

1 **Effects of the COVID-19 pandemic on the environment, waste management, and energy**
2 **sectors: A deeper look into the long-term impacts**

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26 **Effects of the COVID-19 pandemic on the environment, waste management, and energy**
27 **sectors: A deeper look into the long-term impacts**

28 **Abstract**

29 The COVID-19 pandemic has not only caused a global health crisis but also has significant
30 environmental consequences. Although many studies are confirming the short-term improvements
31 in air quality in several countries across the world, the long-term negative consequences outweigh
32 all the claimed positive impacts. As a result, this review highlights the positive and the long-term
33 negative environmental effects of the COVID-19 pandemic by evaluating the scientific literature.
34 Remarkable reduction in the levels of CO (3–65%), NO₂ (17–83%), NO_x (24–47%), PM_{2.5}
35 (22–78%), PM₁₀ (23–80%), and VOCs (25–57%) was observed during the lockdown across the
36 world. However, according to this review, the pandemic put enormous strain on the present waste
37 collection and treatment system, resulting in ineffective waste management practices, damaging
38 the environment. The extensive usage of face masks increased the release of
39 microplastics/nanoplastics (183 to 1247 particles/piece) and organic pollutants in land and water
40 bodies. Further, the significant usages of antibacterial hand-sanitizers, disinfectants, and
41 pharmaceuticals, have increased the accumulation of various toxic emerging contaminants (e.g.,
42 triclocarban, triclosan, bisphenol-A, hydroxychloroquine, etc.) in the treated sludge/biosolids and
43 discharged wastewater effluent, posing great threats to the ecosystems. This review also suggests
44 strategies to create long-term environmental advantages. Thermochemical conversions of solid
45 wastes including medical wastes and for treated wastewater sludge/biosolids offer several
46 advantages through recovering the resources and energy, and stabilizing/destroying the
47 toxins/contaminants and microplastics in the precursors.

48 **Keywords:** COVID-19; Air quality; Face masks; Emerging contaminants; Solid wastes; Pyrolysis;
49 Hydroxychloroquine/chloroquine

50 **Introduction**

51 COVID-19 has spread over the world, causing a global health emergency. It was first reported in
52 Wuhan, China, in December 2019 and soon spread throughout the country, becoming a global
53 disease (Ali and Alharbi 2020; Zambrano-Monserrate et al. 2020). Coronavirus may have been
54 transmitted from bats to people, according to one hypothesis (Brennecke et al. 2020). Humans
55 were infected mostly by droplets, interactions, and airborne transmissions. COVID-19 has made a
56 great impact all across the world. Almost all countries are currently focusing their efforts to prevent
57 the spread of COVID-19 disease by enacting policies such as the complete closure of public places
58 (Liu et al. 2021; Muhammad et al. 2020; Naethe et al. 2020). Though these policies or regulations
59 may have a significant impact on the economies of most countries, they have some negative or
60 positive effects on the environment, such as a substantial drop in the greenhouse gas emissions
61 that have not been seen since World War II, as many industries in the world halted production and
62 vehicle usage drastically decreased. These variables resulted in a significant drop in nitrogen
63 dioxide (NO₂) and particulate matter (PM_{2.5}, diameter smaller than 2.5 μm) concentrations all
64 across the world (Cui et al. 2020; Muhammad et al. 2020; Liu et al. 2021).

65 To prevent the rise of coronavirus sickness, every country's government has established social
66 distancing regulations and forced its inhabitants to remain in their homes. Noise pollution levels
67 have decreased dramatically in most countries due to reduced public and private transportation and
68 other business operations (Jairoun et al. 2021). A critical link exists between emergency control
69 and the betterment of pristine beaches, air quality, and environmental noise depletion. Conversely,
70 additional harmful indirect factors, like minimizing recycling, and the rise in effluents, threaten
71 the pollution of land and water bodies even more, besides air (Zambrano-Monserrate et al. 2020).

72 For instance, the demand for conventional energy has decreased by approximately 30% in
73 several regions, and there has been a decrease in power consumption ranging from 12 to 20% for

74 most countries (Mousazadeh et al. 2021). Although the pandemic had some direct and short-term
75 positive consequences, it has more severe negative effects on plastic production and solid waste
76 management. Unfortunately, in many countries, the COVID-19 pandemic has wreaked havoc on
77 municipal solid waste management (Knowlton 2019). Some states in the United States, for
78 example, temporarily closed recycling centers, ceased curbside recycling collection, and stopped
79 onboard recyclables pickups owing to fears about the virus's spread. Pesticides, detergents, soaps,
80 single-use plastic, and other chemicals have surged dramatically in recent months, wreaking havoc
81 on the ecosystem (Baldasano 2020). In the last few months, roads, buildings, and entire cities have
82 been sanitized. The utilization of hand sanitizers has risen dramatically as well. Personal protective
83 kits, gloves, and face masks have emerged as important safety precautions during the continuing
84 epidemic (Neumeyer et al. 2020).

85 Hand sanitizers containing isopropanol and alcohol are becoming more popular worldwide as
86 a means of mass disinfection. In almost every location where people live, disinfectants such as
87 hypochlorous acids, sodium hypochlorite, and chlorine are utilized in large quantities (Kumar et
88 al. 2020; Tripathi et al. 2020; Capoor and Parida 2021). In addition, the excessive usage of
89 antibiotics, disinfectants, antibacterial hand sanitizers and soaps during the COVID-19 pandemic
90 will raise the concentration of these pharmaceutical and personal care products and other emerging
91 pollutants in the discharged wastewater effluent from wastewater treatment plants (WWTPs),
92 posing great threats to the aquatic life and ecosystems (Neumeyer et al. 2020; Rugani and Caro
93 2020). The extraordinary use of pharmaceuticals (e.g., antibiotics, disinfectants) and personal care
94 products (e.g., antibacterial soaps and hand-sanitizers) which contains triclocarban (antibacterial)
95 and triclosan (fungicide) (Du et al. 2019; Patel et al. 2019; Brennecke et al. 2020), will noticeably
96 increase their concentrations in the WWTPs influent and effluent. For the microorganisms and
97 environment, sodium hypochlorite is extremely toxic. Endocrine disruption and various other

98 neurological consequences are all linked to these substances. Both compounds are harmful to the
99 environment since they are hard to break down and makeup 60% of all drugs identified in sewage
100 sludge. The aquatic fauna is adversely affected by these chemical pollutants (Kumar et al. 2020;
101 Rizou et al. 2020; Capoor and Parida 2021).

102 Several studies and reviews discussing the positive impacts of COVID-19 on the environment
103 have been punished, while limited published studies are studying the long-term negative impacts
104 of COVID-19 on the environment and waste management. The present review paper discusses the
105 positive and long-term negative impacts of COVID-19 on the environment, waste management,
106 and energy sectors. It is very important to evaluate the long-term negative impacts of COVID-19
107 on the environment for future strategies and regulations. Furthermore, the information presented
108 in this review can be used to assess short and long-term options for mitigating future harmful
109 effects.

110

111 **Short-term positive impacts of COVID-19 on the environment**

112 Many studies confirmed that air quality improved only during the lockdown period in various
113 countries all over the world. Major reductions in the levels of NO₂, CO, NO_x, SO₂, VOCs, and
114 PM_{2.5} were reported, which is mainly due to the lockdown and closure of many industrial and
115 commercial activities in different countries including the United States (Chen et al. 2020; Son et
116 al. 2020), Canada (Adams, 2020), Mexico, Italy, Germany, China (Jia et al. 2020a,b; Wang et al.
117 2021a), India (Selvam et al. 2020; Yunus et al. 2020), Iraq (Hashim et al. 2021), Egypt (Mostafa
118 et al. 2021), etc (Table S1, Supplementary material). The details regarding the impacts of
119 COVID-19 on greenhouse gases emissions, VOCs, particulate matter, and air quality index are
120 given in the following sections.

121

122 *Greenhouse gases (GHGs) emissions*

123 In the United States, the pollution levels in the COVID-19 phase from March 13 to April 21, and
124 before the COVID duration, from January 8 to March 11 were recorded in 2020. Notable NO₂
125 reductions were detected in 2020 than from 2017 to 2019, where a 25.5% reduction, including
126 absolute mitigation of 438 ppb was recorded (Berman and Ebisu 2020). Another study showed
127 that the data obtained from 28 long-term air quality sites demonstrated varying reduction levels of
128 CO and NO₂ in the initial lockdown period of March 15 to April 25, 2020, in the US, corresponding
129 to a reference duration, before lockdown, including previous baselines instituted from 2017 to
130 2019. The approximate reduction levels of CO by 37% and NO₂ by 49% are notable in around
131 67% of the regions and principally proliferate with the regional population density (Chen et al.
132 2020). Declines of 49 and 38% in CO and NO₂ levels were found at base level analyses in
133 California, throughout the lockdown from March 19 to May 7 than before the lockdown on January
134 26 to March 18 in 2020 (Figure 1). Similarly, in Ontario, Canada, NO₂ and NO_x exhibited their
135 lowest levels for 22 of the 29 recorders. Distinctive readings varied from 1 ppb of NO₂ and 5 ppb
136 of nitrogen oxides (NO_x), which was above average to 4.5 ppb of NO₂ and 7.1 ppb of NO_x, which
137 was lower than average (Adams 2020).

