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1	Effect of diesel particulate filter regeneration on fuel consumption				
2	and emissions performance under real-driving conditions				
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Abstract

22 Diesel particulate filters (DPF) are widely adopted in diesel vehicles to meet the increasingly 23 stringent emission regulations, which require continuous passive regenerations or/and periodic 24 active regenerations to burn off the accumulated particulate matter (PM). In spite of many laboratory 25 studies using DPF benches and engine/chassis dynamometers, there is currently a lack of 26 investigation on DPF regeneration under real-world conditions. Therefore, this study was conducted 27 to investigate the impact of active DPF regenerations on the fuel consumption and gaseous and 28 particulate emissions performance of a diesel light goods vehicle under real-driving conditions by 29 using the state-of-the-art portable emission measurement system. In total, 60 real-driving emission 30 (RDE) tests (~1200 km in total) were performed on the same route during the same periods of a day, 31 to minimise the effect of uncontrollable real-world factors on the performance evaluation. The 32 results showed that real-world active DPF regenerations occurred every 130 km for the studied 33 vehicle. Although they did not occur frequently, DPF regenerations increased the trip-averaged fuel 34 consumption rate by 13% on average. CO and THC emission factors tended to increase with DPF 35 regenerations because the post combustion used to achieve the high exhaust temperature for 36 regeneration of the filter occurred under oxygen-lean conditions. Total NO_x emissions were not 37 affected but NO₂/NO_x ratio was greatly reduced by DPF regeneration due to lower NO oxidation by 38 the diesel oxidation catalyst and higher NO₂ reduction by the DPF. Finally, DPF regenerations 39 sharply increased PM emission factors by 27 times compared with a trip without DPF regeneration, 40 resulting in significant exceedance of the emission limit.

Keywords: Diesel particulate filter; Regeneration; Fuel consumption; Gaseous and particulate
 emissions; Portable emission measurement system

1. Introduction

44 Diesel vehicles are widely used for commercial road transport worldwide due to their high 45 thermal efficiency and durability. However, a major challenge of diesel engines is that their diffusion 46 combustion mode is prone to produce high particulate matter (PM) and nitrogen oxides (NO_x) 47 emissions [1, 2], and the latter are also key precursors for the formation of secondary air pollutants 48 of PM and ozone (O₃) [3, 4]. Despite representing a small proportion (<5%) of the total fleet, heavy-49 duty diesel vehicles account for 40-60% of global on-road NO_x and PM emissions [5-8], as well as 50 a significant share (>40%) of global on-road energy consumption [9]. These pollutants have caused 51 significant health and economic damages. It was estimated that 385,000 premature deaths and US\$1 52 trillion health losses were associated with global transport emissions (including on-road vehicles, 53 non-road mobile sources, and international shipping) in 2015 [10]. Among various emission sources, 54 on-road diesel vehicles were the biggest contributor to these health burdens, which alone were responsible for approximately 110,000 premature deaths in 2015 [5, 11]. 55

56 Increasingly stringent regulations have been enforced to limit the pollutant emissions from 57 diesel vehicles. For example, the PM emission limits for light-duty diesel vehicles were significantly 58 reduced from 0.025 g/km in Euro 4 to 0.0045 g/km in Euro 5 and 6 standards [12, 13]. Although the 59 limits remained the same after Euro 5, the emission testing methods became stricter in Euro 6 which 60 introduced a more dynamic Worldwide harmonized Light vehicles Test Procedure (WLTP) to 61 replace the outdated New European Driving Cycle (NEDC) for laboratory testing using a chassis 62 dynamometer [14, 15] and a Real Driving Emissions (RDE) test procedure for real-world testing 63 using a portable emission measurement system (PEMS) [16, 17]. The new test procedures aimed to 64 reduce the discrepancies between laboratory and real-world fuel consumption and pollutant 65 emissions [18, 19], which received great research attention after the discovery of the Volkswagen 66 dieselgate in 2015 [20, 21].

