

Elsevier required licence: © 2021

This manuscript version is made available under the  
CC-BY-NC-ND 4.0 license

<http://creativecommons.org/licenses/by-nc-nd/4.0/>

The definitive publisher version is available online at

<https://doi.org/10.1016/j.chemosphere.2021.130910>

---

1 **Free ammonia pretreatment enhances the removal of antibiotic resistance genes in**  
2 **anaerobic sludge digestion**

3  
4 Zehao Zhang <sup>a</sup>, Xuan Li <sup>b</sup>, Huan Liu <sup>a,c</sup>, Li Gao <sup>d</sup>, Qilin Wang <sup>a,\*</sup>  
5

6 <sup>a</sup> Centre for Technology in Water and Wastewater, School of Civil and Environmental  
7 Engineering, University of Technology Sydney, Ultimo, NSW 2007, Australia

8 <sup>b</sup> School of Civil, Mining & Environmental Engineering, University of Wollongong, NSW  
9 2522, Australia

10 <sup>c</sup> State Key Laboratory of Pollution Control and Resources Reuse, College of Environmental  
11 Science and Engineering, Tongji University, Shanghai 200092, China

12 <sup>d</sup> South East Water, 101 Wells Street, Frankston, VIC 3199, Australia  
13

14  
15 \*Corresponding author.

16 E-mail: [Qilin.Wang@uts.edu.au](mailto:Qilin.Wang@uts.edu.au) (Q. Wang)  
17  
18  
19  
20  
21  
22  
23  
24  
25

---

26 **ABSTRACT:** Sludge has been recognized as a reservoir of antibiotic resistance genes (ARGs)  
27 in the wastewater treatment plants. Our previous study has demonstrated that free ammonia  
28 (FA, i.e., NH<sub>3</sub>-N) pretreatment is an effective method for enhancing anaerobic digestion of  
29 sludge. However, the effect of FA pretreatment on the removal of ARGs in the anaerobic sludge  
30 digestion is still unknown. In this study, several ARGs representing various antibiotic classes  
31 and integrase gene (*intI1*) which is crucial for horizontal transfer of ARGs were chosen. This  
32 study demonstrated for the first time that combined FA pretreatment (420 mg NH<sub>3</sub>-N/L for 24  
33 h) and anaerobic digestion could enhance the removal of *aac(6')-Ib-cr*, *blaTEM*, *sul2*, *tetA*, *tetB*  
34 and *tetX* from sludge by 17~74% compared with anaerobic digestion without FA pretreatment,  
35 resulting in a lower ARGs abundance in the anaerobically digested sludge. This is caused by  
36 the removal of tested ARGs during FA pretreatment and the reduced abundance of potential  
37 bacterial hosts of ARGs due to FA pretreatment during anaerobic digestion. The removal of  
38 **IntI1** was not significantly affected by FA pretreatment and *intI1* only had a significant  
39 correlation with one ARG *sul1* in this study, indicating that *intI1* did not play a large role in the  
40 fate of the tested ARGs in this study. This study indicated that FA pretreatment for anaerobic  
41 digestion could potentially reduce the spread of ARGs from the sludge to the natural  
42 environment during sludge disposal or reuse.

43

#### 44 **Keywords**

45 Antibiotic resistance genes; Free ammonia; Anaerobic digestion; Wastewater treatment plants

46

47

48

49

50

---

51 **1. Introduction**

52 Antibiotics have been widely used for human and livestock to treat infectious diseases and  
53 promote the growth of livestock (Zhuang et al., 2015). Though the effectiveness of antibiotics  
54 has significantly benefited mankind, the intensive use of antibiotics has led to the spread of  
55 antibiotic resistance among microorganisms. Antibiotic resistance genes (ARGs), as the main  
56 reason for microorganisms to be able to withstand the bacteriostatic or bactericidal effects of  
57 antibiotics (Martínez et al., 2014), have been widely found in soil, surface water, groundwater,  
58 PM 2.5, and even deep ocean sediments (Allen et al., 2010; Brown and Balkwill, 2009; Ouyang  
59 et al., 2020). The spread of ARGs not only posed a global threat to the public well-being, but  
60 also affected the development of industries such as veterinary medicine and agriculture (Teuber,  
61 2001).

62

63 Generally, the spread of ARGs is considered a result of natural selection when microorganisms  
64 are exposed to antibiotic compounds (Negreanu et al., 2012). In unaltered states, up to 95% of  
65 antibiotics can be excreted from humans or animals, which enter into the sewers and are  
66 eventually collected by wastewater treatment plants (WWTPs) (Negreanu et al., 2012). This  
67 alters the selective pressure and leads to the occurrence of ARGs in the WWTPs microbial  
68 community. Furthermore, the ARGs can disseminate through vertical gene transfer by cell  
69 reproduction, and/or horizontal gene transfer by mobile genetic elements in the WWTPs  
70 microbial community (Shao et al., 2018; Xue et al., 2019).

71

72 Recent studies suggested that sludge from WWTPs might be a crucial source of ARGs to the  
73 environment because most of the ARGs in a WWTP eventually end up in the sludge (Xue et  
74 al., 2019; Yang et al., 2014). In particular, ARGs in sludge can reach a very high concentration.  
75 For instance, the abundances of tetracycline resistance genes *tetA* and *tetX* in sludge reached

---

76  $1 \times 10^{12}$  and  $1 \times 10^{13}$  gene copies/g-TS (total solids), respectively (Auerbach et al., 2007), which  
77 is about eight orders of magnitude higher than that of soil samples ( $1 \times 10^4$  -  $1 \times 10^5$  gene  
78 copies/g-TS) (Duan et al., 2017; Nölvak et al., 2016). Therefore, the existence of ARGs in  
79 sludge may lead to the spread of the ARGs from sludge to soil, considering a large amount of  
80 sludge is reused for agriculture in many regions such as Australia (>65%) and the United States  
81 (>50%) (Australian and New Zealand Biosolids Partnership, 2019; the United States  
82 Environmental Protection Agency, 2019). The potential spread of ARGs may have an adverse  
83 effect on the environment (e.g., soil) and increase the risk of sludge reuse to human health  
84 (Chen et al., 2016; Ross and Topp, 2015).

85

86 Anaerobic digestion is a typical sludge treatment method, which achieves sludge reduction and  
87 energy recovery (Batstone et al., 2002). Recently, additional benefits of anaerobic digestion  
88 have been observed with ARGs removal, where the total abundances of *tetA*, *tetB*, *tetC*, *tetW*,  
89 *tetX*, *sul1* and *sul2* were reduced in anaerobic digestion (Ma et al., 2011; Pei et al., 2016).  
90 However, the efficiency of anaerobic digestion is often limited by the poor biodegradability of  
91 the sludge (Carrère et al., 2010). Thus, various pretreatment methods such as ultrasonic and  
92 thermal pretreatment have been applied to enhance sludge biodegradability and sludge  
93 reduction (Bougrier et al., 2006). Additional benefits of pretreatment have been observed  
94 recently on the removal of ARGs. For instance, ultrasonic pretreatment enhanced the removal  
95 of the targeted ARGs by 50% during anaerobic digestion and further reduced the abundance of  
96 the bacterial hosts of ARGs (Wang et al., 2019).

97

98 Free ammonia (FA., i.e.,  $\text{NH}_3$ ) pretreatment was found as an economically attractive and  
99 environmentally friendly approach to enhance the anaerobic digestion efficiency, as it only  
100 relies on a by-product (i.e., FA) of wastewater treatment (Wang, 2017; Wei et al., 2017a). Our

---

101 previous study demonstrated that FA pretreatment at 420 mg N/L for 24 h enhanced sludge  
102 biodegradability by 20% (Wei et al., 2017a). However, the effect of FA pretreatment on the  
103 removal of ARGs in anaerobic sludge digestion is still largely unknown. Understanding this  
104 effect will be beneficial and essential for the practical application of the FA pretreatment  
105 strategy.

106

107 This study aimed to assess the effect of FA pretreatment on the abundances of ARGs in  
108 anaerobic sludge digestion. The full-scale secondary sludges with and without FA pretreatment  
109 were subjected to anaerobic digestion tests. Nine ARGs representing various antibiotic classes  
110 were quantified by real-time quantitative polymerase chain reaction (RT-qPCR). The integrase  
111 gene (*intI1*) of Class 1 integrons was also quantified as a representative of mobile genetic  
112 elements which could affect the horizontal transfer of ARGs. The abundances of ARGs and  
113 *intI1* in FA-pretreated sludge were quantified to evaluate the fate of ARGs and *intI1* during the  
114 FA pretreatment. Microbial community analysis and the correlation between ARGs and  
115 *intI1*/microbial community were also determined to reveal the abundance change of the  
116 potential bacterial hosts of ARGs, which aimed to further understand the potential mechanisms  
117 underpinning the effect of FA pretreatment on the removal of ARGs in anaerobic sludge  
118 digestion.

119

## 120 **2. Materials and methods**

### 121 **2.1 Sludge sources**

122 Both secondary sludge and inoculum were used to conduct the experiments. The secondary  
123 sludge was collected from the thickener of a WWTP conducting biological nitrogen and  
124 phosphorus removal. The WWTP has a sludge retention time (SRT) of 12 - 16 d. The inoculum  
125 was collected from a lab-scale mesophilic anaerobic digester receiving real secondary sludge

---

126 from the same WWTP. The mesophilic anaerobic digester has an SRT of 15 d. The secondary  
127 sludge and inoculum were used for the anaerobic digestion tests to be described below.