138 In Düsseldorf, Germany, the continuous observations of NO₂ level variations corresponding to
139 the reduction of city traffic because of the lockdown enforcement were performed using down-
140 welling light including RoX automated field-spectrometer. A decision tree was established by
141 primary constituents that were disintegrated from down-welling radiance spectra, which showed
142 to be the highest durable methodology to obtain NO₂ readings. Improved differentiation of the
143 NO₂ readings-scale was derived with a partial least square regression model, and down-welling
144 radiance measurements can be used to observe NO₂ levels continuously (Naethe et al. 2020).

145 In Italy, a study was conducted to measure the carbon footprint (CF) related to energy reduction
146 because of economic undertakings and locations throughout the lockdown in the nation, and to
147 equate these environmental obstacles with the quantified CF for corresponding durations around
148 March and April from 2015 to 2019. The study demonstrates that CF during lockdown reduced by
149 approximately 20% more than the average computed CF before. This indicates that greenhouse
150 gases or GHGs were regulated within approximately 5.6 and 10.6 Mt CO_{2e}. Further studies
151 indicate an inclination that happens toward larger impact savings in the Northern provinces, which
152 is approximately 230 kt CO_{2e} of GHGs regulated by provinces on average, than approximately
153 110 to 130 kt CO_{2e} in the Southern and Central provinces (Rugani and Caro 2020).

154 In Madrid and Barcelona, Spain, the magnitude of the air pollution decline was measured in
155 March and April 2020, during the lockdown, which demonstrated readings of a notable reduction
156 of approximately 75%. The NO₂ levels in these two cities recorded a 50 and 62% reduction for
157 Barcelona and Madrid, respectively. In March 2020, the hourly quantification readings were
158 derived from 9 and 24 air quality sites from the recording systems, which provided data of the
159 regulations, that can be accomplished by imposing low emission zones (LEZ), including the
160 pollution level to be eradicated, which was 55% for Barcelona and Madrid (Baldasano 2020).
161 Average monthly concentrations of pollutants in Tunisia examined from January 1 to April 30,
162 2020, revealed that actions taken to reduce the spread of the virus have a major impact on emission
163 levels. In March, there have been 51% decreases for NO₂ and 52% for SO₂ emissions for most of
164 the cities compared to January, while the levels of NO₂ and SO₂ decreased by about 40% (Chekir
165 and Ben Salem 2021).

166 To investigate the effect of the lockdown on air pollution in Egypt with a focus on Cairo and
167 Governorates of Alexandria, the data for the lockdown period in 2020 were extracted and
168 contrasted with the corresponding month for the specified baseline period (2015–2019). The

169 absorbing aerosol index (AAI) was reduced by around 30%, NO₂ declined by 15 and 33%
170 respectively in Cairo and Alexandria Governorates, while CO reduced by around 5% in both
171 Governorates. Furthermore, during the epidemic, GHG emissions have been cut by at least 4%,
172 while the amount of ozone in Cairo and Alexandria rose by around 2% (Mostafa et al. 2021). In
173 the United Arab Emirates, core data reveal that the AOD, NO₂, and Surface Urban Heat Island
174 Intensity (SUHII) levels decreased in the lockdown period by 3.7, 23.7, and 19.2% in comparison
175 with the same time of 2019 correspondingly. This study has shown that measurements of chosen
176 air contaminants and SUHII data from satellites correspond highly to actual data (Alqasemi et al.
177 2021).

178 In São Paulo, Brazil, the lockdown was enforced on March 24, 2020, to reduce COVID-19
179 cases. The Business as Usual or BAU days with homogeneous weather situations were
180 incorporated to decrease the impact of meteorology on ambient levels. The largest average level
181 or \pm standard deviation, on BAU days, was recorded at Congonhas station at $125.8 \pm 27.4 \mu\text{g m}^{-3}$,
182 based 500 m from Congonhas airport; and the lowest value was recorded at Pico de Jaragua at 17.3
183 $\pm 5.2 \mu\text{g m}^{-3}$ in the city suburbs. The highest absolute reduction in NO_x levels was observed at
184 Congonhas at $-79.6 \mu\text{g m}^{-3}$, succeeded by Marginal Tiete at $-59.2 \mu\text{g m}^{-3}$ at the curbside of a
185 notable-trafficked region. A significant decline was recorded for all sites with depreciation ranging
186 from 34% at Osasco to 68% in Pinheiros. The traffic emission depletion led to air quality
187 amelioration because it constitutes 67% of NO_x emissions (Krecl et al. 2020). Other studies in
188 Brazil showed significant decreases in NO levels of 77.3%, CO of 64.8%, and NO₂ of 54.3%, were
189 detected throughout the incomplete lockdown in urban regions than the five-year monthly average
190 (Nakada and Urban 2020; Urban and Nakada, 2021).

191 In Central Eastern China (CEC), the angstrom exponent (AE) and the aerosol optical depth
192 (AOD) in the lockdown from January 24 to February 29, 2020, rose and fell in many CEC places

193 compared with the 21-year climatological mean from 2000 to 2020. In Wuhan and Hubei, the
194 AOD or AE outputs reduced or rose by 31.0 to 45.3% and 39.2 to 29.4%, due to strict lockdown
195 (Figure 1) (Shen et al. 2021a). The industrial activities, ongoing constructions, vehicle kilometers
196 traveled, etc., fell tremendously in China, thereby causing less NO_x, and SO₂ levels by roughly 29
197 to 47%, and 16 to 26% throughout phase 1 and phase 2 response duration and ultimately enhanced
198 the air quality during the national lockdown (Li et al. 2020). In Taiwan, the NO₂ column levels
199 also fell by 24% in the third week of the Chinese New Year in 2020 than earlier (Griffith et al.
200 2020). In Shijiazhuang, China, nitrous acid (HONO) levels were quantified by implementing a
201 recorder for aerosols and gases in ambient air or MARGA from December 15, 2019, to March 15,
202 2020, which included the Chinese New Year (CNY), the high air pollution period, and the COVID-
203 19 lockdown phase. The air quality notably ameliorated due to the diffusion propensity
204 proliferation of air clouds and the emission depletion in and post CNY, which incorporated the
205 COVID-19 lockdown. The average HONO readings were 2.43 ± 1.08 ppbv before CNY; it abated
206 to 1.53 ± 1.16 ppbv in the CNY and further mitigated to 0.97 ± 0.76 ppbv post CNY. Following
207 the incorporation of the amelioration of the diffusion propensity, an approximate depletion of 31%
208 of ambient HONO, 36% of NO₂, and 62% of NO (Figure 1 and Table S1), were recorded in the
209 lockdown phase, throughout and post CNY, then with those before CNY (Jia et al. 2020b). Another
210 study was conducted to monitor NO₂ levels in 367 main cities in Northern China, which were
211 separated into low, moderate, and extreme emission categories, as per their four-year mean NO₂
212 rating. NO₂ levels were determined every day for all emission categories, categorized the days as
213 good, medium, and bad-meteorological conditions (Jia et al. 2020).

214 The measurement of NO₂, HCHO, SO₂, and CO levels, including the Aerosol Optical Depth
215 (AOD) over Tokyo, Seoul, Wuhan, and Beijing-Tianjin-Hebei (BTH) region was undertaken in
216 February 2019 and February 2020. NO₂ recorded the highest critical reductions of approximately

217 19, 33, 83, and 54% in Tokyo, Seoul, Wuhan, and BTH, respectively. Wuhan experienced the
218 highest reductions in contaminants, by roughly 62% in AOD, with a 4, 71, 11, and 83% fall in
219 column densities of CO, SO₂, HCHO, and NO₂, respectively. SO₂ levels climbed in Tokyo and
220 Seoul because of vehicles coming from contaminated upwind areas, whereas the formaldehyde,
221 CO, and NO₂ levels declined. A large rise in surface ozone in East China from February 19 to
222 February 2020 was due to remarkable falls in NO₂ levels, which is possibly due to lower O₃ and
223 NO reaction, by high NO_x dips and low NO_x saturation (Ghahremanloo et al. 2021).

224 In Hat Yai, Thailand, The NO₂ levels abated by 33.7% (Figure 1), respectively in the initial
225 three weeks of the lockdown period than the similar phase, before the lockdown period. The NO₂
226 tropospheric readings recorded over the ground site and regionally including the spatial mean over
227 the urban regions of the city were obtained from Sentinel-5P, which corresponds with the depletion
228 levels derived from the ground site (Stratoulis and Nuthammachot 2020).