Various technologies including in-cylinder combustion optimization and exhaust after treatment have been adopted to meet the stringent emission regulations. For PM emission control,

69 the most effective method involves using a diesel particulate filter (DPF) which has become a 70 mandatory emission control device after Euro 5 [9, 22, 23]. A DPF traps PM when exhaust gas flows 71 through the porous substrate walls by two filtration mechanisms, i.e. deep bed and surface filtration 72 [24]. The deep bed filtration dominates if the pore size of filter media is larger than that of PM. 73 However, surface filtration becomes important as the filter pores are filled. The materials used for 74 manufacturing DPFs include cordierite, silicon carbide, acicular mullite, aluminum titanate, metal 75 foams and fibers, with the first two materials being the most widely used [24-27]. Modern DPFs are 76 highly efficient in reducing PM emissions, which could achieve a high filtration efficiency of over 77 99% [22, 28]. However, back pressure builds up across DPF as PM accumulates, which may 78 deteriorate engine combustion and consequently worsen fuel consumption and emissions 79 performance. Therefore, continuous or periodic DPF regenerations are needed to remove the 80 accumulated PM in order to restore its filtration capacity.

81 DPF regenerations oxidize PM to carbon dioxide (CO₂) using either oxygen (O₂) or nitrogen 82 dioxide (NO₂). The former is abundant in diesel exhaust gas due to the lean combustion mode, while 83 the latter is produced by diesel oxidation catalyst (DOC) via the oxidation of nitric oxide (NO) which 84 is the dominant species in raw diesel NO_x emissions [29]. Since NO₂ is a much stronger oxidizer 85 than O₂, NO₂-based regenerations can occur at a lower temperature (260-300 °C) than O₂-based regenerations (~600 °C) [22]. NO₂-based regenerations (usually referred as passive regenerations) 86 87 occur continuously during normal engine operation. The main advantages of passive regenerations 88 are no fuel consumption penalty and low thermal stress on DPF, while the disadvantage is that 89 regeneration efficiency is limited by a number of factors including NO_x/PM ratio, engine conditions 90 (i.e. exhaust temperature) and fuel sulfur content (i.e. catalyst poisoning) [30, 31]. O₂-based 91 regenerations (usually referred as active regenerations) periodically realize a high exhaust 92 temperature via extra efforts such as in-cylinder post-injection, in-exhaust fuel injection before the 93 DOC and electric heating. Active regenerations are more independent of engine conditions and more 94 effective than passive regenerations, while the main drawbacks are increased fuel consumption and higher thermal stress. Pure passive regenerations may be insufficient to remove the accumulated PM
under all engine conditions, and thus modern diesel vehicles mostly adopt active or hybrid
regenerations [22].

98 Effective and reliable DPF regenerations are of great importance for maintaining the desired 99 fuel consumption and emissions performance of modern diesel vehicles. Many studies have been 100 carried out to investigate the performance of DPF regenerations. Rossomando et al. [32] measured 101 the particle size distributions of a Euro 5 light-duty diesel engine during both DPF accumulation and 102 regeneration phases using an engine dynamometer. The results showed that the regeneration phase 103 increased the particle number concentration by up to two orders of magnitude than the accumulation 104 phase. Meng et al. [33] explored the formation mechanisms and influencing factors of PM emissions 105 during DPF regenerations using both DPF and engine test benches. They found that 10 nm 106 nucleation mode particles were emitted by high temperature inlet flows during the DPF heating-up 107 stage, leading to 2-3 orders of magnitude increase in particle number concentration. These particles 108 were mainly from the penetration of exhaust particles, as well as the blow-out of the deposited soot 109 layer. Using a longer DPF inlet transition section could improve the particle deposition and 110 regeneration performance [34]. Beatrice et al. [35] measured the PM characteristics of a Euro 5 111 diesel engine under real regeneration strategies using an engine dynamometer. The results showed that a large number of small particles were produced by DPF regenerations controlled by multiple 112 113 in-cylinder post-injections. Engine dynamometer experiments on a diesel-methanol dual fuel engine 114 [36] showed that the dual fuel mode had better active and passive regeneration performance than pure diesel mode, in terms of regeneration rates, fuel penalty, and PM and NO₂ emissions. This was 115 116 mostly likely due to the oxygen in the methanol. Yoon et al. [37] measured the PM emissions of 117 parked active DPF regenerations from two heavy-duty diesel trucks using an ambient air dilution 118 tunnel. They found that PM number was highly dominated by nucleation mode particles (i.e. < 50 119 nm) for both trucks, while accumulation mode particles (i.e. 50-500 nm) contributed significant percentages (varied from 7%-28% to 78%-96%) of PM mass. Pechout et al. [38] compared the 120