128

## 129 **2.2 Pretreatment of secondary sludge with FA**

130 Batch experiments were conducted to evaluate the impact of FA pretreatment on the  
131 abundances of ARGs and *intI1*. One liter of secondary sludge was evenly distributed into two  
132 batch reactors as control and experimental reactors, respectively. For the experimental reactor  
133 with FA pretreatment, a certain amount of ammonium stock solution (3.0 M) was added to  
134 obtain an ammonia concentration of 500 mg N/L. pH was adjusted and maintained at  $10.0 \pm$   
135  $0.1$  using NaOH solution. The pretreatment was conducted in a temperature-controlled room  
136 ( $25\text{ }^{\circ}\text{C}$ ) with a stirring speed of 500 rpm using a magnetic stirrer for 24 h. The total ammonia  
137 nitrogen, pH and the temperature collectively resulted in an FA concentration of 420 mg  $\text{NH}_3$ -  
138 N/L, which was determined by the formula  $S_{(\text{NH}_4^+-\text{N}+\text{NH}_3-\text{N})} \times 10^{\text{pH}}/(\text{K}_b/\text{K}_w+10^{\text{pH}})$  (Anthonisen et  
139 al., 1976). The  $S_{(\text{NH}_4^+-\text{N}+\text{NH}_3-\text{N})}$  is the total ammonia nitrogen concentration. The  $\text{K}_b/\text{K}_w$  is equal  
140 to  $e^{6.344/(273+T)}$ . This ammonia concentration (i.e., 420 mg  $\text{NH}_3$ -N/L) was selected based on our  
141 previous study, which demonstrated that FA pretreatment at 420 mg  $\text{NH}_3$ -N/L for 24 h led to  
142 the highest methane production potentially with a large economic advantage (Wei et al., 2017a).  
143 The control reactor was set up without ammonium addition or pH control. For both reactors,  
144 sludge samples were taken both before and after pretreatment for the determination of ARGs  
145 and *intI1* using RT-qPCR, as described below.

146

## 147 **2.3 Anaerobic digestion tests**

148 Anaerobic digestion tests were performed to evaluate the effect of FA pretreatment on the  
149 removal of ARGs during anaerobic sludge digestion. The serum vials (160 mL) with a working  
150 volume of 100 mL were used for the anaerobic digestion tests. The inoculum and the secondary

---

151 sludges with and without FA pretreatment were added into two separate serum vials, resulting  
152 in a VS (volatile solids) based inoculum to sludge ratio of approximately 2.0. The vials were  
153 flushed with helium gas for 2 min (1 L/min) to ensure an anaerobic condition. After that, a  
154 rubber stopper with an aluminum crimp cap was used to seal the vials. The sealed vials were  
155 then put into an incubator operated at 37 °C. Blank was also set up, which only contained  
156 inoculum and MilliQ water (i.e., without secondary sludge). The anaerobic digestion tests  
157 lasted for 45 days. At the end of the anaerobic digestion tests, the anaerobically digested sludges  
158 with and without FA pretreatment as well as blank were sampled to analyze ARGs, *intI1*, and  
159 microbial community as described in the following sections.

160

#### 161 **2.4. Quantification of ARGs and *intI1***

162 Sludge samples were collected into centrifuge tubes and then centrifuged at 10,000 rpm and  
163 4 °C for 10 min to collect the pellet for DNA extraction. 0.25 g pellet of each sample was used  
164 for DNA extraction using the Fast DNA Spin Kit for Soil (MP Biomedicals, USA) according  
165 to the manufacture's instruction. The integrity of extracted DNA was measured by gel  
166 electrophoresis (1% agarose). The concentration and purity of extracted DNA were confirmed  
167 by a NanoDrop ND-1000 (NanoDrop, USA).

168

169 RT-qPCR was used to quantify the abundance of the target genes in this study. One  
170 aminoglycoside and fluoroquinolone resistance gene (*aac(6')-Ib-cr*), one beta-lactamase  
171 resistance gene (*blaTEM*), two sulfonamide resistance genes (*sul1* and *sul2*), and five  
172 tetracycline resistance genes (*tetA*, *tetB*, *tetG*, *tetM* and *tetX*) were chosen to represent ARGs  
173 in this study. The selected ARGs represent various antibiotic classes. The *intI1* was also  
174 included in the quantification to represent the change of mobile genetic elements. The *intI1* is  
175 prevalent in the bacterial community and plays a key role in the spread of ARGs (Amos et al.,



---

176 2018). The 16s rRNA gene was also quantified to represent the total bacterial biomass in the  
177 sludge (Burch et al., 2013). Each DNA sample was amplified in triplicate. The detailed  
178 information of annealing temperatures, primers and q-PCR reaction matrix of target genes is  
179 listed in the supporting information (Tables S1 and Text S1, Supporting Information).

180

181 In this study, the absolute abundances of the target genes were normalized to the gram of TS  
182 (i.e., gene copies/g-TS). The relative abundances of the target genes were normalized to 16S  
183 rRNA genes (i.e., gene copies/16s rRNA) as an indicator of the proportion of bacteria carrying  
184 ARGs and *intI1*.

185

## 186 **2.5 Microbial community analysis**

187 The microbial community was analyzed by a high-throughput sequencing method. PCR  
188 primers 341F (5'-CCTACGGGAGGCAGCAG-3') and 806R (5'-  
189 ATTACCGCGGCTGCTGG-3') were used to amplify the V3–V4 regions of 16S rRNA genes.  
190 The extracted DNA in each sample was amplified in triplicate in PCR to acquire amplicons  
191 (see Tables S1 for the details of PCR). The amplicons were further purified with GeneJET™  
192 Gel Extraction Kit (Thermo Scientific, USA). A fluorometer of Qubit 2.0 (Thermo Scientific,  
193 USA) was used to determine the concentrations of the amplicons. Amplicons were then  
194 subjected to small fragment library construction and pair-end sequencing using the Illumina  
195 NovaSeq PE250 platform (Illumina, USA).

196

197 The sequencing reads of each sample was performed under specific filtering conditions to  
198 obtain the high-quality reads according to the Cutadapt (v1.9.1) quality-controlled process  
199 (Martin, 2011). Subsequently, the PCR chimeras were filtered using UCHIME (v4.2) algorithm

---

200 (Edgar et al., 2011) by comparing it with Silva (v132) database (Quast et al., 2012). Then, the  
201 clean reads of each sample were finally obtained.

202  
203 Sequences analysis was performed by UPARSE software (v7.0.1001) (Edgar, 2013).  
204 Sequences with  $\geq 97\%$  similarity were assigned to the same operational taxonomic units  
205 (OTUs). The representative sequence for each OTU was screened for further annotation. For  
206 each representative sequence, the Silva (v132) database was used based on the Mothur  
207 algorithm to annotate taxonomic information (Quast et al., 2012).

208

## 209 **2.6 Data analysis**

210 The correlation between ARGs and *IntI1*/microbial community was determined by spearman's  
211 rank correlation coefficient (R) using SPSS 25.0 (IBM, USA), which is a powerful tool for  
212 providing new insights into ARGs and their potential bacterial hosts in complex environmental  
213 examples (Song et al., 2017). A correlation between bacterial genus and an ARG with a  $R >$   
214  $0.8$  and  $p < 0.05$  was considered as a potential bacterial host for the selected ARGs (Li et al.,  
215 2015). Visualization of the correlation between ARGs and *IntI1*/microbial community was  
216 produced by Matlab R2020a (MathWorks, USA) platform. The abundance of bacteria was  
217 normalized to gram of TS to obtain its absolute abundance and was also visualized using  
218 heatmap by Heml (v1.0).

219

## 220 **3 Results**

### 221 **3.1 Effects of FA pretreatment on the fate of ARGs and *intI1* in the sludge prior to** 222 **anaerobic digestion**

223 The abundances of ARGs and *intI1* in the sludge with and without FA pretreatment were  
224 quantified to evaluate the effect of FA pretreatment on the fate of ARGs and *intI1*. The absolute

---

225 abundances of *aac(6')-Ib-cr*, *blaTEM*, *sul1*, *sul2*, *tetA*, *tetB*, *tetG*, *tetM*, *tetX* and *intI1* were  
226 ranged from  $7.8 \times 10^6$  to  $3.4 \times 10^9$  gene copies/g-TS (Fig. 1). FA pretreatment decreased the  
227 absolute abundances of *blaTEM*, *sul1*, *sul2*, *tetA*, *tetB* and *tetX* by 21%, 8%, 16%, 73%, 38%  
228 and 76%, respectively (Fig. 2). Although *tetA*, *tetB*, *tetX*, *tetM* and *tetG* all belong to the  
229 tetracycline resistance genes, different impacts of FA pretreatment were observed on *tetM* and  
230 *tetG* (Figs. 1 and 2). FA pretreatment increased the absolute abundance of *tetM* by more than  
231 three times, while an insignificant change was observed in the absolute abundance of *tetG*. The  
232 insignificant abundance change of *tetG* could be attributed to the fact that *tetG* fragment was  
233 more conservative to FA pretreatment in comparison to *tetA*, *tetB* and *tetX* (Zhang and Zhang,  
234 2011). The reason for the increased absolute abundance of *tetM* during FA pretreatment  
235 required further study. In addition, FA pretreatment did not significantly ( $p > 0.05$ ) affect the  
236 absolute abundances of *aac(6')-Ib-cr* and *intI1* (Figs. 1 and 2).