229

230 *Particulate matter and VOCs*

231 Several studies across the world reported a major reduction in particulate matter (PM) and VOCs.
232 PM_{2.5} levels demonstrated a reduction in the COVID-19 phase in the United States (Berman and
233 Ebisu 2020). In California, the variations in PM_{2.5} levels, which were around 2.5 µm were
234 quantified in the mitigation period, versus the baseline or pre-mitigation quantification phase, by
235 implementing the difference-in-difference methodology, and the approximate avoided overall and
236 cause-specific fatality, due to PM_{2.5} variations by district or state in the urban regions of 10 US
237 states and the District of Columbia. The PM_{2.5} levels declined in seven states, including the capital,
238 in the abatement phase. The mean PM_{2.5} level depreciation was approximately 0.25 µg m⁻³ or 4.3%
239 in Maryland to 4.20 µg m⁻³ or 45.1% in California. PM_{2.5} levels mitigated by 12.8% on average,
240 in the capital, including the seven states. An approximate 483 or 95% CI: 307,665 fatalities,

241 corresponding to PM_{2.5} levels, were avoided in the urban region of California (Son et al. 2020). A
242 decline of 31% in PM_{2.5} level was found in California throughout the lockdown from March 19 to
243 May 7 than before the lockdown on January 26 to March 18 in 2020 (Liu et al. 2021). Similarly,
244 the PM₁₀ and PM_{2.5} levels notably reduced in the Northeast cities of California and Nevada (Chen
245 et al. 2020).

246 In India, the analysis of the air pollutants falls throughout the lockdown of March 24, 2020, and
247 April 20, 2020, was done by differentiating against the pre-lockdown phase of January 1, 2020,
248 and March 23, 2020. The Central Pollution Control Board or CPCB reported lower air pollution
249 throughout the lockdown with significant falls in air pollution in zone 2 comprising Gandhinagar
250 and Ahmedabad and zone 3 of Rajkot and Jamnagar and slight fall in zone 1 comprising Vadodra,
251 Surat, and Ankleshwar and zone 4 including Palanpur and Bhuj in the industrialized state of
252 Gujarat. The concentration of PM_{2.5} and PM₁₀, plummeted in the range of 38 to 78% and 32 to
253 80%, respectively (Selvam et al. 2020). In Ghaziabad, the lockdown led also to a substantial
254 amelioration in environmental variables like air quality. The PM_{2.5} and PM₁₀ levels demonstrated
255 reductions of approximately 85.1 and 50.8%, respectively in the city than on January 14, 2020,
256 before the lockdown (Lokhandwala and Gautam 2020).

257 A separate spatial dispersion of the surface PM_{2.5} was observed over Central Eastern China
258 (CEC) throughout the lockdown, including increased levels in East and North China.
259 Comparatively, high PM_{2.5} levels were observed in the low flatlands of Hubei, where six distinct
260 occasions of PM_{2.5} contamination were discovered. Half of the occurrences of contamination were
261 linked to long-range transport (LRT) of air contaminants from upstream areas of CEC (Shen et al.
262 2021a). Similarly, the bad-meteorological conditions experienced a 56% greater PM_{2.5} level in
263 extreme emission cities than low emission cities. Concerning the good meteorological conditions,
264 the extreme emission category rose to 8.8 $\mu\text{g m}^{-3}$ in regular mean PM_{2.5}, from 2017 to 2019,

265 including a 2.6% rise in the likelihood of extreme PM_{2.5}. But, the extreme emission, bad
266 meteorological category witnessed a 24% dip in standard mean PM_{2.5} levels from 2017, an extreme
267 emission year compared to 2019, which was a low emission year (Jia et al. 2020a). Taiwan avoided
268 an excess PM_{2.5} of 19.2 µg m⁻³, on average, throughout the event, similar to a 0.5 µg m⁻³ fall for
269 the entire three-month winter season (Griffith et al. 2020). Similarly, a decrease in industrial
270 activities, ongoing constructions, vehicle kilometers traveled, etc., caused reduced emission levels
271 of VOCs, and PM_{2.5} levels by roughly 37 to 57%, and 27 to 46% throughout phase 1 and phase 2
272 response duration and ultimately enhanced the air quality during the national lockdown (Li et al.
273 2020).

274 The air contaminants readings of PM_{2.5} and VOCs in the non-regulatory phase or NCP from 24
275 December 2019 to 23 January 2020, and the control phase or CP from 24 January to 23 February
276 2020 were examined at Pudong Supersite or PD and the Dianshan Lake Supersite or DSL. The
277 prevalence of fog or haze-like incidents, VOCs, and PM_{2.5} levels fell significantly from NCP to
278 CP and the mean percentages were 35.1, 38.9, and 31.6% at PD, and 37.9, 50.7, and 34.5% at DSL.
279 The emission regulation of major sources like vehicle fumes and coal combustion caused low
280 forerunner readings of NO_x and SO₂, which resulted in the decline of PM_{2.5} from NCP to CP. The
281 inadequate nitrogen monoxide titration, low relative humidity, and greater visibility than NCP may
282 have caused extreme ozone levels at DSL and PD in CP. The VOCs also fell by controlled
283 vehicular fumes and fugitive emissions, and the regulation of locally circulating air pollutants led
284 to fog level dips (Jia et al. 2020b; Hashim et al. 2021).

285 The hourly PM_{2.5} levels of related elements were quantified at a rural area between Beijing and
286 Tianjin, China, before 12 to 25 January 2020; from 26 January to 9 February 2020, and post 22
287 March to 2 April 2020 in the control phase. Fe, Ca, K, Zn, Ba, and Cu were the primary elements,
288 before the control period; and Zn smelter, vehicle emissions, and fireworks combustion constituted

289 the largest proportions of the overall element mass of 12.1, 10.3, and 55%, respectively. K, Fe, Ba,
290 Cu, and Zn were the principal elements in the control period; and vehicle emissions and fireworks
291 combustion constituted 27 and 55% of the overall element mass. Fe, K, Ca, Zn, and Ba were the
292 primary elements, post lockdown; and steel, dust, and the iron sector constituted 21 and 56% of
293 the overall element mass (Cui et al. 2020).

294

295 *Air quality index and O₃*

296 The Ontario region in Canada announced a State of Emergency (SOE) in March 2020 to restrict
297 the COVID-19 cases. A five-week phase in the SOE period was analyzed against an earlier five-
298 week duration as a benchmark. Ozone concentrations at 12 of the 32 recorders were less than the
299 preceding five-year phase. The mean ozone levels were 1 ppb less in the SOE phase. However, it
300 varied from 1.5 ppb up to 4.2 ppb, which was lower than the long-lasting circumstances (Adams
301 2020). Conversely, increase in ozone levels was found in Rio de Janeiro, Brazil, due to the rise in
302 NMHC/NO_x throughout social distancing, enforced due to the atmospheric chemistry in Rio de
303 Janeiro, regulated within VOC-monitored circumstances. The proportions of non-methane
304 hydrocarbons and nitrogen oxides or NMHC/NO_x increased around 37.3% throughout the
305 incomplete lockdown, according to the surveying data derived from two automated surveying
306 sites. However, the rise was substantial when air clouds appeared from industrial regions, due to
307 the large rise in VOC reactivity from the clouds, which are high in aromatic compounds (Siciliano
308 et al. 2020). Similarly, an approximate rise of 30% in ozone was noticed in urban regions, largely
309 determined by automobile traffic, possibly associated with nitrogen monoxide reduction (Nakada
310 and Urban 2020; Urban and Nakada, 2021).

311 The effect of COVID-19 regulations was also investigated in Europe from 15 March to 30 April
312 2020, according to maximal regular 8-h running average ozone or MDA8 O₃ from European

313 Environment Agency's air quality database. The MDA8 O₃ levels were reduced in Iberia because
314 of high specific humidity and low solar radiation, while ozone escalated in different regions.
315 Northwestern to Central European area showed a notable rise in O₃ levels of 10 to 22% in urban
316 background sites. The measurements of the approximate O₃ demonstrated that declined due to
317 emission reductions, and O₃ variations were determined by meteorology (Ordóñez et al. 2020). In
318 India, power plant operations probably caused a minimal fall in CO by 3 to 55% and the reduced
319 discharge of NO enhanced the O₃ levels by 16 to 48% (Figure 1). AQI recovery of 58%, in general,
320 was reported for the initial four months of 2020 than in 2019 (Selvam et al. 2020).

321

322 *Freshwater quality*

323 There is a noteworthy correlation between emergency procedures, enhancement of air quality,
324 pristine beaches, and environmental sound depletion. Freshwater quality for some lakes was also
325 improved regarding the suspended particulate matter or SPM in certain countries, including India,
326 and the SPM output decreased up to 34% than the earlier years before the lockdown (Yunus et al.
327 2020). Figure 2 shows the tracking of the spread of SARS-CoV-2 in near real-time with a map-
328 centric dashboard developed by Johns Hopkins University's Center for Systems Science and
329 Engineering dashboard leading the pack (Johns Hopkins 2021). In Vembanad Lake of Southern
330 India, which is the longest freshwater lake in India, the surface water quality recovered regarding
331 the suspended particulate matter or SPM. SPM dropped by 15.9% on average from 10.3% to
332 36.4%, which is up to an 8 mg/l decline than the pre-lockdown phase (Figure 3). The computed
333 SPM for April 2020 is the least for 11 of the 20 Vembanad lake zones. The SPM output decreased
334 up to 34% than the earlier years from the preceding minima (Yunus et al. 2020). Conversely, there
335 are adverse indirect characteristics like decrease in recycling and rise in waste, that jeopardizes the
336 water and land pollution, including air, in the highly affected nations like Spain, Italy, USA, and

337 China, when analyzing the adverse and productive secondary outcomes of COVID-19 (Zambrano-
338 Monserrate et al. 2020).