121 effects of traditional biodiesel and hydro-treated vegetable oil (HVO) on a Euro 6 diesel car using a 122 chassis dynamometer. It was reported that, during DPF regeneration, HVO had comparable/lower 123 particle loading while traditional biodiesel had considerably lower particle loading than diesel fuel. 124 Fuel oxygen content has consistently been found to reduce particle mass. Smith et al. [39] analyzed 125 the real-time PM emissions of active and passive DPF regenerations of two heavy-duty diesel 126 vehicles under steady-state driving conditions using a chassis dynamometer. The results showed that 127 regeneration PM emissions were dominated by particles < 100 nm, and passive regenerations had 128 lower PM emissions than active regenerations. All these experiments have well revealed that fuel 129 consumption and emissions could be significantly increased during active DPF regenerations. To 130 address this issue, various solutions have been studied, such as more accurate air-fraction and soot 131 loading modelling [40, 41], microwave [42], hydrocarbon injection [43], and non-thermal plasma 132 technology [44].

133 The above studies all adopted laboratory testing methods such as DPF test benches, parked 134 testing and engine/chassis dynamometers. Although laboratory experiments are highly accurate and 135 repeatable, they may not fully replicate the real-world conditions which are far more varied and thus 136 could produce very different fuel consumption and emissions performance than laboratory results. So far, very few studies have investigated DPF regenerations under real driving conditions. Ruehl 137 138 et al. [45] investigated the emission performance of two heavy-duty diesel vehicles during parked 139 active, driving active and passive regenerations using PEMS. They found that 0.2-16.3 g of PM was 140 emitted after each regeneration, under the frequency of one real-world active regeneration every 141 28.0 h or 599 miles. Ko et al. [46] evaluated the on-road NO_x emissions from a 2.2 L diesel vehicle 142 using NO_x sensors and an exhaust flow meter (not PEMS). They found that NO_x emissions were 30% higher under DPF regeneration than normal conditions. Papadopoulos et al. [47] measured the 143 144 emission factors of 14 Euro IV-VI diesel medium goods trucks under real-world conditions using 145 PEMS. It was found that DPF regenerations could significantly increase total hydrocarbons (THC) 146 and PM emissions by eight and 34 times, respectively than normal vehicle operation.

147 As reviewed above, there is a lack of investigation on DPF regeneration performance under real 148 driving conditions. Although DPF regenerations do not occur frequently or last long, they may 149 significantly affect the vehicle performance for a given trip. Therefore, this study was carried out to 150 comprehensively evaluate the effects of DPF regeneration on fuel consumption and gaseous and 151 particulate emissions under RDE conditions using a state-of-the-art PEMS. A total of 60 RDE trips, 152 which incurred nine active DPF regeneration trips, were performed using the same diesel vehicle on 153 the same route during the same periods of a day, which minimised the effect of uncontrollable real-154 world factors on performance evaluation of the test vehicle. The results reveal the impact of a DPF 155 regeneration event on a typical RDE trip and provide valuable information for developing more 156 effective DPF technologies.

157

2.

Experimental methods

158 2.1. Test vehicle and emission sensors

Fig. S1 shows the test vehicle installed with a set of on-board emission test equipment. The experiments were performed on a Toyota HiAce diesel light goods vehicle (LGV), which is the most common diesel vehicle model in Hong Kong [48]. The test vehicle was powered by a 2.98 L turbocharged diesel engine and a four-speed automatic transmission. The model year was 2014 and it was certified to Euro 5 standard which was achieved by the widely adopted exhaust after-treatment systems of diesel oxidation catalyst (DOC), a DPF that actively regenerated and exhaust gas recirculation (EGR).