237

238 (Position for Fig. 1)

239 (Position for Fig. 2)

240

241 Overall, FA pretreatment reduced the total absolute abundance of the tested ARGs by ~10% in  
242 the sludge, indicating that some ARGs could be removed by FA directly in the sludge  
243 pretreatment. In contrast, the absolute abundance of *intI1* in the sludge was not significantly  
244 affected by FA pretreatment.

245

246 To determine the proportion of bacteria carrying ARGs and *intI1*, ARGs were normalized to  
247 16S rRNA gene abundance to get the relative abundance (Fig. 3). Similar trends to the absolute  
248 abundances were found in the relative abundances of ARGs and *intI1* (Fig. 3). FA pretreatment  
249 decreased the relative abundances of *blaTEM*, *sul1*, *sul2*, *tetA*, *tetB* and *tetX* by 15~76%. In

---

250 contrast, FA pretreatment increased the relative abundance of *tetM* by more than three times  
251 (Fig. 3). In terms of the relative abundances of *aac(6')-Ib-cr*, *tetG* and *intI1*, FA pretreatment  
252 showed insignificant impact ( $p>0.05$ ). In total, FA pretreatment slightly decreased the relative  
253 abundance of the total tested ARGs by ~10%. These results indicate that FA pretreatment could  
254 reduce the proportion of the targeted ARGs in the biomass but had a negligible effect on *intI1*.

255

256 (Position for Fig. 3)

257

### 258 **3.2 FA pretreatment enhanced the removal of ARGs in anaerobic sludge digestion**

259 The absolute abundances of ARGs and *intI1* in untreated sludge and anaerobically digested  
260 sludge without FA pretreatment were compared in Fig. 1. Compared with untreated sludge,  
261 anaerobic digestion reduced the absolute abundance of *aac(6')-Ib-cr*, *sul1*, *sul2*, *tetA*, *tetB*, *tetG*,  
262 *tetX* and *intI1* by 22~89% (Fig. 4), resulting in an overall reduction of about 30% in the total  
263 absolute abundance of the tested ARGs. It is evident that FA pretreatment enhanced the removal  
264 of ARGs in anaerobic sludge digestion. Compared with untreated sludge, the absolute  
265 abundance of *aac(6')-Ib-cr*, *sul1*, *sul2*, *tetA*, *tetB*, *tetG*, *tetX* and *intI1* was removed by 25~95%  
266 in the anaerobically digested sludge with FA pretreatment (Fig. 4). Overall, combined FA  
267 pretreatment and anaerobic digestion could reduce the total absolute abundance of the tested  
268 ARGs by 40%, from  $6.0 \times 10^9$  gene copies/g-TS in the untreated sludge to  $3.6 \times 10^9$  gene copies/g-  
269 TS in the anaerobically digested sludge with FA pretreatment.

270

271 (Position for Fig. 4)

272

273 In particular, the removal of *aac(6')-Ib-cr*, *blaTEM*, *sul2*, *tetA*, *tetB* and *tetX* was enhanced by  
274 17%, 58%, 19%, 52%, 42% and 74%, respectively, in the anaerobically digested sludge with

---

275 FA pretreatment, in comparison to anaerobically digested sludge without FA pretreatment (Fig.  
276 5). In addition, FA pretreatment did not ( $p>0.05$ ) significantly affect the abundances of *sull*,  
277 *tetM* and *intII*, and slightly increased the abundance of *tetG* by 14% in the anaerobically  
278 digested sludge (Fig. 5). Overall, FA pretreatment enhanced the removal of the total absolute  
279 abundance of the tested ARGs by ~15% in the anaerobic digestion.

280

281 (Position for Fig. 5)

282

283 The effects of FA pretreatment on the relative abundances of the tested ARGs and *intII* in the  
284 anaerobic digestion were shown in Fig. 3. The results were consistent with the trend of absolute  
285 abundance. Compared with the anaerobically digested sludge without FA pretreatment, FA  
286 pretreatment reduced the relative abundances of *aac(6')-Ib-cr*, *blaTEM*, *sul2*, *tetA*, *tetB* and  
287 *tetX* by 20~75% in the anaerobically digested sludge. On the contrary, FA pretreatment  
288 increased the relative abundances of *tetG* by 14% in the anaerobically digested sludge. The  
289 relative abundance of *tetM*, *sul* and *intII* remained unchanged ( $p>0.05$ ) at around  $2.5\times 10^{-4}$ ,  
290  $9.2\times 10^{-3}$  and  $7.7\times 10^{-4}$  gene copies/16s rRNA. Overall, FA pretreatment decreased the relative  
291 abundance of the total tested ARGs in the digested sludge by 15%. These indicate that FA  
292 pretreatment could decrease the proportion of the targeted ARGs in the anaerobically digested  
293 biomass, but does not significantly affect *intII*.

294

### 295 **3.3. Correlation between ARGs and *intII*/microbial community and effect of FA** 296 **pretreatment on microbial community in anaerobic sludge digestion**

297 The correlation between the nine tested ARGs and *intII*/microbial community (at the genus  
298 level) was shown in Fig. 6. The results suggested that fourteen bacteria (among the top 30 in  
299 the absolute abundance) had a significant positive correlation with nine ARGs selected in this

---

300 study ( $R>0.8$ ,  $p<0.05$ ) (Fig. 6). *TetB* and *tetX* shared seven potential bacterial hosts, i.e.,  
301 *Thauera*, *unidentified\_Nitrosomonadaceae*, *Halomonas*, *Acidaminobacter*, *Sulfuritalea*,  
302 *Aquabacterium* and *Microbacterium*. In addition, *tetA* and *sul2* shared most of their potential  
303 bacterial hosts with *tetB* and *tetX*, except *Anaeromyxobacter* and *Sulfuritalea*. Four bacterial  
304 hosts, i.e., *Thauera*, *Aquabacterium*, *Microbacterium* and *Sulfuritalea* were identified for *sull*.  
305 *TetM* had three potential bacterial hosts, which were *Methanoculleus*, *Smithella* and  
306 *Methanosarcina*. In addition, *aac(6')-ib-cr*, *blaTEM* and *tetG* each had one sole potential  
307 bacterial host, which were *unidentified\_Archaea*, *Lactobacillus* and *Dechloromonas*,  
308 respectively. The results also revealed that *intI1* only had a significant correlation ( $p < 0.05$ )  
309 with one ARG *sull* in this study (Fig. 6). This demonstrated that *intI1* may not be the main  
310 factor affecting the fate of the tested ARGs.

311

312 (Position for Fig. 6)

313

314 The discrepant changes of the bacterial abundances may explain the different ARGs  
315 abundances in the anaerobically digested sludges with and without FA pretreatment. In our  
316 study, most of the bacterial hosts of the ARGs were affiliated into three phyla, which were  
317 *Proteobacteria*, *Bacteroidetes* and *Euryarchaeota*. *Proteobacteria* was the predominant  
318 bacterial phylum, which included eight bacterial hosts (genus level) in this study. As shown in  
319 Fig. S1, FA pretreatment significantly decreased the absolute abundance of the *Proteobacteria*  
320 from  $2.0 \times 10^{10}$  to  $1.1 \times 10^{10}$  gene copies/g-TS during anaerobic sludge digestion. Fig. 7 further  
321 showed the difference in the abundance of the bacteria (at the genus level) in the anaerobically  
322 digested sludges with and without FA pretreatment. Compared to the digested sludge without  
323 FA pretreatment, FA pretreatment reduced the abundance of most potential bacterial hosts in  
324 the digested sludge. For instance, the abundances of *Thauera* and *Lactobacillus* decreased from

---

325  $1.8 \times 10^8$  and  $5.0 \times 10^8$  gene copies/g-TS in the digested sludge without FA pretreatment to  
326  $1.4 \times 10^8$  and  $7.4 \times 10^6$  gene copies/g-TS in the digested sludge with FA pretreatment,  
327 respectively. Furthermore, compared to anaerobically digested sludge without FA pretreatment,  
328 FA pretreatment reduced the total absolute abundances of the bacteria genera (i.e.,  
329 *unidentified\_Archaea*, *Lactobacillus*, *unidentified\_Nitrosomonadaceae*, *Thauera*, *Halomonas*,  
330 *Acidaminobacter*, *Anaeromyxobacter*, *Aquabacterium*, *Sulfuritalea*), that were associated with  
331 the decreased ARGs in the digested sludge from  $1.8 \times 10^9$  to  $1.0 \times 10^9$  gene copies/g-TS. These  
332 may explain the lower abundance of ARGs in the anaerobically digested sludge with FA  
333 pretreatment. In contrast, the total abundances of potential bacterial hosts of *sulI* and *tetM* were  
334 relatively stable in digested sludges with and without FA pretreatment, which was consistent  
335 with the negligible changes of their abundances in anaerobically digested sludge due to FA  
336 pretreatment. In addition, the abundance of the potential bacterial host of *tetG* (i.e.,  
337 *Dechloromonas*) increased from  $1.7 \times 10^8$  gene copies/g-TS in the digested sludge without FA  
338 pretreatment to  $6.3 \times 10^8$  gene copies/g-TS in the digested sludge with FA pretreatment. This  
339 may contribute to the increased abundance of *tetG* in the digested sludge with FA pretreatment  
340 compared to the digested sludge without FA pretreatment.