339 In summary, mostly all studies across the world confirmed remarkable improvements in air
340 quality during the lockdown period because of the closure of several industrial and commercial
341 activities which rely on fossil fuels, leading to a remarkable reduction in GHG emissions,
342 particulate matter, and VOCs. Additionally, the global lockdown drastically affected energy
343 resources as discussed in the section “Impacts on the energy sources”, resulting in a remarkable
344 decrease in energy usage and reduction in GHG emissions. The results clearly show that there is
345 an urgent need to reduce the dependence on all fossil fuel-related energy sources, including oil,
346 coal and natural gas, and increase dependence on other renewable resources to have positive long-
347 term impacts on the global warming potential.

348

349 **Long-term negative impacts of COVID-19 on the environment**

350 *Generation of biomedical waste and management issues*

351 In several countries, the COVID-19 epidemic has resulted in not only health-related challenges
352 such as job, mental illness, economic losses, socio-economic burdens, and waste management
353 challenges (Le et al. 2020; Mofijur et al. 2021; Mousazadeh et al. 2021). It has put enormous strain
354 on the present waste collection and treatment system, resulting in ineffective waste management
355 practices like mobile incinerations and direct landfills for wipes, bottles of sanitizers, single-use
356 masks, and gloves, which are crucial for the safety of frontline Corona pandemic workers. In the
357 face of increasing panic, the single-use plastic manufacturing industry is attempting to seize the
358 chance and resurrect an otherwise dwindling industry (Somani et al. 2020; Mousazadeh et al.
359 2021). Many supermarkets no longer allow consumers to bring reusable bags instead of delivering
360 things in single-use plastic bags. There has been a rise in online shopping of meals in restaurants,

361 resulting in a per capita rise in plastic usage, indicating that global plastic pollution has increased
362 due to the COVID-19 pandemic. This will certainly result in a sharp increase in the use of plastics
363 and intensify the release of microplastics to the surrounding ecosystems. Plastic consumption has
364 increased dramatically (40%) and other applications (17%), including medical applications
365 (Boroujeni et al. 2021).

366 Hospitals generate many infectious and biological wastes for sample collection of suspicious
367 COVID-19 patients, testing, treating a significant number of people, and disinfection purposes.
368 COVID-19 patients can produce around 3.4 kilograms of medical waste per day. Thus, during the
369 epidemic, the amount of health waste has grown progressively (Das et al. 2021). For example,
370 during the outbreak in Wuhan, China, over 245 metric tons of biological waste were created every
371 day, over 190 metric tons more than usual. The number of healthcare waste generated in the Indian
372 city of Ahmedabad grew from 500-600 kg day⁻¹ to roughly 1000 kg day⁻¹ during the first phase of
373 the shutdown. COVID-19 generates approximately 206 million tons of biological waste a day in
374 Dhaka, Bangladesh's metropolis (Celis et al. 2021). Other cities, including Kuala Lumpur, Manila,
375 Hanoi, and Bangkok, had similar increases, creating between 154 and 280 million tons of medical
376 waste each day compared to before the pandemic (Chakraborty and Maity 2020). Healthcare waste
377 grew by 600% in Hubei, from 40 to 240 metric tons, overwhelming the infrastructure of transit
378 and disposal. Similar difficulties face other countries as far as treating the enormous volume of
379 garbage is concerned (Das et al. 2021).

380 Based on the review of the different papers, it can be assumed that perhaps the COVID-19
381 pandemic can have short-term favorable environmental benefits. However, the negative effects are
382 much more severe. Many countries, particularly Australia, are grappling with the disposal of used
383 personal protective equipment (PPE) composed of non-biodegradable plastics that really can take
384 hundreds of years to degrade in the environment. According to the current study's estimates, in

385 the first and second waves of the pandemic in Victoria, approximately 104-160 tons of the users'
386 face masks were manufactured daily (Neumeyer et al. 2020). During the 2nd wave of the epidemic
387 in Victoria, the respective mobility patterns of public transportation hubs, retail and entertainment
388 venues, and workplaces decreased by 85, 83, and 76%, respectively, when contrasted to the period
389 of 5 weeks between 3 January and 6 February 2020. PM_{2.5} levels were also reduced by 23% at
390 Alphington and 24% at Footscray (Aragaw 2020).

391

392 ***Personal protective equipment is a significant source of microplastics***

393 Proper plastic disposal after they have been used has become a major concern. It is a setback in
394 our campaign against plastic pollution as a whole. The products end up in locations where they are
395 not supposed to be. They are common on the streets, in natural habitats, and in the oceans. People
396 are desperately defending themselves from the illness; thus, masks are strewn across highways,
397 sidewalks, and parks. If the masks make it to the oceans, they could endanger marine life. The
398 proper disposal of PPE is also a concern (Oyedotun et al. 2020; Vanapalli et al. 2021). SARS-
399 CoV-2 is more persistent on stainless steel and plastic than on cardboard and copper, and live virus
400 particles can indeed be found up to 72 hours after being sprayed to these surfaces (Knowlton 2019).
401 Although latex rubber gloves are natural items, there are concerns that they are not necessarily
402 environmentally friendly. Chemicals employed in their production are hazardous to the
403 environment and disposing of all such wastewater is another issue (Selvaranjan et al. 2021;
404 Sridharan et al. 2021).

405 Face masks and other plastic-based protection equipment have been proposed as a possible
406 source of microplastic fibers within the environment. N-95 masks are usually made up of
407 polypropylene while Tyvek is used in protective gloves, suits, and medical face shields. Both the
408 microplastics can last a long time and leak dioxin and harmful substances into the environment

409 (Selvaranjan et al. 2021; Sridharan et al. 2021; Wang et al. 2021b). Though specialists and
410 responsible authorities recommend that domestic organic waste and plastic-based protective
411 equipment be properly disposed of and segregated, mixing these wastes increases the danger of
412 disease transmission and waste workers' exposure to the virus (Selvaranjan et al. 2021; Sridharan
413 et al. 2021).

414 The usage of face masks triggered the microplastic release from 183 to 1247 particles/piece
415 (Figure 4). The majority of microplastics liberated from face masks were clear medium-sized
416 polypropylene fibers from nonwoven materials (Chen et al. 2021). Due to the wide distribution of
417 microplastics likely to be exposed to heavy metals and organic pollutants in the natural
418 environment, microplastics exhibit substantial potential for adsorption of harmful substances.
419 Hydrophobic chemicals are adsorbed from environmental pollution onto the surface of the plastics
420 (Ye et al. 2020). Some of those intentional chemical additives in plastics with toxic and endocrine-
421 disrupting properties might be present at levels of 1-500 g kg⁻¹ (Gallo et al. 2018). The estimated
422 cumulative plastics marine debris can reach 250 million metric tons by 2025 (Jambeck et al. 2015).

423

424 **Impacts of COVID-19 on water and soil ecosystems**

425 Since the outbreak of the pandemic, non-steroidal anti-inflammatory prescription drugs (NSAIDs)
426 are the most widely used pharmaceuticals in the globe, including nonselective cyclooxygenase
427 (COX) inhibitors such as ibuprofen, naproxen, aspirin (acetylsalicylate), and diclofenac (WHO,
428 2020). Furosemide, a medication used for treating fluid retention, has been suggested for COVID-
429 19 patients as a possible therapeutic disease through a primarily anti-inflammatory action
430 mechanism (Brennecke et al. 2020). The ecosystem's components are intertwined with one
431 another. People washing their hands with soap more frequently, government and local government
432 mass disinfection, and the creation of single-use polymers with bisphenol A (BPA) all are destined

433 to have severe effects on water and soil quality (Kim et al. 2021; Sridharan et al. 2021). Several
434 laboratory experiments showed that human exposure to BPA would induce an endocrine disruption
435 in many organ systems (Dodson et al. 2012). Alcohol-containing products spilled in water are
436 dangerous to aquatic life, while spilled alcohol-containing products in the soil can pollute
437 groundwater (Selvaranjan et al. 2021). Figure 5 shows the possible effects of emerging
438 contaminants (e.g., anti-inflammatory chemicals, hydroxychloroquine/chloroquine, BPA,
439 triclocarban, triclosan, etc.) on our soil and water environments, which are further triggered by the
440 negative effects of COVID-19.

441 Triclocarban and triclosan are chemical compounds present in cleaning and washing products
442 where they are used as antimicrobial and fungicides due to their high microbicide spectrum (Ley
443 et al. 2018; Ion et al. 2019). These chemicals produce a barrier at the air-water contact by forming
444 a protective surface coating. Both compounds are considered dangerous substances due to their
445 potentially harmful effects on humans and marine ecosystems, as well as the possibility of creating
446 antibiotic-resistant strains and have been banned in the USA by the Food and Drug Administration
447 (FDA), but still appear on WWTP (Pycke et al. 2014; Sun et al. 2014).

