G. Mondal, Removal and uptake of copper (II) by *Salvinia natans* from
 326 wastewater. Wat. Air Soil Pollut. 49 (1990) 1-6.
- 327 [29] S. C. Lenore, E. G. Arnold, D. E. Andrew, (Eds.), Standard methods for the examination of
 328 water and wastewater, 20th ed. APHA, AWWA and WEF. 1998.
- [30] C. Tu, L.Q. Ma, Effects of arsenic and phosphate on their accumulation by an arsenic hyperaccumulator *Pteris vittata* L. Plant Soil, 249 (2003) 373-382.
- [31] J. Wang, F. J. Zhao, A. A. Meharg, A. Raab, J. Feldman, S. P. McGrath, Mechanism of
 arsenic hyperaccumulation in *Pteris vittata* L. uptake kinetics, interaction with phosphate
 and arsenic speciation. Plant Physiol. 130 (2002) 1552-1561.
- [32] M. Patra, N. Bhowmil, B. Bandopadhyay, A. Sharma, Comparison of mercury, lead and
 arsenic with respect to genotoxic effects on plant systems and the development of genetic
 tolerance. Environ. Exp. Bot. 52 (2004) 199-223.

- 337 [33] B. Robinson, N. Kim, M. Marchetti, C. Moni, L. Schroeter, C. van den Dijssel, G. Milne,
 338 B. Clothier, Arsenic hyperaccumulation by aquatic macrophytes in the Taupo Volcanic
 339 Zone, New Zealand. Environ Exper Bot. 58 (2006) 206-215.
- 340 [34] S. Mukherjee, S. Kumar, Adsorptive uptake of arsenic (V) from water by aquatic fern
 341 Salvinia natans. J. Water Supply: Research and Technology- AQUA. 54(1) (2005) 47-53.
- 342 [35] N. K. Blute, D. J. Brabander, H. F. Hemond, S. R. Sutton, M. G. Newville, M. L. Rivers,
- 343 Arsenic sequestration by ferric iron plaque on cattail roots. Environ. Sci. Technol. 38
 344 (2004) 6074-6077.
- 345 [36] G. Guzman, E. Alcantara , V. Barron, J. Torrent, Phytoavailability of phosphate adsorbed
 346 on ferrihydrite, hematite, and goethite. Plant and Soil. 159 (1994) 219-225.
- 347 [37] U. Schwertmann, R. M. Taylor, Iron oxides. In: Minerals in Soil Environments (2nd ed.),
- 348 B. Dixon, S. B. Weed (Eds.), Soil Sci. Soc. Am. Madison, Wisconsin, USA. (1989) 378349 438.
- [38] O. K. Borggaard, S. S. Jdrgensen, J. P. Moberg, B. Raben-lange, Influence of organic
 matter on phosphate adsorption by aluminium and iron oxides in sandy soils. European J.
 Soil Science. 41 (1990) 443–449.
- 353 [39] M. E. Hamad, D. L. Rimmer, J. K. Syers, Effect of iron oxide on phosphate sorption by
 354 calcite and calcareous soils. European J. of Soil Science. 43 (1992) 273–281.
- [40] M. D. Krom, R. A. Berner, Adsorption of phosphate in anoxic marine sediments. Limnol.
 Oceanogr. 25 (1980) 797-806.
- [41] P. A. Moboornea, D. M. Miller, Decreased phosphorus solubility in poultry litter with
 aluminum, calcium, and iron amendments. J. Environ. Qual. 23 (1994) 325–330.
- [42] H. A. Elliott, G. A. O'Connor, P. Lu, S. Brinton, Influence of water treatment residuals on
 phosphorus solubility and leaching. J. Environ. Qual. 31 (2002) 1362–1369.

- 361 [43] L. E. Gallimore, N. T. Basta, D. E. Storm, M. E. Payton, R. H. Huhnke, M. D. Smolen,
- Water treatment residual to reduce nutrients in surface runoff from agricultural land. J.
 Environ. Qual. 28 (1999) 1474–1478.
- 364 [44] R. L. Parfitt, Anion adsorption by soils and soil materials. Adv. Agron. 30 (1978) 1–50.
- 365 [45] J. B. Harrison, V.B. Berkheiser, Anion interactions with freshly prepared hydrous iron
 366 oxides. Clays Clay Miner. 30 (1982) 97–102.
- [46] C. C. Ainsworth, M. E. Sumner, V. J. Hurst, Effect of aluminum substitution in goethite on
 phosphorus adsorption: I. Adsorption and isotopic exchange. Soil Sci. Soc. Am. J. 49
 (1985) 1142–1149.
- 370 [47] V. Barron, M. Herruzo, J. Torrent, Phosphate adsorption by aluminous hematites of
 371 different shapes. Soil Sci. Soc. Am. J. 52 (1988) 647–651.
- 372 [48] J. Torrent, V. Barron, U. Schwertmann, Phosphate adsorption and desorption by goethites
 373 differing in crystal morphology Soil Sci. Soc. Am. J. 54 (1990) 1007–1012.
- 374 [49] F. E. Rhoton, J. M. Bigham, D. L. Lindbo, Properties of iron oxides in streams draining the
- loess uplands of Mississippi. Appl. Geochem. 17 (2002) 409–419.
- 376
- 377
- 378
- 379
- 380
- 381
- 382
- 383
- 384

Table 1: Modified ^a Murashige and Skoog (MS) culture solution used for Salvinia nate	ıns L.
cultivation.	