The real-driving fuel consumption and emissions performance were measured by a state-of-theart PEMS which consisted of an AVL M.O.V.E Gas PEMS 493, an AVL M.O.V.E PM PEMS 494, a 2.5-inch EFM-2 flow meter, a weather station, a Peiseler MT wheel speed sensor and a Garmin global positioning system (GPS). The Gas PEMS unit measured CO and CO₂ concentrations by a non-dispersive infra-red (NDIR) analyzer, NO and NO₂ concentrations by a non-dispersive ultraviolet (NDUV) analyzer, and THC concentrations by a heated flame ionization detector (FID). The 172 PM PEMS unit measured time-resolved PM mass by a photo-acoustic measurement unit and a 173 gravimetric filter module. The EFM-2 flow meter was installed in the tailpipe exit to measure the 174 flow rate and temperature of the exhaust gas. The weather station and GPS were installed on the 175 vehicle roof to measure the ambient (i.e. temperature, relative humidity and atmospheric pressure) 176 and route (i.e. elevation and ground speed) conditions, respectively. In addition to GPS, the wheel 177 speed sensor also measured the vehicle driving speed to provide a quality control check on the 178 vehicle speed data provided by the GPS and on-board diagnostics (OBD). The PEMS system was 179 mounted in the cabin of the test vehicle and was supplied by a Honda EU 30IS generator and three 180 lead acid batteries, so that no additional load was put on the vehicle engine. In this study, all the 181 above measurements were taken at a sampling frequency of 10 Hz.

182 2.2. Experimental conditions and procedures

183 Previous experiments were usually conducted using simulated PM loading and regenerations. 184 In this study, real-world DPF regenerations were studied. Since DPF regenerations do not occur 185 frequently and depend on a number of factors such as engine load and exhaust temperature, 186 significant RDE tests were carried out: i.e. 60 RDE trips which were completed by 30 drivers (two 187 trips per driver) and accounted for ~1200 km in total. To minimise the impact of uncontrollable real-188 world conditions on performance evaluation (e.g. traffic and road conditions), all drivers used the 189 same test vehicle to drive on the same route during the same periods of a day (i.e. 11.00-12.00 and 190 14.00-15.00). In addition, all drivers did not receive any training or use any driver assistance device 191 during the tests, so that they followed their normal driving style. The RDE test route (Fig. S2) was 192 representative of daily driving in Hong Kong. It was a round trip between Tsing Yi and Sham Tseng 193 in New Territories of Hong Kong. The total distance was 19 km, which consisted of 5, 6 and 8 km 194 of urban, rural and highway driving, respectively. It took approximately 25-30 minutes to complete 195 one trip. All RDE tests were carried out under hot start conditions when the engine coolant 196 temperature was over 80 °C in the OBD system. To assure data quality, the PEMS was zeroed using 197 pure nitrogen gas before each RDE test and was calibrated using standard EPA protocol standard 198 gas before each test day. In addition, the system was purged before each test day and a leak check199 was performed before each test according to the RDE test procedure.

200

201 2.3. Data treatment and identification of DPF regenerations

In this study, both the instantaneous emission concentrations (% or ppm) and trip-averaged distance specific emission factors (EFs, g/km) were analysed to evaluate the effect of DPF regeneration on RDE performance. The instantaneous emission concentrations were the raw RDE data measured by PEMS at a frequency of 10 Hz. The distance-specific EFs were calculated from the instantaneous emission concentrations, exhaust flow rates and driving speeds for each RDE trip according to the Euro 6 RDE standard [49]. The fuel consumption rates (L/100 km) were calculated from the EFs of CO_2 , CO and THC by applying the principle of carbon balance [50, 51].



- Fig. 1. Instantaneous driving speed, exhaust temperature and CO₂ concentration of one pair of RDE
- tests completed by Driver 1: **a**) morning test without DPF regeneration, **b**) afternoon test with DPF

regeneration. The regeneration window is identified in shadow band.

214 Active DPF regenerations can only occur under relatively high exhaust temperatures that are 215 usually realized by extra fuel combustion such as post in-cylinder or in-exhaust injections. Thus, 216 exhaust temperature and CO₂ concentration are useful indicators of active DPF regenerations. Fig. 217 1 shows the variations of instantaneous driving speed, exhaust temperature and CO₂ concentration 218 of one pair of RDE tests that were completed by Driver 1. The morning test (Fig. 1a) was a normal 219 trip with no DPF regeneration occurring. Although the driving conditions (i.e. speed and acceleration) 220 are highly variable during a RDE trip, the exhaust temperature and CO₂ concentration remained in 221 the ranges of 150-300 °C and 3-12%, respectively during the whole test. Then, a DPF regeneration 222 event occurred in the afternoon test (Fig. 1b). Comparing with a normal trip (Fig. 1a), the exhaust 223 temperature (350-500 °C) and CO₂ concentration (6%-15%) increased significantly when DPF 224 regeneration occurred during 1000-1500 s (Fig. 1b), although the driving conditions were similar to 225 those in the morning test.