448 Large quantities of effluent wastewater intoxicated with chloroquine and hydroxychloroquine
449 were being discharged into the environmental waters, with potential ecotoxicological impacts on
450 living organisms. According to Kuroda et al. 2021, the predicted environmental concentration
451 (PEC) in raw wastewater for chloroquine, chloroquine metabolite (N-desethylchloroquine), and
452 HCLQ were 857, 171, and 833 ng L⁻¹, respectively. PEC in secondary effluents for chloroquine,
453 chloroquine metabolite (N-desethylchloroquine), and hydroxychloroquine were 320, 135, and 783
454 ng L⁻¹, respectively. Due to the assumed dilution, concentrations in the river waters were lowered
455 by a factor of 10. Subsequently, the PEC for chloroquine, chloroquine metabolite (N-
456 desethylchloroquine), and hydroxychloroquine in river water reduced to 32, 13, and 78.3 ng L⁻¹.

457 Moreover, eco-toxicological assessment for metabolites of both chloroquine and
458 hydroxychloroquine found them to be highly persistent in the aquatic environment. Owing to their
459 strong anti-viral and anti-bacterial properties, chloroquine, and its derivatives bioaccumulate their
460 toxic forms in aquatic organisms. In terms of toxicological impact on the environment, the medium
461 risk was predicted for hydroxychloroquine and N-desethylchloroquine (Kuroda et al. 2021).

462 **Impacts of COVID-19 on aquatic life**

463 Soaps are the earliest detergents known to man. Discharged detergents produce foam in bodies
464 of water. Detergents and other soaps lower the surface tension of water, resulting in foam. Soaps
465 can cut re-aeration by up to 40% (Capoor and Parida 2021). A study found that 120 mg L⁻¹ of soap
466 could reduce algal growth and development. Soaps can harm aquatic vegetation. Plants such as
467 *Ranunculus aquatilis* and *Potamogeton* cannot thrive in detergent concentrations of 2.5 ppm
468 (Ankit et al. 2021). The accumulation of toxic chemicals in the soil resulting from widespread soap
469 use may degrade soil quality. The sudden rise in soapy discharge from every household over a
470 short amount of time could increase the number of contaminants and change the composition of
471 greywater. Domestic waste will damage river water, which will eventually damage lakes and seas.
472 This unwelcoming series of events will become a severe problem (Leal Filho et al. 2021). Studies
473 indicated that pharmaceuticals including naproxen, ibuprofen, diclofenac, and ketoprofen have
474 been detected in the plasma of fish and marine organisms after they are exposed to treated
475 wastewater (Vasquez et al. 2014). In the majority of traditional WWTPs, most of these emerging
476 contaminants are not efficiently removed from wastewater. Effluent discharge and waste disposal
477 on land might have long-term repercussions on our water systems and biota. Water and wastewater
478 treatment systems for the removal of organic pollutants and emerging contaminants are thus
479 necessary to be more effective and easily implemented. During the first stage of the COVID-19
480 pandemic, ibuprofen has been prescribed as an anti-inflammatory medicine to treat the symptoms

481 of COVID-19 disease (De Girolamo et al. 2020; Yousefifard et al. 2020). Low levels of ibuprofen
482 ($\sim 0.01 \mu\text{g L}^{-1}$) have been observed to raise the danger to aquatic living things in the environment
483 and can produce a serious persistent toxic impact on the reproduction of aquatics (Carlsson et al.
484 2006).

485

486 **Impacts on the energy sources**

487 The virus outbreak created several significant issues in the green energy industry, such as supply
488 chain disruptions and tax stock market challenges (Jefferson 2020). Coal is one of the most
489 significant fuels in the global energy market, occurring in up to 40% of electricity generation.
490 Global coal output rises by 2.7% in 2018, with a projected annual production of 8.1 billion tons
491 by 2019 (Rizou et al. 2020). Figure 6 (Ghosh 2020) shows that in 2020 the use of coal has
492 decreased dramatically to 40 K metric tons, from 10–30 days following the new year, compared
493 with 80 K metric tons before the COVID occurrence (Mousazadeh et al. 2021). Three major coal-
494 producing countries, India, China, and Australia were primarily responsible for the growth,
495 accounting for 70% of global production. The coronavirus lockdown is anticipated to increase
496 world output by 0.5% in 2020 (Eroğlu 2020). However, the worldwide coal market is expected to
497 shrink from \$816.5 billion in 2019 to \$722.8 billion in 2020 due to further lockdown and other
498 government restrictions amid the ongoing COVID-19 outbreak (Chakraborty and Maity 2020).
499 Several factors from the COVID-19 pandemic that have seriously disrupted oil demand would
500 only trigger gradual recovery, thus curbing significant oil price increases due to market
501 vulnerabilities for at least three years (Jefferson 2020).

502 The global economic slowdown impact of global lockdown to stop the spread of the Covid-19
503 epidemic is largely to blame for this huge drop in global output. Similarly, global oil consumption
504 was severely impacted, with a 5% drop in the first quarter of 2020 (Atolani et al. 2020). The drop

505 was primarily due to restrictions on transportation and aircraft, which together account for more
506 than 60% of world oil demand. Similarly, lockdown measures have resulted in a large reduction
507 in electricity usage (>20%), with knock-on implications on the energy mix (Boroujeni et al. 2021).
508 After 40 days of lockdown, Italy showed up to a 30% reduction in energy usage. In France,
509 Germany, Spain, India, and the United Kingdom, there are 15, 12, 15, 20, and 16% reductions in
510 power usage, as shown in Figure 7 (IEA 2021; Mousazadeh et al. 2021). Less energy usage resulted
511 in some positive impacts on the environment regarding various air pollutants (Figure 1). Due to
512 less energy usage, reduced emissions of air pollutants such as greenhouse gases as discussed
513 previously (Adams 2020; Chen et al. 2020; Li et al. 2020; Liu et al. 2021) and particulate matter
514 were noted during COVID-19 (Shen et al. 2021a). Reduced emissions of NO₂ in the atmosphere
515 resulted in an improved air quality index around the world (Lian et al. 2020).

516

517

518 **Mitigating the negative impacts of the pandemic**

519 *Hazardous wastes guidelines and policies during the pandemic*

520 Various national and international agencies have given the guidelines to mitigate the negative
521 impacts of pandemic waste. For example, WHO (2020) has advised the workers to wear proper
522 personal protective equipment (PPE) for the safe collection and disposal of pandemic wastes. US-
523 EPA (2020) has focused on recycling and sustainable management of food wastes at public and
524 private institutions during the pandemic. The U.S Occupational Safety and Health Administration
525 (OSHA), 2020 has advised to have strict engineering and administrative controls, safe work
526 practices, and proper PPE during pandemic waste handling. European Commission, 2020 has also
527 stressed on safe handling of pandemic waste generated from households with confirmed cases and
528 the safety of workers involved in pandemic waste collection. In the United Kingdom, waste stream
529 prioritization, expansion in the temporal waste capacity, waste segregation, an adaptation of MSW
530 incinerators to dispose of COVID-19 infectious waste, and communication with residents were
531 recommended (DEFRA, 2020).

532 In India, the safety of workers with proper PPE, the dedicated vehicle involved in pandemic
533 waste collection, quarantine facilities at Common Bio-medical Waste Treatment Facilities, and
534 proper labeling was recommended by CPCB (2020). The Ministry of Ecology and Environment
535 of the People's Republic of China recommended proper packing of pandemic wastes, the
536 establishment of medical waste disposal units, incinerators, furnaces, emergency disposal
537 methods, and standards for pollution control on medical waste treatment and disposal (GB39707-
538 2020). According to Nigeria's Center for Disease Control (NCDC) guidelines, special waste
539 collection bins, daily disposal, and decontamination as per standard protocols were recommended.
540 Despite these recommendations by different agencies, various strategies have been employed to
541 properly the pandemic waste. The details about these strategies are given in the following sections.

542 ***Strategies for the management of solid wastes***

543 Hospitals and health care centers have generated a tremendous amount of infectious and biological
544 waste for sample collection of suspicious COVID-19 patients, testing, treating a significant
545 number of people, and disinfection purposes. Thus, more efficient and cost-effective disposal
546 methods for the disposal of medical wastes are urgently needed during the pandemic and post-
547 pandemic periods. Effective healthcare waste management is entirely based on a well-organized
548 and well-executed management strategy. To create and maintain an effective waste management
549 strategy, a waste management team or panel should be created. An infection control committee,
550 with one individual involved in healthcare waste management in healthcare institutions, should be
551 established in low-income regions (Das et al. 2021; Hantoko et al. 2021). After the safe disposal
552 of waste, the PPE must be removed safely, and the sanitizer used for disinfecting hands after the
553 waste has been disposed of. For safe cleaning off-site or on-site, the soiled PPE must be placed in
554 a sealed bag and should be cleaned with an agent containing 10% lime slurry (WHO, 2020).
555 Several countries have current regulations and protocols for the disposal of the hospital or domestic
556 medical waste. During the COVID-19 epidemic, certain modifications were made to the waste
557 management method (Hantoko et al. 2021). The trends and practices for solid waste management
558 during the COVID-19 pandemic are presented in Figure 8 (ACRPlus 2020).