Nutrients	Concentrations (mg l ⁻¹)
KNO ₃	1900
NH ₄ NO ₃	1650
$CaCl_2 \cdot 2H_2O$	440
MgSO ₄ .7H ₂ O	370
K ₂ HPO ₄	Modified ^a
FeSO ₄ ·7H ₂ O	27.80
MnSO ₄ ·5H ₂ O	22.30
$ZnSO_4 \cdot 7H_2O$	8.60
H_3BO_3	6.20
KI	0.83
$Na_2MoO_4 \cdot 2H_2O$	0.25
$CuSo_4 \cdot 5H_2O$	0.025
CoCl ₂ ·6H ₂ O	0.025
Na ₂ -EDTA	37.30

390 ^a The control culture solution did not contain phosphate. The other solutions were modified either with 50 or $100 \,\mu\text{M}$ of phosphate.

Table 2: Pearson correlations co-efficient (r) between arsenic (arsenate and DMAA) and
 phosphate; arsenic (arsenate and DMAA) and iron concentrations in *Salvinia natans* L.
 401

Exposure time	Pearson Correlation (r)	Significance (<i>p</i>)
As(V) & P	-0.662*	0.019
DMAA & P	-0.076	0.814
As(V) & Fe	0.662*	0.019
DMAA & Fe	0.233	0.466

* Correlation is significant at the 0.05 level

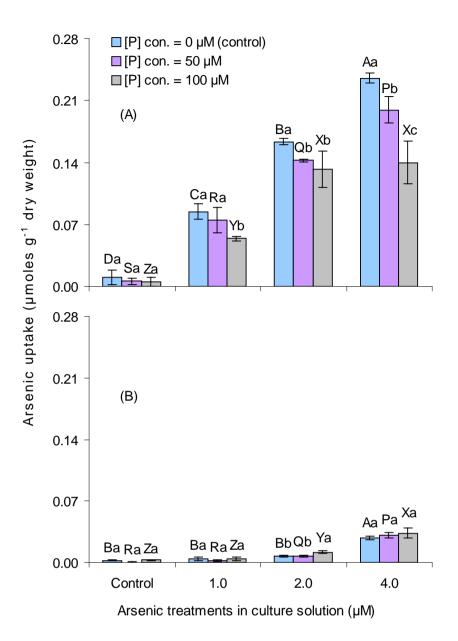




Figure 1: Arsenic uptake in Salvinia natans L. affected by the phosphate concentrations in culture solution. Error bars represent \pm S.D. (n = 3). Arsenate (A); DMAA (B). Different lowercase letters indicate statistically significant differences (p < 0.05) between phosphate treatments and different uppercase letters indicate statistically significant differences (p < 0.05) between different arsenic treatments.

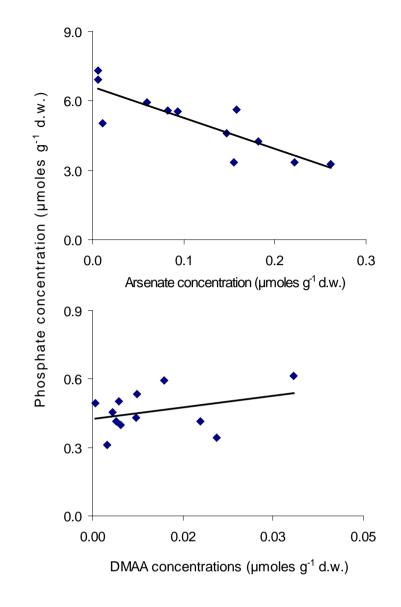
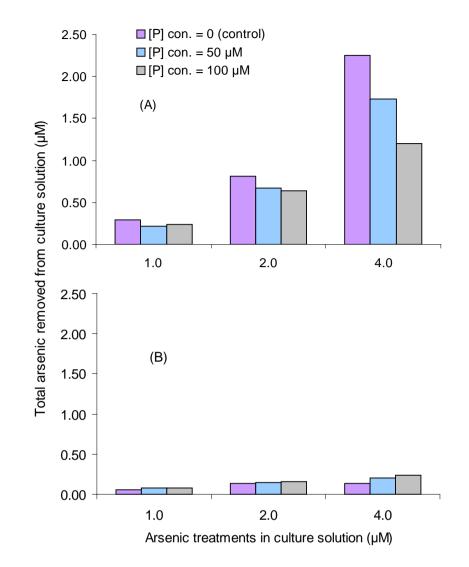


Figure 2: Correlation between arsenic and phosphate in of *Salvinia natans* L.





450 Figure 3: Arsenic removal efficiency of *Salvinia natans* L. from culture solutions containing different
451 phosphate concentrations. The duration of exposure was 5 days. Arsenate (A); DMAA (B).

- ...

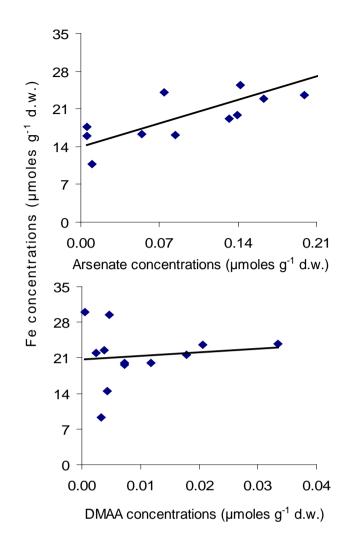


Figure 4: Correlation between arsenic and iron in *Salvinia natans* L.