341

342 (Position for Fig. 7)

343

#### 344 4. Discussion

345 This study reveals for the first time that FA pretreatment could enhance ARGs removal in  
346 anaerobic sludge digestion. This was experimentally demonstrated through the anaerobic  
347 digestion tests with or without FA pretreatment. Nine widely detected ARGs were tested in the  
348 experiments. The combined use of FA pretreatment and anaerobic digestion reduced the total  
349 absolute abundance of the tested ARGs by ~15% in comparison to anaerobic digestion without

---

350 FA pretreatment. This is caused by the removal of tested ARGs during FA pretreatment and  
351 the reduced abundance of potential bacterial hosts of ARGs due to FA pretreatment during the  
352 anaerobic digestion process. The removal of *IntI1* was not significantly affected by FA  
353 pretreatment, indicating that *intI1* was not the main factor affecting the fate of the tested ARGs  
354 in this study.

355

#### 356 **4.1 Potential mechanisms of enhanced removal of ARGs in combined use of FA** 357 **pretreatment and anaerobic digestion**

358 Our results indicated that some ARGs were removed during FA pretreatment. FA is the  
359 unionized form of ammonium ( $\text{NH}_4^+$ ). Previous studies have shown that FA could inactivate a  
360 variety of microorganisms, as FA could easily diffuse through the cell membrane to reach the  
361 cytoplasm, resulting in proton imbalance or potassium deficiency (Martinelle and Högström,  
362 1993). Therefore, FA could potentially inactivate or kill antibiotic resistant bacteria during the  
363 pretreatment. Additionally, a recent study has indicated that FA could induce DNA damage  
364 (Zhang et al., 2020). As a consequence, the DNA that carries ARGs might be damaged, which  
365 leads to a decrease in the absolute abundances of ARGs including *blaTEM*, *sul1*, *sul2*, *tetA*,  
366 *tetB* and *tetX* in this study.

367

368 The relative abundances of ARGs also decreased after FA pretreatment, which may be because  
369 16S rRNA genes were less affected by FA treatment than ARGs (Fig. 1). This might be related  
370 to different positions of ARGs and 16s rRNA genes in the microorganisms. 16S rRNA genes  
371 are typically located at bacterial genomic DNA. They are tightly coiled and are around by the  
372 nucleotide-related proteins (Olins and Olins, 2003). These proteins may protect the 16S rRNA  
373 gene from being damaged by FA to some extent. In terms of ARGs, many of them are located  
374 at plasmids that are widely distributed in the cytoplasm (Pogliano, 2002). Therefore, ARGs are



---

375 more vulnerable to FA attack than 16S rRNA genes. In terms of *intII*, though some of them  
376 could also be attacked by FA, DNA damage may result in SOS response (a stress-response  
377 system in bacteria) of cell, increase the conjugative transfer frequency and activate *intII*  
378 expression (Erill et al., 2007; Hocquet et al., 2012). As a consequence, FA pretreatment did not  
379 significantly affect the abundance of *intII* in sludge.

380

381 It has been reported that the microbial community is the primary determinant of ARGs  
382 abundance in the anaerobically digested sludge (Guo et al., 2017; Ma et al., 2011). In this study,  
383 FA pretreatment changed the microbial community in the anaerobically digested sludge, which  
384 could be caused by the different selective pressure due to the effect of FA on microbes (Wang.,  
385 2017; Wang et al., 2017). For example, our study found that FA reduced the abundance of  
386 *Thauera* in the anaerobically digested sludge (Fig. 7). As *Thauera* was the potential bacteria  
387 host of *sul2*, *tetA*, *tetB* and *tetX* (Fig. 6), the enhanced removal of these ARGs was clearly  
388 linked with the FA pretreatment. Moreover, the total abundance of the potential bacterial hosts  
389 of the decreased ARGs was lower in the digested sludge with FA pretreatment than that in the  
390 digested sludge without FA pretreatment, contributing to the reduced abundance of the ARGs  
391 (Figs. 6 and 7). As described in Section 3.3, the relative stable total abundances of potential  
392 bacterial hosts may result in negligible changes in the abundances of *sulI* and *tetM* in the  
393 anaerobically digested sludges with and without FA pretreatment. In terms of *tetG*, the  
394 abundance of its potential bacterial host (i.e. *Dechloromonas*) was higher in the digested sludge  
395 without FA pretreatment than that in the digested sludge with FA pretreatment, which could  
396 potentially lead to an increased *tetG* abundance after implementing FA pretreatment (Figs. 6  
397 and 7).

398

---

399 *IntI1* plays an important role in the horizontal gene transfer among bacteria which could also  
400 be a mechanism affecting the fate of ARGs in the anaerobic sludge digestion under certain  
401 conditions (Zhang et al., 2019). For instance, Zhang et al. (2019) concluded that the changes  
402 of *intI1* could be related to the decrease of *sul1*, *tetG*, *ereA* and *sul2* in the anaerobically  
403 digested sludge with microwave pretreatment. In our study, the abundance of *intI1* remained  
404 similar in the anaerobically digested sludges both with and without FA pretreatment and was  
405 only found to be positively correlated with *sul1*. This suggested that some *sul1* genes may be  
406 harbored in *intI1* and *intI1* may not affect the removal of most of ARGs in anaerobic sludge  
407 digestion in this study.

408

#### 409 **4.2 FA pretreatment as a potential strategy for enhancing ARGs removal in the** 410 **anaerobically digested sludge**

411 Our previous study demonstrated that FA pretreatment could enhance methane production  
412 during anaerobic digestion of primary and secondary sludges (Wang, 2017; Wei et al., 2017a,  
413 2017b). This study showed for the first time that, in addition to the enhanced methane  
414 production, FA pretreatment could also enhance the ARGs removal in anaerobic sludge  
415 digestion.

416

417 Sludge from WWTPs has been regarded as an essential resource for agriculture (Sharma et al.,  
418 2017). For example, more than 65% and 50% of the sludge are been reused for agriculture in  
419 Australia and the United States, respectively (Australian and New Zealand Biosolids  
420 Partnership, 2019; the United States Environmental Protection Agency, 2019). Sludge is the  
421 major by-product of the wastewater treatment process. It has a nutrient value that can condition  
422 soils and improve their structure and water retention. These benefits create strong motivations  
423 for the reuse of sludge. However, such an opportunity may disappear if the risk to human health

---

424 due to ARGs spread through sludge outweighs the fertilizing benefits. This study indicates that  
425 the combined use of FA pretreatment and anaerobic digestion could enhance the removal of  
426 ARGs, thereby mitigating the risk of sludge reuse to human health. This provided promising  
427 support for the safe application of reused sludge as an organic fertiliser globally to support the  
428 transition to a circular economy.

429

430 Importantly, FA is a by-product of wastewater treatment and can be attained directly from the  
431 anaerobic digestion liquor of the WWTPs, which contain an FA concentration of 30-560 mg  
432 NH<sub>3</sub>-N/L (Wang et al., 2017). Therefore, the FA pretreatment strategy requires negligible  
433 external chemical input. The FA pretreatment strategy is also sophisticated in its simplicity,  
434 easing its uptake. Its implementation only requires the installation of a small, simple mixing  
435 tank and minor retrofitting of existing WWTPs. Therefore, this FA pretreatment strategy is  
436 potentially economically attractive and environmentally friendly.

437

438 This study tested nine representatives of ARGs that are widely detected in the wastewater  
439 environment. The results indicated that FA pretreatment may potentially be able to enhance the  
440 removal of the other un-tested ARGs in anaerobic sludge digestion. It has been reported that  
441 the potential bacterial hosts of certain ARGs could be commonly shared with a variety of other  
442 ARGs (Tian et al., 2019). For example, *Thauera* was the host of *strB* (a streptomycin resistance  
443 gene), *aadA2* (an aminoglycoside resistance gene) and *qacH* (a fluoroquinolone resistance gene)  
444 (Tian et al., 2019). Therefore, the decrease in the abundance of *Thauera* in the anaerobically  
445 digested sludge with FA pretreatment (Fig. 7) observed in this study may also result in an  
446 enhanced removal of *strB*, *aadA2* and *qacH*.

447

448 It should be highlighted that this is only a proof-of-concept study that demonstrated the

---

449 feasibility of the FA pretreatment strategy in enhancing ARGs removal in anaerobic digestion.  
450 Therefore, the optimization of the FA pretreatment strategy (e.g., FA concentration) was not  
451 conducted in this study. A higher ARGs or *intI1* removal may be achievable after optimizing  
452 the FA pretreatment strategy. Furthermore, though promising results were achieved in our  
453 laboratory studies, full-scale trials are required to fully reveal the potential of this FA  
454 pretreatment strategy. In addition, this study only investigated a selective of the known ARGs  
455 using RT-qPCR to represent various antibiotic classes. Future studies should focus on  
456 conducting metagenomic sequencing technique to reveal the broad spectrum profile of ARGs.