226

 Table 1. Characteristics of identified real-driving DPF regenerations.

Driver No.	RDE test session	DPF regeneration duration (length), s	Exhaust temperature range (average), °C	CO ₂ concentration range (average), %
1	Afternoon	953-1450* (497)	330-494 (412)	4.9-14.5 (10.0)
4	Afternoon	232-963 (731)	318-522 (426)	1.6-14.1 (9.5)
6	Morning	901-1526* (625)	329-495 (421)	3.9-14.0 (9.0)
9	Morning	1434-1544* (110)	333-407 (367)	6.6-14.0 (10.1)
12	Afternoon	913-1518* (605)	342-507 (413)	3.3-14.3 (9.3)
16	Morning	703-1261* (558)	332-508 (427)	3.7-13.5 (9.4)
22	Afternoon	198-776 (578)	319-498 (404)	1.5-14.1 (9.4)
26	Morning	788-1429 (641)	338-462 (396)	4.6-13.7 (9.5)
29	Afternoon	597-1132 (535)	334-487 (420)	3.1-13.9 (9.5)

Notes: * indicates that DPF regeneration was still ongoing when the RDE trip was finished (i.e. incomplete DPF regeneration).

227 Therefore, this study defined that an active DPF regeneration occurred if exhaust temperature 228 continuously exceeded 350 °C and CO₂ concentration frequently exceeded 13%. Compared with 229 CO₂ concentration, exhaust temperature was a clearer and more reliable indicator and thus was used 230 as the primary parameter for determining the occurrence of DPF regenerations. In addition, the 231 colour of the PEMS PM filter changed from white to grey after a RDE test if a DPF regeneration 232 occurred. Using the above criteria, nine out of 60 RDE trips were found to have DPF regenerations 233 in this study. On average, real-world active DPF regenerations occurred every 130 km for the studied 234 diesel LGV. Table 1 shows the characteristics of the identified DPF regenerations. Their 235 instantaneous exhaust temperature profiles and the identification of DPF regenerations are presented 236 in Fig. S3. Among the nine DPF regenerations, four occurred in the middle of RDE tests and were 237 completed before the trips were finished. The other five occurred at the end of RDE tests and were 238 still ongoing when the trips finished, and thus were considered as incomplete DPF regenerations. 239 Nevertheless, four of these five incomplete regenerations lasted for over eight minutes and were 240 almost completed (Fig. S3). Specially, the DPF regeneration of Driver 9 occurred at the very end of 241 that RDE test, resulting in a very short regeneration time (~2 minutes) and a much lower average 242 exhaust temperature (367 °C) than others (396-427 °C). As shown in Table 1, a complete real-243 driving DPF regeneration on average took around 10 minutes, with an average exhaust temperature 244 and CO₂ concentration of 412 °C and 9.5%, respectively. The DPF regeneration time of the studied 245 diesel LGV was comparable to that of medium diesel trucks (550 s) [47]. It should be noted that the 246 exhaust temperature measured by the PEMS was at the tailpipe exit and was thus lower than that in 247 the DPF [39].