559

560 ***Thermochemical conversion processes for treating medical wastes***

561 COVID-19 contributed to the worldwide massive increase of medical waste, generated largely by
562 clinics, hospitals, and other healthcare facilities. This poses another hurdle in the management of
563 medical waste, especially in poor nations. Inappropriate medical wastes management might have
564 severe public health problems and major environmental consequences (Das et al. 2021; Dharmaraj
565 et al. 2021). Thermochemical conversions (i.e., pyrolysis, gasification, hydrothermal liquefaction)

566 of solid wastes including medical wastes, offer several advantages through recovering the
567 recourses and energy and stabilizing/destroying the toxins/contaminants and microplastics in the
568 precursors (Mohamed et al. 2017, 2022a). Pyrolysis is generally non-selective, and many products,
569 including biochar, bio-oil, and gases, can be created in a very short time among all other
570 thermochemical conversion processes (Mohamed et al. 2021; Huang et al. 2022; Periyasamy et al.
571 2022). Fast pyrolysis has been explored mainly to produce bio-oil, with a maximum yield of >
572 70%. Furthermore, the method is more environmentally friendly, reliable, and cost-effective, needs
573 almost no landfill capacity, and produces fewer pollutants (Leng et al. 2022; Mohamed et al.
574 2022a, b). The use of catalytic pyrolysis to improve the quality of pyrolysis vapors, especially bio-
575 oil and noncondensable gases, has been widely researched. Natural and synthetic zeolites are
576 among the numerous catalysts available that have been investigated for improving bio-oil quality
577 (Pütün et al. 2006; Sulman et al. 2009; Veses et al. 2015).

578 During the pyrolysis process, the plastic waste itself has an impact on the product composition.
579 The primary components found in the waste of COVID-19 plastics included polypropylene (PP),
580 polyethylene (PE), polystyrene (PS), polyvinyl chloride (PVC), and polyethylene terephthalate
581 (PET). These cross-linked polymers in the plastic wastes do not melt during high-temperature
582 pyrolysis; instead of melting, they produce valuable energy-bearing products including bio-oil-
583 rich hydrocarbons, gases rich in hydrogen gas content, and charcoal (Dharmaraj et al. 2021; Zhou
584 et al. 2021). The generation of lighter and thermodynamically stable hydrocarbons including
585 aromatic hydrocarbons and C₁-C₄ gases was triggered by higher pyrolysis temperatures (Figure 9).
586 The resistance to thermal decomposition of three distinct plastic feedstocks was ranked as follows
587 based on product yields and chemical composition: HDPE, PP, and PP with fillers, with the
588 mineral filler, talc, acting as a catalyst and showing considerable cracking activity (Zhou et al.
589 2021).

590 The pyrolysis treatment utilizing the face masks was presented by the authors as a safe and cost-
591 effective way to dispose of face masks that protect against Covid-19 (Yousef et al. 2021). Pyrolysis
592 of face masks was conducted using TG-FTIR-GC-MS method, and the results revealed that the
593 face mask has high volatiles with high mass loss of 67–96%, which decomposed in three stages
594 between 360 and 500 °C with activation energy ranging from 231 to 281 kJ mole⁻¹. Based on the
595 GC/MS results, the bio-oil produced from the face mask is composed mainly of furan, propene,
596 isopropylcyclobutane, and 2,4-dimethyl- 1-heptene (Yousef et al. 2021).

597

598 *Thermochemical conversion processes for treating sewage sludge*

599 The majority of the traditional WWTPs are not much efficient in removing most of the emerging
600 contaminants and microplastics from wastewater effluent and sludge/biosolids (Mohamed et al.
601 2022a). Therefore, traditional WWTPs increase the possibility of spreading hundreds of
602 contaminants and microplastics/nanoplastics in the aquatic ecosystems and agricultural lands in
603 case of using biosolids as fertilizer (Kimbell et al. 2018; Patel et al. 2019; Rodríguez-Narvaez et
604 al. 2021). Thus, advanced treatment methods for removing organic pollutants and emerging
605 contaminants are needed, which should be more effective and can be easily implemented without
606 creating secondary wastes and/or toxic pollutants. However, some advanced treatment methods
607 could further generate more byproducts with high acute toxicity for marine creatures and humans
608 such as benzoquinone (Reddy et al. 2018).

609 Among several treatment methods, thermochemical conversion processes i.e., pyrolysis,
610 gasification, hydrothermal liquefaction have proved more effective for the destruction and/or
611 stabilization of toxins (Kimbell et al. 2018), viruses, emerging contaminants (Xu et al. 2015;
612 Murtaza et al. 2021; Mehmood et al. 2022), and microplastics (Rodríguez-Narvaez et al. 2021) as
613 shown in Figure 10. Microplastics in the sludge are considered a good source for carbon-based

614 materials that can be utilized as a precursor for the manufacturing of sludge-based adsorbents (e.g.,
615 hydrochar, biochar, activated carbon) with possibilities for environmental usages for treating
616 contaminated sites or wastewater effluents (Xu et al. 2015; Rodríguez-Narvaez et al. 2021).
617 Several studies successfully produced sludge-based adsorbents with very high BET surface area
618 $>1800 \text{ m}^2 \text{ g}^{-1}$ (Xu et al. 2015). Furthermore, during pyrolysis operations, the generated volatiles
619 and gases could be collected, condensed into biofuels, and used as a source of sustainable energy
620 to maximize profits (Figure 10) (Mohamed et al. 2019). Li and coworkers developed a method
621 combining hydrothermal pretreatment, anaerobic digestion, and pyrolysis to fully utilize sewage
622 sludge from WWTPs, and the process was evaluated at the pilot scale. It was found that the heavy
623 metals were stabilized and the process is effective for resources and energy recovery from sewage
624 sludge (Li et al. 2018). However, it is necessary to analyze the potential of releasing original
625 contaminations in the sludge, which could transform, volatilize or decompose to new forms of
626 hazardous substances during the sludge-based adsorbents production process (Xu et al. 2015).

627

628 *Strategies for the management of pharmaceutical and personal care products*

629 Various strategies such as photocatalytic degradation (Kargar et al. 2021), electrochemical
630 advanced oxidation (Tella et al. 2018; Rizvi and Ahammad, 2022), metal-organic framework,
631 and carbon biomasses adsorbents (Rasheed et al. 2020; Gümüş and Gümüş, 2021; Januário et al.
632 2022) and microbial degradation (Lindroos et al. 2019) have been employed for the removal of
633 various pharmaceutical and personal care products in water and soil ecosystem. Previous studies
634 have utilized different sorbent materials (biochar, activated carbon) to remove ibuprofen and
635 naproxen. The corresponding removal efficiencies range from 83 to 99% (Chakraborty et al.
636 2018; Tomul et al. 2020; Kim et al. 2021), with the highest removal observed for biochar,
637 produced from peanut shell (Tomul et al. 2020). A sludge-based biochar was produced from the

638 co-pyrolysis of sewage sludge and bamboo waste and was successfully used to remove 95% of
639 antibiotic residues (e.g., ciprofloxacin) (Li et al. 2020). Sludge-based biochar was also
640 successfully used as a heterogeneous Fenton-like catalyst for the degradation of organic
641 contaminants (Li et al. 2019).

642 Despite there are several studies that report the presence of pharmaceutical and personal care
643 products in effluents, an international framework regarding WWTP effluents guidelines is needed.
644 In fact, just a few countries regulate wastewater discharges. For example, Canada inspects BPA
645 (3.5 µg/L) and triclosan (0.47 µg/L), in Federal Environmental Quality Guidelines, which are
646 recommendations in quantitative or qualitative terms to support federal environmental quality
647 monitoring (FEQGs, 2020); China regulates only BPA (0.01 mg/L) (EPPRC, 2002) as an
648 ecological hazard value. Thus, it is essential to regulate the concentrations of the harmful
649 pharmaceutical and personal care products in the final discharged wastewater effluent before being
650 discharged into receiving the receiving waters.

651

652 **Outlook and future directions**

653 A cross-disciplinary collaborative strategy is urgently required to address the existing
654 environmental concerns created by the indirect detrimental effects of the COVID-19 pandemic.
655 There is a need to lessen reliance on single-use PPE or trash caused by pandemics. Repurposing
656 worn PPE is an excellent strategy to avoid dumping a large volume of pandemic-related garbage
657 in landfills (Dharmaraj et al. 2021). The researchers have already conducted a feasibility study on
658 using shredded face masks combined with recycled concrete as a road surface material. According
659 to the findings, constructing a two-lane 1-kilometer road would necessitate the usage of nearly 3
660 million old face masks, which would otherwise end up in landfills. It was also discovered that
661 reusing the used masks improved the quality of the recycled masks (Somani et al. 2020).

662 To accommodate surplus healthcare waste, mobile treatment and temporary storage strategies
663 may aid sustainable management of healthcare waste without further spreading the virus. Proper
664 healthcare waste management can also help to recycle waste or convert it into valuable products,
665 e.g., energy. Hazardous medical wastes should be treated through thermal treatment such as
666 pyrolysis or gasification to recover more energy with less carbon footprint, and care should be
667 taken because of the possible release of toxic chlorinated gases and other toxic gases.