457

## 458 **5. Conclusions**

459 In this study, the effect of FA pretreatment on the removal ARGs in the anaerobic sludge  
460 digestion was evaluated by anaerobic digestion tests. The following conclusions can be drawn  
461 from this study:

462

- 463 • FA pretreatment at 420 mg NH<sub>3</sub>-N/L for 24h enhanced the removal of the total tested  
464 ARGs by ~15% in the anaerobically digested sludge. This revealed that FA pretreatment  
465 could potentially reduce the spread of ARGs from the sludge to the natural environment  
466 through sludge reuse.
- 467 • FA pretreatment had a negligible effect on the abundance of *intI1*.
- 468 • The enhanced removal of ARGs was likely attributed to the removal of tested ARGs  
469 during FA pretreatment and the reduced abundance of potential bacterial hosts of ARGs  
470 due to FA pretreatment during the anaerobic digestion process.

471

## 472 **Acknowledgments**

---

473 The authors acknowledge the Australian Research Council (ARC) Discovery Project  
474 (DP200100933) awarded to Qilin Wang. Qilin Wang acknowledges the ARC Future  
475 Fellowship (FT200100264).

476

477 **References**

478 Allen, H.K., Donato, J., Wang, H.H., Cloud-Hansen, K.A., Davies, J., Handelsman, J., 2010.  
479 Call of the wild: antibiotic resistance genes in natural environments. *Nat. Rev. Microbiol.*  
480 8 (4), 251-259.

481 Amos, G.C., Ploumakis, S., Zhang, L., Hawkey, P.M., Gaze, W.H., Wellington, E.M., 2018.  
482 The widespread dissemination of integrons throughout bacterial communities in a riverine  
483 system. *ISME J.* 12 (3), 681-691.

484 Anthonisen, A.C., Loehr, R.C., Prakasam, T., Srinath, E., 1976. Inhibition of nitrification by  
485 ammonia and nitrous acid. *J. Water Pollut. Control Fed.* 835-852.

486 Auerbach, E.A., Seyfried, E.E., McMahon, K.D., 2007. Tetracycline resistance genes in  
487 activated sludge wastewater treatment plants. *Water Res.* 41 (5), 1143-1151.

488 Australian and New Zealand Biosolids Partnership., 2019. Australian Biosolids Statistics. URL:  
489 [https://www.biosolids.com.au/guidelines/australian-biosolids-](https://www.biosolids.com.au/guidelines/australian-biosolids-statistics/)  
490 [statistics/](https://www.biosolids.com.au/guidelines/australian-biosolids-statistics/).

491 Batstone, D.J., Keller, J., Angelidaki, I., Kalyuzhnyi, S., Pavlostathis, S., Rozzi, A., Sanders,  
492 W., Siegrist, H., Vavilin, V., 2002. The IWA anaerobic digestion model no 1 (ADM1).  
493 *Water Sci. Technol.* 45 (10), 65-73.

494 Bougrier, C., Albasi, C., Delgenès, J.P., Carrère, H., 2006. Effect of ultrasonic, thermal and  
495 ozone pre-treatments on waste activated sludge solubilisation and anaerobic  
496 biodegradability. *Chem. Eng. Process.* 45 (8), 711-718.

497 Brown, M.G., Balkwill, D.L., 2009. Antibiotic resistance in bacteria isolated from the deep

---

498 terrestrial subsurface. *Microb. Ecol.* 57 (3), 484.

499 Burch, T.R., Sadowsky, M.J., LaPara, T.M., 2013. Aerobic digestion reduces the quantity of  
500 antibiotic resistance genes in residual municipal wastewater solids. *Front. Microbiol.* 4, 17.

501 Carrère, H., Dumas, C., Battimelli, A., Batstone, D.J., Delgenes, J.P., Steyer, J.-P., Ferrer, I.,  
502 2010. Pretreatment methods to improve sludge anaerobic degradability: a review.  
503 *J. Hazard. Mater.* 183 (1-3), 1-15.

504 Chen, Q., An, X., Li, H., Su, J., Ma, Y., Zhu, Y.G., 2016. Long-term field application of sewage  
505 sludge increases the abundance of antibiotic resistance genes in soil. *Environ Int.* 92, 1-10.

506 Duan, M., Li, H., Gu, J., Tuo, X., Sun, W., Qian, X., Wang, X., 2017. Effects of biochar on  
507 reducing the abundance of oxytetracycline, antibiotic resistance genes, and human  
508 pathogenic bacteria in soil and lettuce. *Environ. Pollut.* 224, 787-795.

509 Edgar, R.C., 2013. UPARSE: highly accurate OTU sequences from microbial amplicon reads.  
510 *Nat. Methods* 10 (10), 996-998.

511 Edgar, R.C., Haas, B.J., Clemente, J.C., Quince, C., Knight, R., 2011. UCHIME improves  
512 sensitivity and speed of chimera detection. *Bioinformatics* 27 (16), 2194-2200.

513 Erill, I., Campoy, S., Barbé, J., 2007. Aeons of distress: an evolutionary perspective on the  
514 bacterial SOS response. *FEMS Microbiol. Rev.* 31 (6), 637-656.

515 Guo, J., Li, J., Chen, H., Bond, P.L., Yuan, Z., 2017. Metagenomic analysis reveals wastewater  
516 treatment plants as hotspots of antibiotic resistance genes and mobile genetic elements.  
517 *Water Res.* 123, 468-478.

518 Hocquet, D., Llanes, C., Thouverez, M., Kulasekara, H.D., Bertrand, X., Plésiat, P., Mazel, D.,  
519 Miller, S.I., 2012. Evidence for induction of integron-based antibiotic resistance by the  
520 SOS response in a clinical setting. *PLoS Pathog.* 8 (6), e1002778.

521 Karkman, A., Do, T.T., Walsh, F., Virta, M.P., 2018. Antibiotic-resistance genes in waste water.  
522 *Trends Microbiol.* 26 (3), 220-228.

---

523 Li, B., Yang, Y., Ma, L., Ju, F., Guo, F., Tiedje, J.M., Zhang, T., 2015. Metagenomic and  
524 network analysis reveal wide distribution and co-occurrence of environmental antibiotic  
525 resistance genes. *ISME J.* 9 (11), 2490-2502.

526 Ma, Y., Wilson, C.A., Novak, J.T., Riffat, R., Aynur, S., Murthy, S., Pruden, A., 2011. Effect  
527 of various sludge digestion conditions on sulfonamide, macrolide, and tetracycline  
528 resistance genes and class I integrons. *Environ. Sci. Technol.* 45 (18), 7855-7861.

529 Martin, M., 2011. Cutadapt removes adapter sequences from high-throughput sequencing reads.  
530 *EMBnet J.* 17 (1), 10-12.

531 Martinelle, K., Häggström, L., 1993. Mechanisms of ammonia and ammonium ion toxicity in  
532 animal cells: transport across cell membranes. *J. Biotechnol.* 30 (3), 339-350.

533 Negreanu, Y., Pasternak, Z., Jurkevitch, E., Cytryn, E., 2012. Impact of treated wastewater  
534 irrigation on antibiotic resistance in agricultural soils. *Environ. Sci. Technol.* 46 (9), 4800-  
535 4808.

536 Nölvak, H., Truu, M., Kanger, K., Tampere, M., Espenberg, M., Loit, E., Raave, H., Truu, J.,  
537 2016. Inorganic and organic fertilizers impact the abundance and proportion of antibiotic  
538 resistance and integron-integrase genes in agricultural grassland soil. *Sci. Total Environ.*  
539 562, 678-689.

540 Olins, D.E., Olins, A.L., 2003. Chromatin history: our view from the bridge. *Nat. Rev. Mol.* 4  
541 (10), 809-814.

542 Osińska, A., Korzeniewska, E., Harnisz, M., Felis, E., Bajkacz, S., Jachimowicz, P., Niestępski,  
543 S., Konopka, I., 2020. Small-scale wastewater treatment plants as a source of the  
544 dissemination of antibiotic resistance genes in the aquatic environment. *J. Hazard. Mater.*  
545 381, 121221.

546 Ouyang, W., Gao, B., Cheng, H., Zhang, L., Wang, Y., Lin, C., Chen, J., 2020. Airborne  
547 bacterial communities and antibiotic resistance gene dynamics in PM<sub>2.5</sub> during rainfall.

---

548 Environ. Int. 134, 105318.

549 Pei, J., Yao, H., Wang, H., Ren, J., Yu, X., 2016. Comparison of ozone and thermal hydrolysis  
550 combined with anaerobic digestion for municipal and pharmaceutical waste sludge with  
551 tetracycline resistance genes. *Water Res.* 99, 122-128.

552 Pogliano, J., 2002. , Dynamic cellular location of bacterial plasmids. *Curr. Opin. Microbiol.* 5  
553 (6), 586-590.

554 Quast, C., Pruesse, E., Yilmaz, P., Gerken, J., Schweer, T., Yarza, P., Peplies, J., Glöckner,  
555 F.O., 2012. The SILVA ribosomal RNA gene database project: improved data processing  
556 and web-based tools. *Nucleic Acids Res.* 41 (D1), D590-D596.

557 Ross, J., Topp, E., 2015. Abundance of antibiotic resistance genes in bacteriophage following  
558 soil fertilization with dairy manure or municipal biosolids, and evidence for potential  
559 transduction. *Appl. Environ. Microbiol.* 81 (22), 7905-7913.