3. **Results and discussion**

250 3.1. Effect of DPF regenerations on fuel consumption

251 Fig. 2 shows the real-driving fuel consumption performance of trips with and without DPF 252 regenerations. As shown in Fig. 2a, although all the tests were performed using the same vehicle on 253 the same route during the same periods of a day, there are noticeable differences in the fuel 254 consumption rates among the 60 RDE trips. Such differences exist among both the individual drivers 255 and the two trips of the same driver. They are mainly caused by the different driving behaviours and 256 traffic conditions of each RDE trip. Each driver adopts different driving styles in response to the 257 dynamic traffic conditions in a RDE trip, which is reflected in the use of the accelerator and brake 258 pedals. Drivers using the accelerator pedal more aggressively generally consume more fuel and 259 produce higher pollutant emissions [50]. In spite of the large differences among individual RDE 260 trips, Fig. 2a shows that the fuel consumption rate of a trip with DPF regeneration is significantly 261 higher than that without DPF regeneration by the same driver, except for Driver 9. This is because 262 Driver 9 had a relatively short incomplete DPF regeneration at the end of that RDE test (Fig. S3), 263 and thus its impact on the trip-averaged fuel consumption rate is insignificant. Overall, the average 264 and median fuel consumption rates of DPF regeneration trips are 13% and 15% higher than those of 265 normal trips, respectively (statistically significant difference exists, $p = 5.40 \times 10^{-5}$) (Fig. 2b). Active 266 DPF regenerations need to burn extra fuel to achieve the high exhaust temperature for soot oxidation, 267 which results in the increased fuel consumption rates. This study demonstrates that the real-driving 268 fuel consumption could be increased by 13% on average for a short trip of 19 km if a DPF 269 regeneration event occurs.



Fig. 2. Comparison of real-driving fuel consumption performance between trips with and without DPF regenerations: a) individual drivers, b) overall statistics (boxplot: 25^{th} and 75^{th} percentiles, box; median, centre line; mean, plus; minimum and maximum, whiskers). * $p = 5.40 \times 10^{-5}$ by unpaired one-tailed t-tests.

275 *3.2. Effect of DPF regenerations on gaseous and particulate emissions*

276 Figs. 3 and 4 compare the real-driving emission factors of CO and THC between trips with and 277 without DPF regenerations, respectively. Generally, CO and THC emissions of diesel vehicles are 278 extremely low due to the lean combustion mechanism of a diesel engine's diffusion flame. During 279 RDE tests, the measured instantaneous CO and THC concentrations were mostly below the detection 280 limits of the PEMS gas analysers. As shown in Figs. 3a and 4a, the CO and THC emission factors 281 are mostly around zero. In addition, no obvious trends can be observed between normal and DPF regeneration trips for the same driver. Among the nine drivers experiencing DPF regenerations 282 283 during their trips (Fig. 3a), five drivers (i.e. Drivers 1, 4, 6, 22 and 29) produced higher CO emission 284 factors while the remaining four drivers (i.e. Drivers 9, 12, 16 and 26) had lower CO emission factors 285 in DPF regeneration trips than normal trips. Regarding THC emissions (Fig. 4a), six drivers (i.e. 286 Drivers 1, 4, 9, 16, 22 and 29) had higher emission factors during DPF regeneration trips than normal 287 trips, while the remaining three drivers (i.e. Drivers 6, 12 and 26) experienced slightly lower 288 emission factors. The overall statistics (Figs. 3b and 4b) show that DPF regeneration trips have 289 higher CO and THC emission factors than the normal trips, by 241%/298% and 82%/47% for the

290 mean/median CO and THC emissions, respectively (statistically significant difference exists for 291 THC ($p = 4.63 \times 10^{-2}$) but not for CO ($p = 1.11 \times 10^{-1}$)). The emission factor of CO is affected by a 292 large scattering, thus a statistically significant difference does not exist between normal and DPF 293 regeneration trips. This can be explained by the fact that the post-combustion used to achieve high 294 exhaust temperature for active DPF regeneration takes place in the burnt gas from the main 295 combustion event, which contains a lower oxygen concentration and thus tends to produce more CO 296 and THC emissions. Nevertheless, such increases are not a concern for the studied vehicle, as even 297 their maximum values are still much lower than the Euro 5 emission limits of 0.74 g/km for CO and 298 0.07 g/km for THC.



Fig. 3. Comparison of real-driving CO emission factors between trips with and without DPF regenerations: **a)** individual drivers, **b)** overall statistics (boxplot: 25^{th} and 75^{th} percentiles, box; median, centre line; mean, plus; minimum and maximum, whiskers). * $p = 1.11 \times 10^{-1}$ by unpaired one-tailed t-tests.



Fig. 4. Comparison of real-driving THC emission factors between trips with and without DPF regenerations: a) individual drivers, b) overall statistics (boxplot: 25^{th} and 75^{th} percentiles, box; median, centre line; mean, plus; minimum and maximum, whiskers). * $p = 4.63 \times 10^{-2}$ by unpaired one-tailed t-

tests.