668 Pollution from the environment is a significant risk factor and helps to increase the prevalence
669 of major chronic diseases. The COVID-19 pandemic determines the need for a constructive
670 approach, in particular the implementation of environment-friendly policies focused on air
671 pollution reduction and the safe disposal of medical hazardous wastes. Further investigation is
672 necessary for the study of the probable presence of coronavirus and other viruses in municipal and
673 drinking water and the development of effective water analytical technologies (La Rosa et al.
674 2020). In addition to health education regarding disinfections and managing municipal solid waste
675 with a scientific basis, clear instructions on the treatment of domestic medical waste are required
676 (Das et al. 2021).

677

678 **Conclusions**

679 Perhaps the COVID-19 pandemic has short-term favorable environmental benefits. However, the
680 environmental issues resulting from this virus outbreak could have long-lasting effects and post
681 challenges for all countries around the world. COVID-19 has also posed negative consequences
682 on the water and soil ecosystem due to the significant usage of antibacterial hand sanitizers,
683 disinfectants, and medications, increasing the accumulation of various toxic emerging
684 contaminants including triclocarban, triclosan, and hydroxychloroquine. Face masks and other
685 plastic-based protection equipment have been found as major sources for the release and

686 accumulation of microplastics in the aquatic/terrestrial ecosystems, posing major threats for
687 aquatics and humans. Proper handling of municipal solid waste, including medical wastes and
688 treated sludge/biosolids, may lead to a viable energy source and thereby will achieve
689 environmental sustainability. Thermochemical conversion processes including pyrolysis and
690 gasification could be utilized effectively for the destruction and/or stabilization of toxins, viruses,
691 emerging contaminants, and microplastics present in the sludge and solid waste, and energy and
692 resources recovery from the precursors. The information presented in this review can be used to
693 assess short and long-term options for mitigating future harmful effects.

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702

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704 **Badr Mohamed:** Conceptualization; Methodology; Data curation; Formal analysis; Writing -
705 original draft; **I M Rizwanul Fattah:** Methodology; Conceptualization, Writing - review &
706 editing; **Balal Yousaf:** Conceptualization; Writing - review & editing; **Selvakumar Periyasamy:**
707 Conceptualization; Writing - review & editing

708 **Data availability**

709 All data analyzed during this study were included in this article.

710 **Compliance with Ethical Standards**

711 **Conflict of interest**

712 The authors declare that he has no known competing financial interests or personal relationships
713 that could have appeared to influence the work reported in this paper.

714 **Ethical approval**

715 This article does not involve any Human Participants and/or Animals.

716 **Informed consent**

717 Not applicable.

718

719 **Consent to participate and publish**

720 All authors participated and approved the final manuscript to be published.

721

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1 **Figures**

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3 countries during lockdown

4 **Fig. 2:** Johns Hopkins University CSSE is tracking the spread of SARS-CoV-2 in near real time
5 with a map-centric dashboard (using ArcGIS Online). Screenshot date: 26 October 2021.

6 **Fig. 3:** Time series suspended particulate matter (SPM) concentrations (2013–2020) estimated
7 for the Vembanad lake, reproduced with permission from Elsevier (Yunus et al., 2020).

8 **Fig. 4:** Abundances and proportions of microplastics in different colors released from the new
9 and used disposable face masks: a) and c) including transparent; b) and d) colored microplastics;
10 types of polymer in different colors of microplastics identified by Raman spectrum (e),
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13 and water environments.

14 **Fig. 6:** Use of coal fuel in China in 2016–2020. The consumption drop for all years during the
15 Chinese new year is explained by holidays with industrial shutdowns (Ghosh, 2020).