560 Shao, S., Hu, Y., Cheng, J., Chen, Y., 2018. Research progress on distribution, migration,  
561 transformation of antibiotics and antibiotic resistance genes (ARGs) in aquatic  
562 environment. *Crit. Rev. Biotechnol.* 38 (8), 1195-1208.

563 Sharma, B., Sarkar, A., Singh, P. and Singh, R.P., 2017. Agricultural utilization of biosolids:  
564 A review on potential effects on soil and plant grown. *Waste Manage.* 64, 117-132.

565 Singh, R., Agrawal, M., 2008. Potential benefits and risks of land application of sewage sludge.  
566 *Waste Manage.* 28 (2), 347-358.

567 Song, W., Wang, X., Gu, J., Zhang, S., Yin, Y., Li, Y., Qian, X., Sun, W., 2017. Effects of  
568 different swine manure to wheat straw ratios on antibiotic resistance genes and the  
569 microbial community structure during anaerobic digestion. *Bioresour. Technol.* 231, 1-8.

570 Teuber, M., 2001. Veterinary use and antibiotic resistance. *Curr. Opin. Microbiol.* 4 (5), 493-  
571 499.

572 The United States Environmental Protection Agency, 2019. Basic information about biosolids,



---

573 URL: <https://www.epa.gov/biosolids/basic-information-about-biosolids#basics>.

574 Wang, M., Li, R., Zhao, Q., 2019. Distribution and removal of antibiotic resistance genes  
575 during anaerobic sludge digestion with alkaline, thermal hydrolysis and ultrasonic  
576 pretreatments. *Front. Environ. Sci. Eng.* 13 (3).

577 Wang, Q., 2017. A roadmap for achieving energy-positive sewage treatment based on sludge  
578 treatment using free ammonia. *ACS Sustain. Chem. Eng.* 5 (11), 9630-9633.

579 Wang, Q., Duan, H., Wei, W., Ni, B.-J., Laloo, A., Yuan, Z., 2017. Achieving stable  
580 mainstream nitrogen removal via the nitrite pathway by sludge treatment using free  
581 ammonia. *Environ. Sci. Technol.* 51 (17), 9800-9807.

582 Wei, W., Zhou, X., Wang, D., Sun, J., Wang, Q., 2017a. Free ammonia pre-treatment of  
583 secondary sludge significantly increases anaerobic methane production. *Water Res.* 118,  
584 12-19.

585 Wei, W., Zhou, X., Xie, G.J., Duan, H., Wang, Q., 2017b. A novel free ammonia based  
586 pretreatment technology to enhance anaerobic methane production from primary sludge.  
587 *Biotechnol. Bioeng.* 114 (10), 2245-2252.

588 Xue, G., Jiang, M., Chen, H., Sun, M., Liu, Y., Li, X., Gao, P., 2019. Critical review of ARGs  
589 reduction behavior in various sludge and sewage treatment processes in wastewater  
590 treatment plants. *Crit. Rev. Environ. Sci. Technol.* 49 (18), 1623-1674.

591 Yang, Y., Li, B., Zou, S., Fang, H.H., Zhang, T., 2014. Fate of antibiotic resistance genes in  
592 sewage treatment plant revealed by metagenomic approach. *Water Res.* 62, 97-106.

593 Zhang, J., Liu, J., Lu, T., Shen, P., Zhong, H., Tong, J., Wei, Y., 2019. Fate of antibiotic  
594 resistance genes during anaerobic digestion of sewage sludge: Role of solids retention  
595 times in different configurations. *Bioresour. Technol.* 274, 488-495.

596 Zhang, T., Yan, Z., Zheng, X., Wang, S., Fan, J., Liu, Z., 2020. Effects of acute ammonia  
597 toxicity on oxidative stress, DNA damage and apoptosis in digestive gland and gill of

---

598 Asian clam (*Corbicula fluminea*). *Fish Shellfish Immunol.* 99, 514-525.

599 Zhang, X.-X., Zhang, T., 2011. Occurrence, Abundance, and Diversity of Tetracycline  
600 Resistance Genes in 15 Sewage Treatment Plants across China and Other Global Locations.  
601 *Environ. Sci. Technol.* 45 (7), 2598-2604.

602 Zhuang, Y., Ren, H., Geng, J., Zhang, Y., Zhang, Y., Ding, L., Xu, K., 2015. Inactivation of  
603 antibiotic resistance genes in municipal wastewater by chlorination, ultraviolet, and  
604 ozonation disinfection. *Environ. Sci. Pollut. Res.* 22 (9), 7037-7044.

605

## 606 **List of figures**

**Figure 1.** Absolute abundances of ARGs and *intI1* in different sludges

**Figure 2.** Removal ratios (%) of ARGs and *intI1* in sludge during FA pretreatment. Negative means the target genes increased after FA pretreatment.

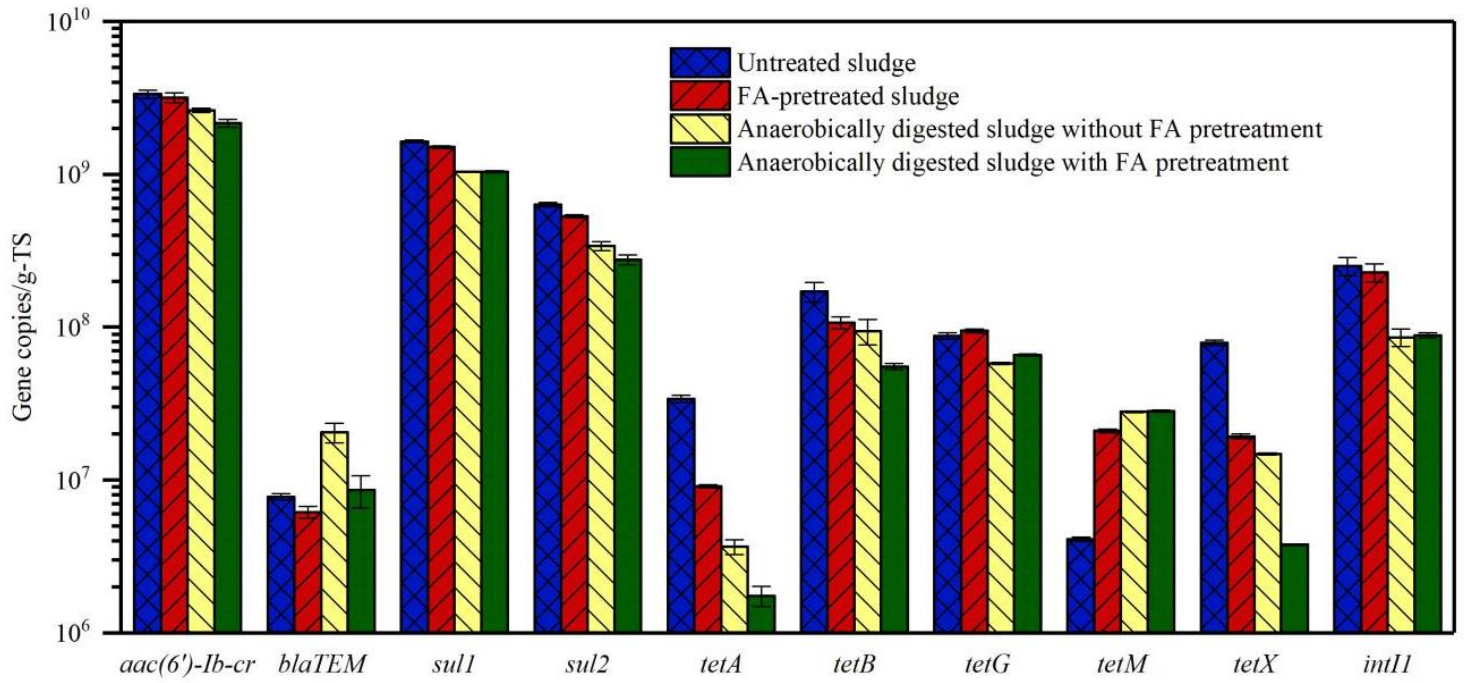
**Figure 3.** Relative abundances of ARGs and *intI1* in different sludges

**Figure 4.** Removal ratios (%) of ARGs and *intI1* in anaerobically digested sludge with and without FA pretreatment compared to untreated sludge. Negative means the target genes increased.