307

308 Diesel vehicles are the dominant sources of NO_x emissions although they only account for a 309 small proportion of the total fleet [5, 6, 11]. In particular, the Volkswagen scandal in 2015 revealed 310 that real-driving diesel NO_x emissions significantly exceeded the laboratory type-approval limits 311 [20, 21]. Following this, increasing studies have reported higher real-driving diesel NO_x emissions 312 than their respective emission limits, which were not limited to Volkswagen diesel vehicles [52-55]. 313 Fig. 5 shows the real-driving NO_x emission performance of trips with and without DPF 314 regenerations. As shown in Fig. 5a, real-driving NO_x emission factors of the studied diesel vehicle 315 vary greatly from 0.31 to 2.37 g/km, which are constantly higher than its NO_x limit of 0.28 g/km. 316 DPF regenerations generally have insignificant influences on NO_x emission factors. For the nine 317 drivers who experienced DPF regenerations (Fig. 5a), three drivers (i.e. Drivers 4, 9 and 29) have 318 noticeably lower NO_x, one driver (i.e. Driver 26) has noticeably higher NO_x, and the rest five drivers 319 (Drivers 1, 6, 12, 16 and 22) have very similar NO_x from DPF regeneration trips than those from normal trips. As a result, no statistically significant difference exists ($p = 4.70 \times 10^{-1}$) between the 320 mean NO_x emission factors of trips with and without DPF regenerations (Fig. 5b). 321

322 It should be noted that NO_x emissions collectively refer to both NO and NO₂. The latter is a 323 major air pollutant and also a much stronger oxidizer than O₂ for burning the PM captured in a DPF. 324 While automotive standards set limits for total NO_x emissions, air quality standards only regulate 325 NO₂ concentrations which are closely linked to a range of respiratory diseases. This may result in a 326 counter-intuitive situation that roadside NO₂ concentrations do not decrease accordingly with the 327 reduction of total NO_x emissions [23]. Therefore, the ratio of NO₂/NO_x is of great importance for 328 air quality management, as well as for DPF regenerations. Fig. 6 compares the instantaneous 329 NO₂/NO_x ratios of trips with and without DPF regeneration. The raw NO_x emissions generated in 330 diesel combustion are mostly (~90%) in the form of NO if without any after-treatment systems [56]. 331 Since NO₂-based regeneration is more energy efficient and can occur under a lower temperature 332 than O₂-based regeneration, diesel vehicles usually use a DOC to promote the conversion of NO to 333 NO₂ for continuous passive DPF regenerations. As a result, the NO₂/NO_x ratio could be relatively 334 high when exhaust temperature is low during normal engine conditions, as shown in Fig. 6a. 335 However, as the exhaust temperature increases for active DPF regenerations, the NO₂/NO_x ratio is 336 sharply reduced to below 10% (Fig. 6b). This could be mainly caused by two reasons. Firstly, DOC 337 starts NO oxidation at around 150-200 °C. The NO-to-NO2 conversion efficiency peaks at around 338 300 °C and reduces rapidly with a further increase of exhaust temperature [57, 58]. Secondly, NO₂ 339 is quickly consumed by PM oxidation before O₂ during DPF regeneration, because NO₂ is a much 340 stronger oxidiser than O₂. Such processes do not affect the total NO_x, but only reduce the NO₂/NO_x ratio via lesser NO oxidation by the DOC (NO + $\frac{1}{2}$ O₂ \rightarrow NO₂) and higher NO₂ reduction by the 341 342 DPF (NO₂ + $\frac{1}{2}$ C \rightarrow $\frac{1}{2}$ CO₂ + NO). The results (Figs. 5 and 6) suggest that DPF regenerations have 343 little influence on the total NO_x emissions, but can significantly lower the NO₂/NO_x ratios.







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350 351

Fig. 6. Variations of instantaneous NO_2/NO_x ratio and exhaust temperature with time of Driver 1: a) morning test without DPF regeneration, b) afternoon test with DPF regeneration.

352 PM is another major air pollutant that is dominated by diesel vehicles [6]. Fig. 7 compares the 353 real-driving PM emission performance of trips with and without DPF regenerations. As shown in 354 Fig. 7a, DPF regenerations significantly increased the trip-averaged PM emission factors, except for 355 Driver 9 who experienced a very short and incomplete DPF regeneration event (Fig. S3). For trips 356 without DPF regeneration events, PM emission factors are relatively low although only a few trips 357 could meet the Euro 5 PM limit of 0.0045 g/km. However, all nine DPF regeneration trips far 358 exceeded the PM limit, by up to 242 times. Overall, the mean and median PM emission factors of 359 DPF regeneration trips are 27 and 46 times higher than those of normal trips, respectively 360 (statistically significant difference exists, $p = 2.74 \times 10^{-4}$) (Fig. 7b). Such increases are similar to a 361 previous study which observed 34 times increases of real-world PM emissions during PM 362 regeneration for Euro IV-VI medium diesel trucks [47]. Three reasons have contributed to the sharp

363 increases of PM emissions during active DPF regenerations [33]. Firstly, DPF regeneration lowers 364 the filtration efficiency and enhances the penetration of particles through the DPF. Secondly, DPF 365 regeneration changes the structure of the deposited soot layer, which produces new or secondary 366 particles to escape from the DPF. Finally, DPF regeneration can cause the nucleation of semi-volatile 367 particles when gaseous sulphuric acid exits the DPF. It should be noted that the test route is relatively 368 short in this study (19 km, 25-30 min) and the DPF regeneration event accounts for over one third 369 of the test time. In the real-world driving, DPF regenerations do not occur frequently, e.g. every 130 370 km or 7 trips for the studied vehicle. The impact of DPF regeneration on trip-averaged PM emission 371 factor should be lower for longer test routes. For example, the overall average PM emission factor 372 of all 60 RDE trips (51 normal + 9 regeneration trips) was significantly reduced to be only 4.1 times 373 higher than that of 51 normal trips.



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Fig. 7. Comparison of real-driving PM emission factors between trips with and without DPF regenerations: a) individual drivers, b) overall statistics (boxplot: 25^{th} and 75^{th} percentiles, box; median, centre line; mean, plus; minimum and maximum, whiskers). * $p = 2.74 \times 10^{-4}$ by unpaired one-tailed t-tests.

378

379 4. Conclusions

This study investigated the impact of DPF regenerations on fuel consumption and emissions performance of a diesel LGV under real-driving conditions. A total of 60 RDE trips (~1200 km in total) were performed with the aid of the state-of-the-art PEMS, among which nine real-driving DPF 383 regeneration events were measured. All the RDE trips were driven using the same vehicle on the 384 same route during the same periods of a day, aiming to minimise the effect of environmental and 385 vehicle configuration factors on the analysis. The results showed that, for a 19-km trip, DPF 386 regeneration increased the trip-averaged fuel consumption rate by 13% on average. Regarding 387 pollutant emissions, DPF regenerations tended to increase CO and THC emission as the post 388 combustion was taking place under oxygen-lean conditions. Nevertheless, such impacts were not a 389 concern as CO and THC emission factors were still well below the emission limits even under the 390 impact of DPF regeneration. NO_x emission factors of all RDE trips exceeded the emission limit. 391 DPF regenerations had little influence on the total NO_x emissions, but significantly reduced the 392 NO₂/NO_x ratio due to lower NO oxidation by the DOC and higher NO₂ reduction at DPF. Finally, 393 PM emission factors were relatively low for normal trips but were sharply increased by 27 times on 394 average when DPF regeneration occurred, leading to significant exceedance of emission limit. The 395 findings therefore suggest that a cleaner combustion control that reduces the regeneration frequency 396 and a better DPF design that improves the regeneration quality are needed to reduce the real-driving 397 fuel consumption and PM emissions of diesel vehicles during DPF regenerations.

398

Declaration of competing interest

400 The authors declare that they have no known competing financial interests or personal401 relationships that could have appeared to influence the work reported in this paper.

402

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407 **Supplementary material**

- 408 Additional information about the test vehicle (Fig. S1), RDE test route (Fig. S2), and exhaust
- 409 temperature profiles of DPF regenerations (**Fig. S3**).
- 410

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