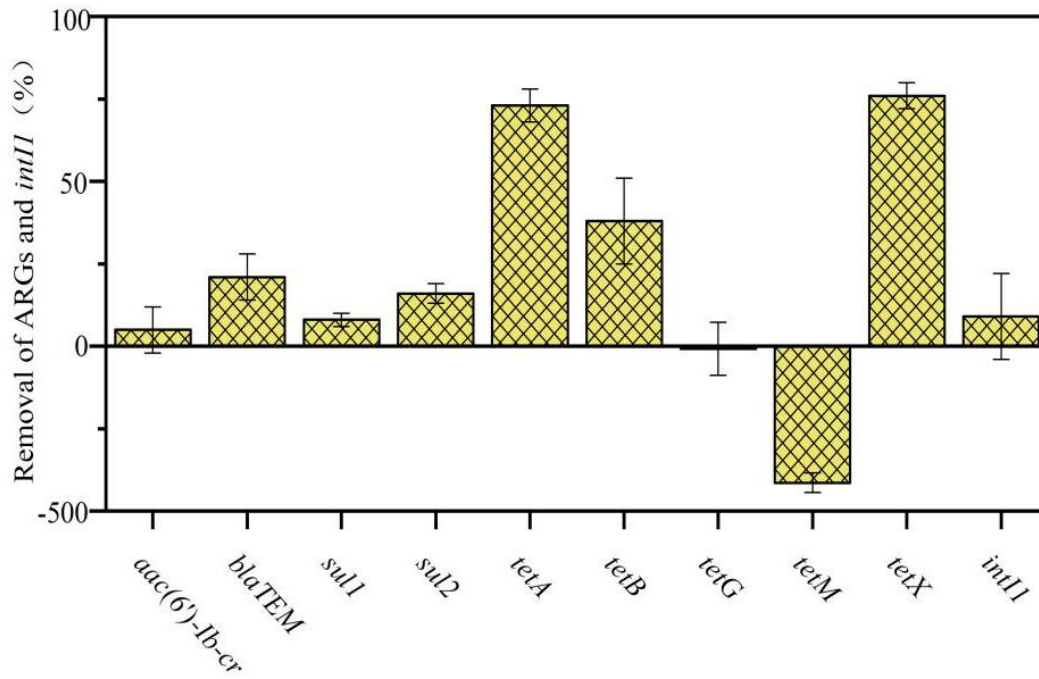
**Figure 5.** Enhanced removal (%) of ARGs and *intI1* in the anaerobically digested sludge with FA pretreatment compared with the anaerobically digested sludge without FA pretreatment. Negative means the target genes increased in the anaerobically digested sludge with FA pretreatment.

**Figure 6.** Correlation between ARGs and *intI1*/microbial community at the genus level. An asterisk (\*) indicates a significant positive correlation ( $R > 0.8$ ,  $P < 0.05$ ). The scale bar showed the R value between ARGs and *intI1*/microbial community.

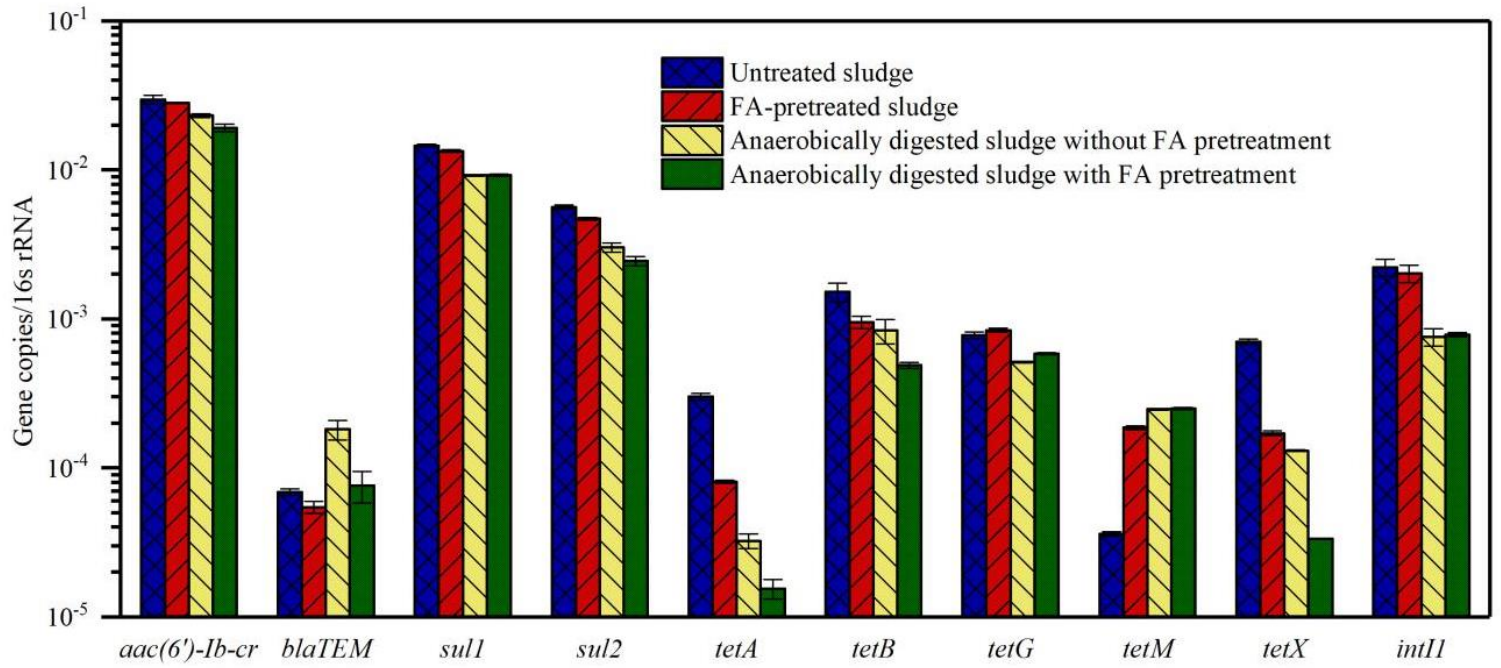
**Figure 7.** Heatmap of the top 30 most abundant bacteria (at the genus level) in the anaerobically digested sludge with and without FA pretreatment



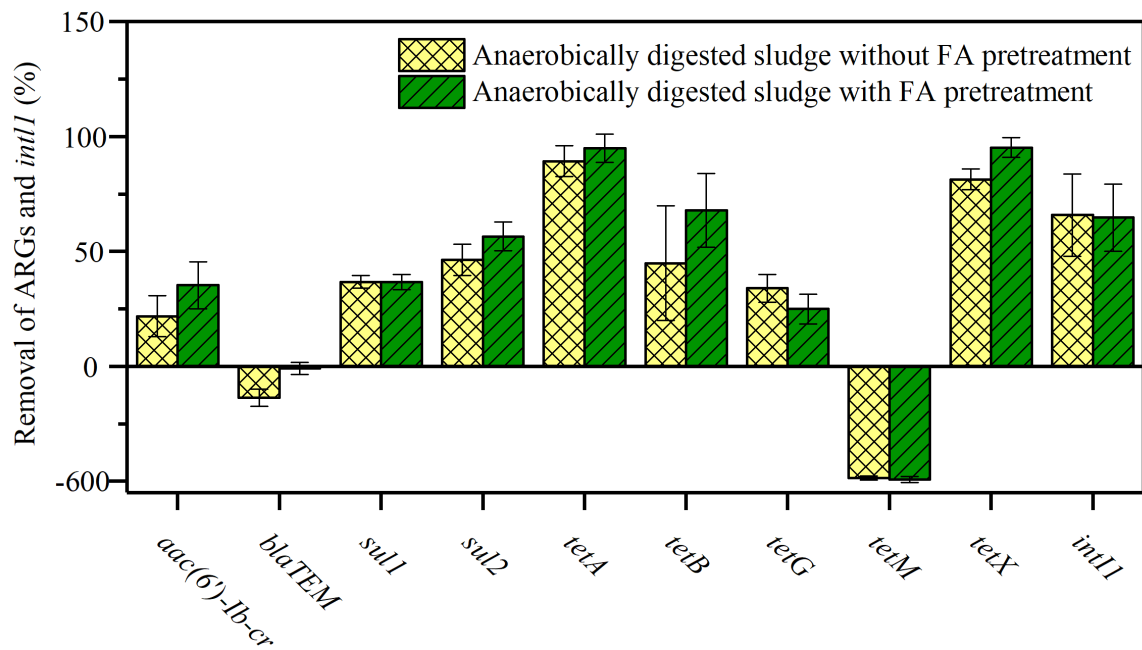
**Figure 1.** Absolute abundances of ARGs and *intI1* in different sludges



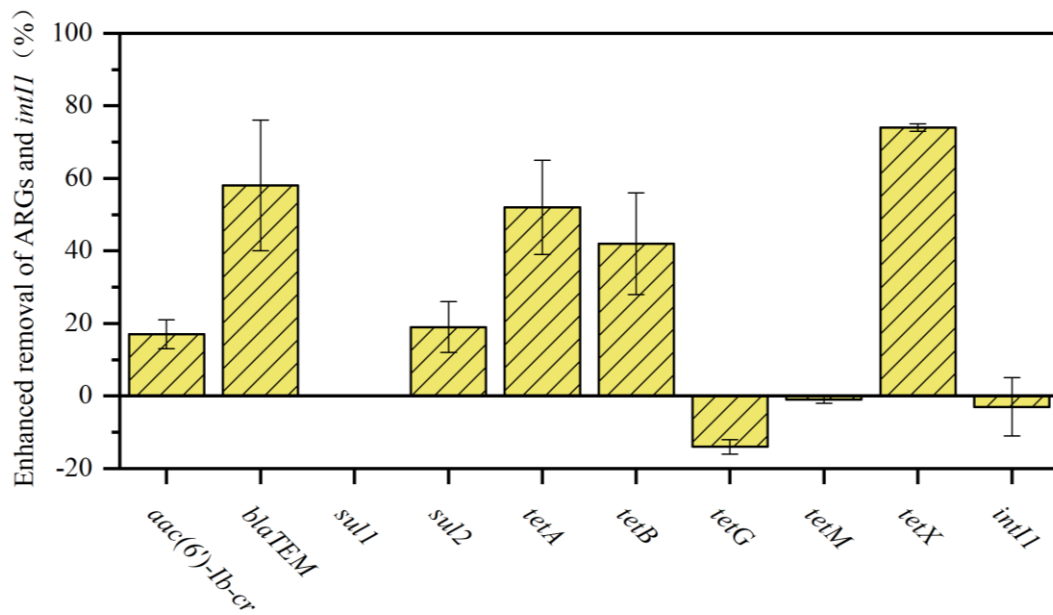
**Figure 2.** Removal ratios (%) of ARGs and *intI1* in sludge during FA pretreatment. Negative means the target genes increased after FA pretreatment.



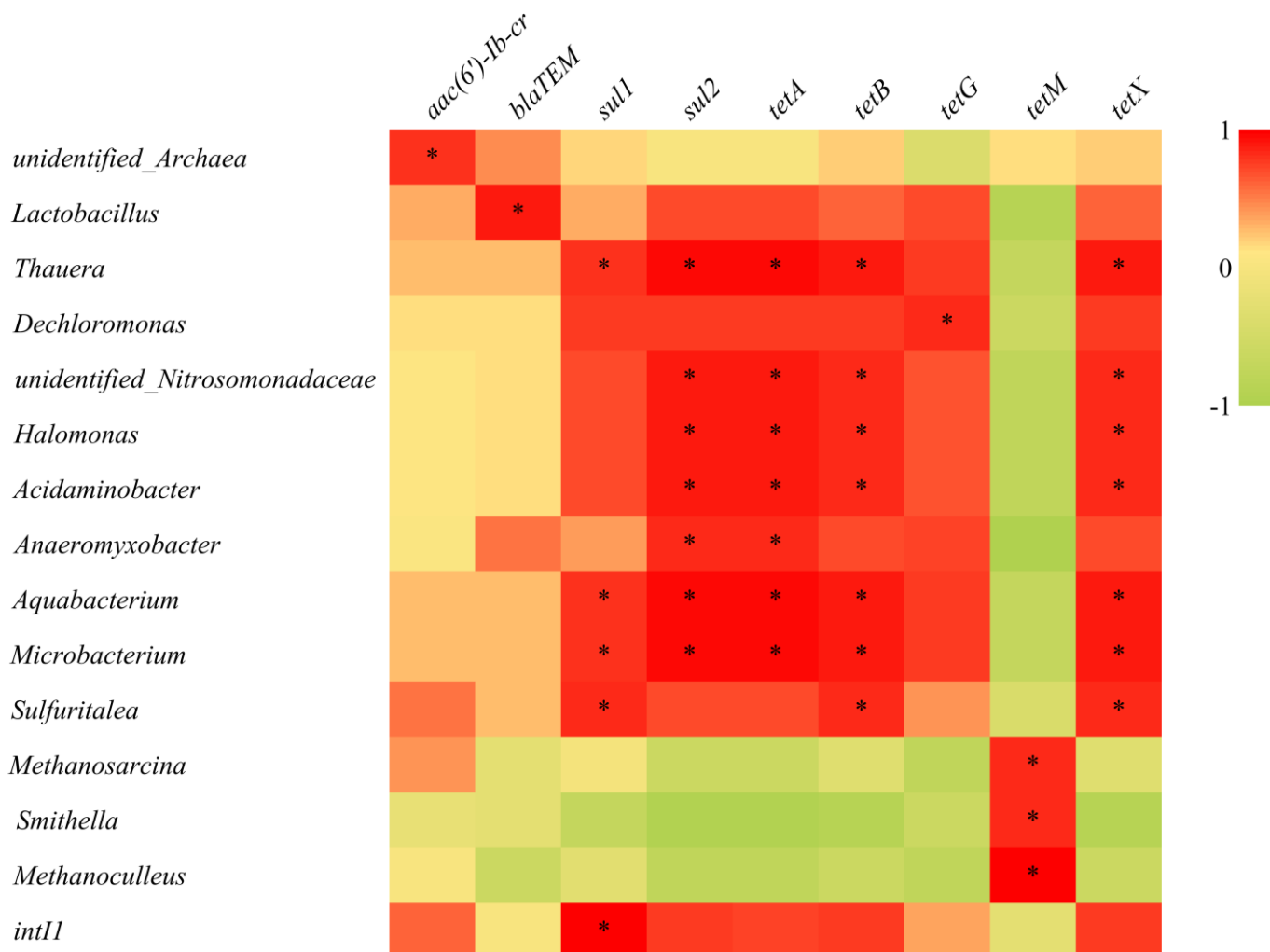
**Figure 3.** Relative abundances of ARGs and *intI1* in different sludges



**Figure 4.** Removal ratios (%) of ARGs and *intI1* in anaerobically digested sludge with and without FA pretreatment compared to untreated sludge. Negative means the target genes increased.

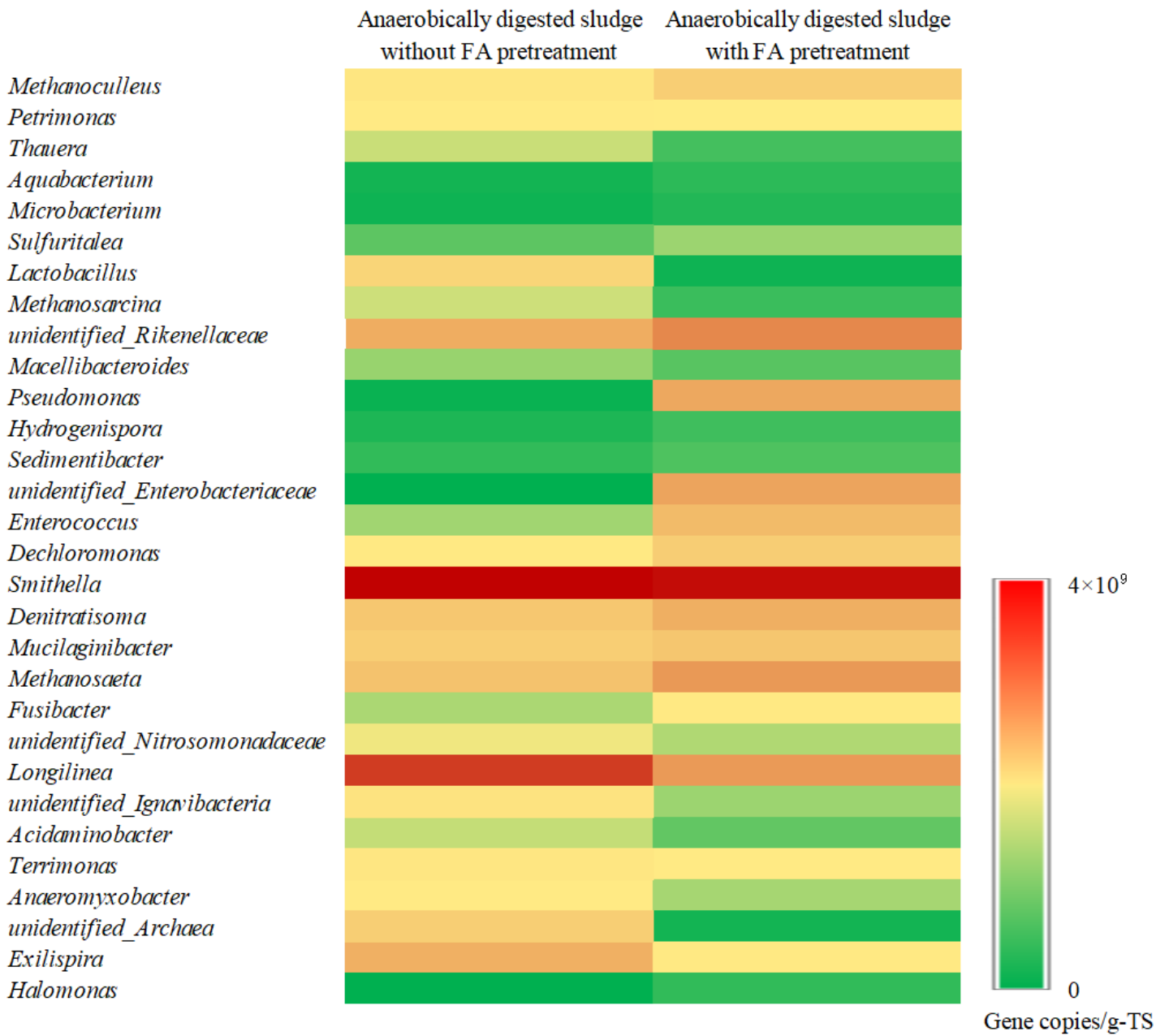


**Figure 5.** Enhanced removal (%) of ARGs and *intl1* in the anaerobically digested sludge with FA pretreatment compared with the anaerobically digested sludge without FA pretreatment. Negative means the target genes increased in the anaerobically digested sludge with FA pretreatment.



**Figure 6.** Correlation between ARGs and *intI1*/microbial community at the genus level. An asterisk (\*) indicates a significant positive correlation ( $R > 0.8$ ,  $P < 0.05$ ). The scale bar showed the R value between ARGs and *intI1*/microbial community.





**Figure 7.** Heatmap of the top 30 most abundant bacteria (at the genus level) in the anaerobically digested sludge with and without FA pretreatment.