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17 demand compared in China, India and Europe (weighted average of France, Germany, Italy,
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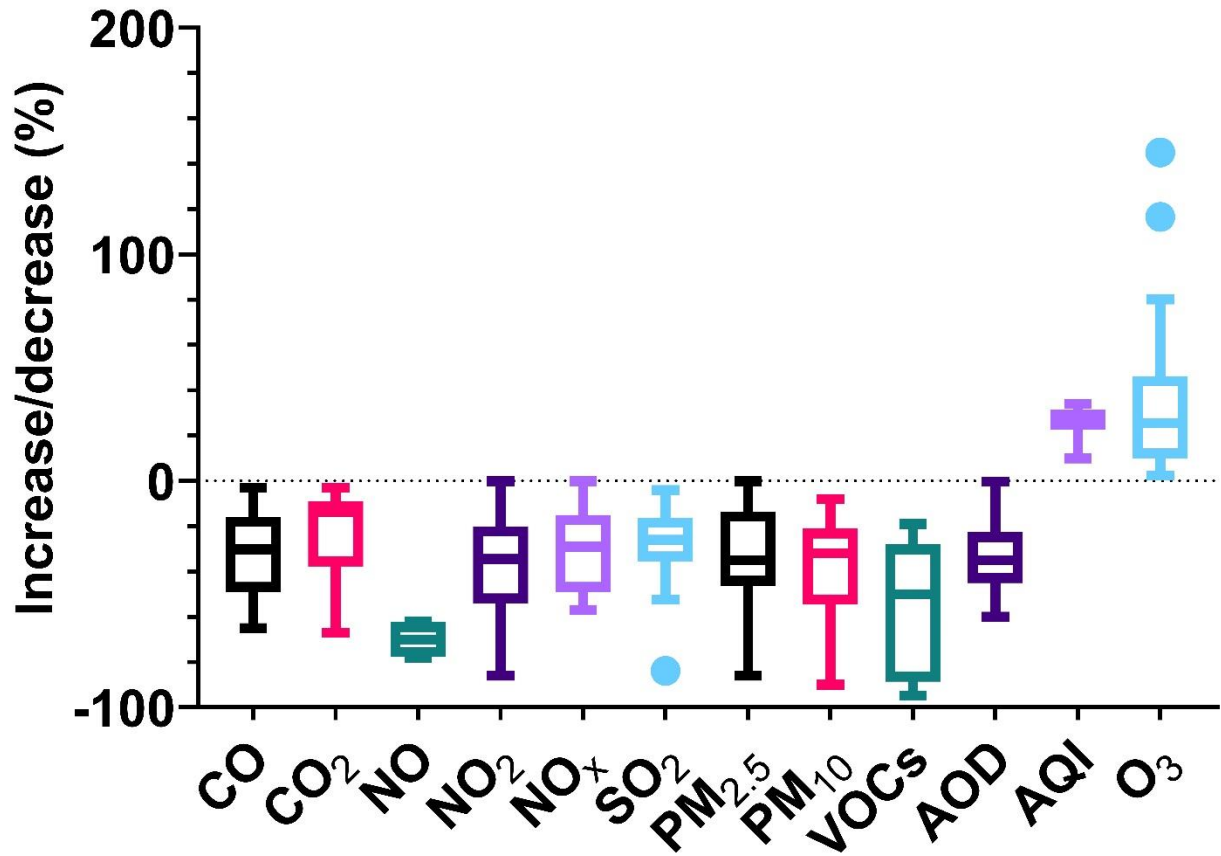
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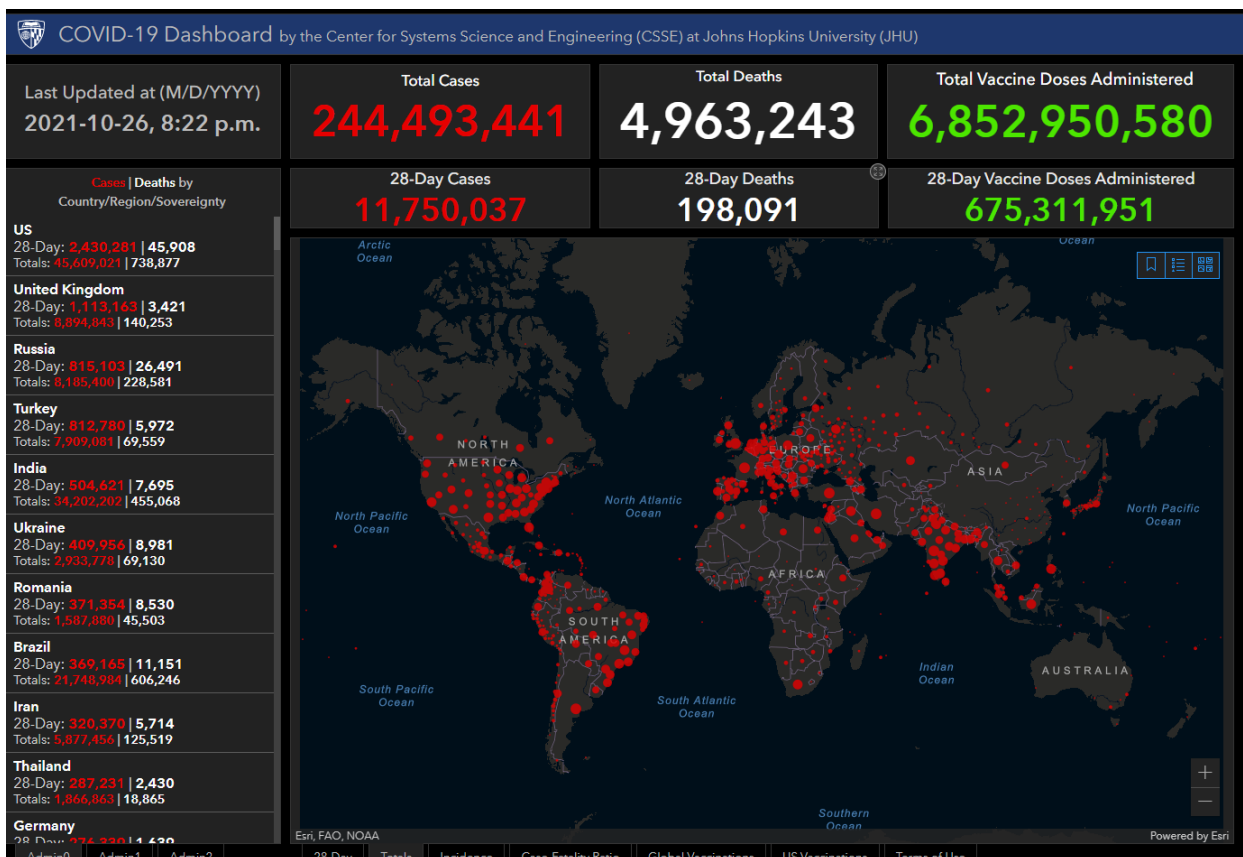
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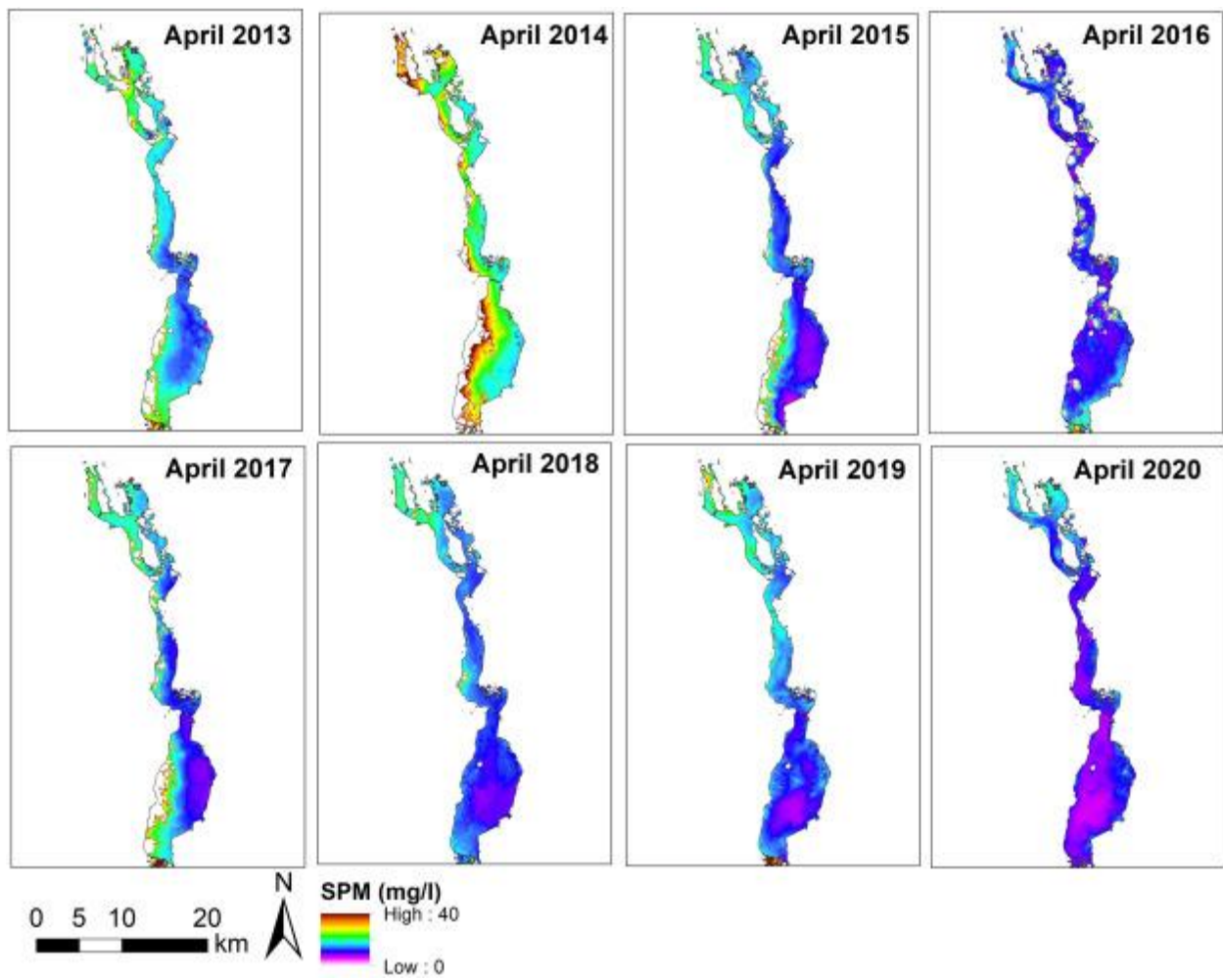
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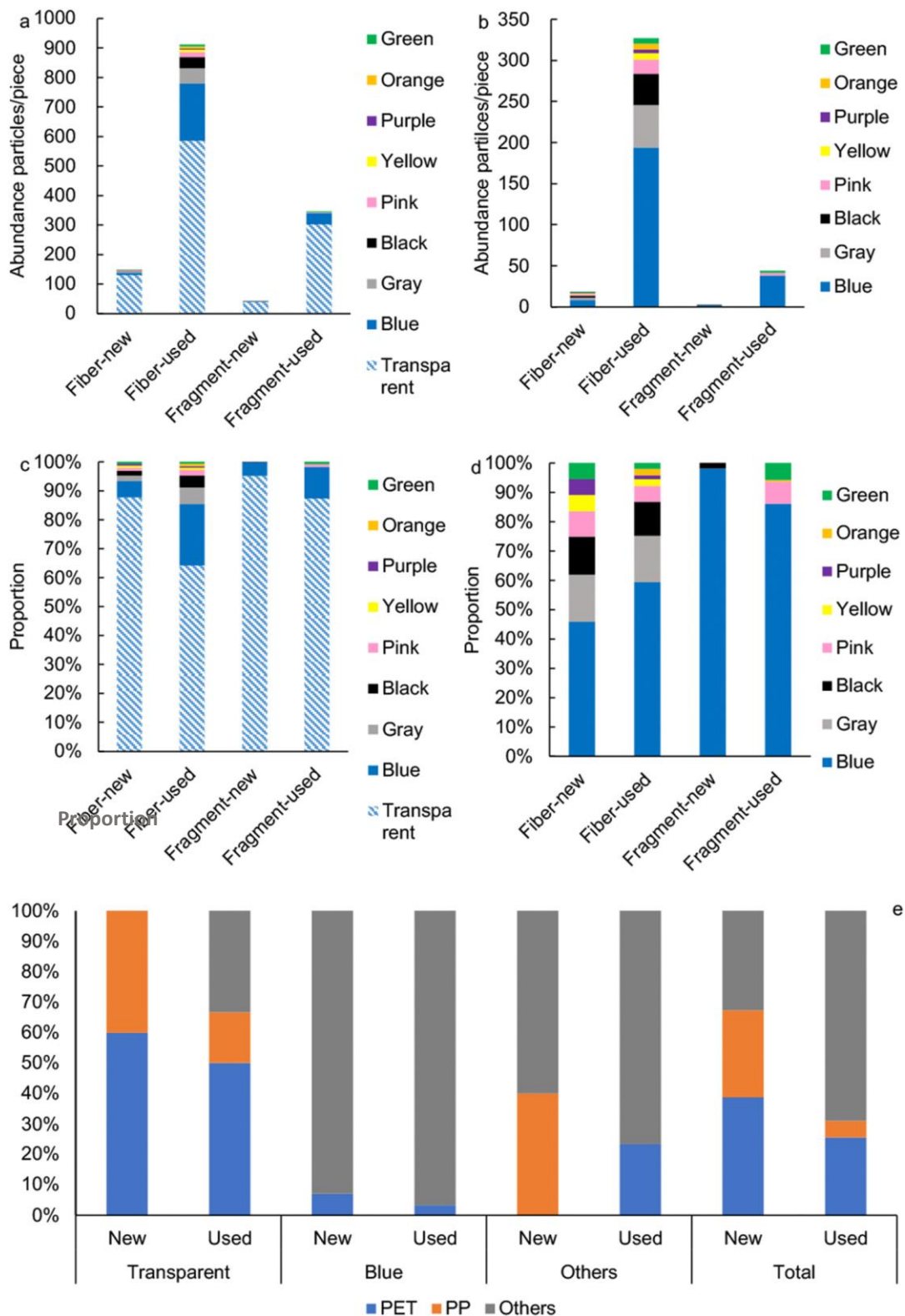
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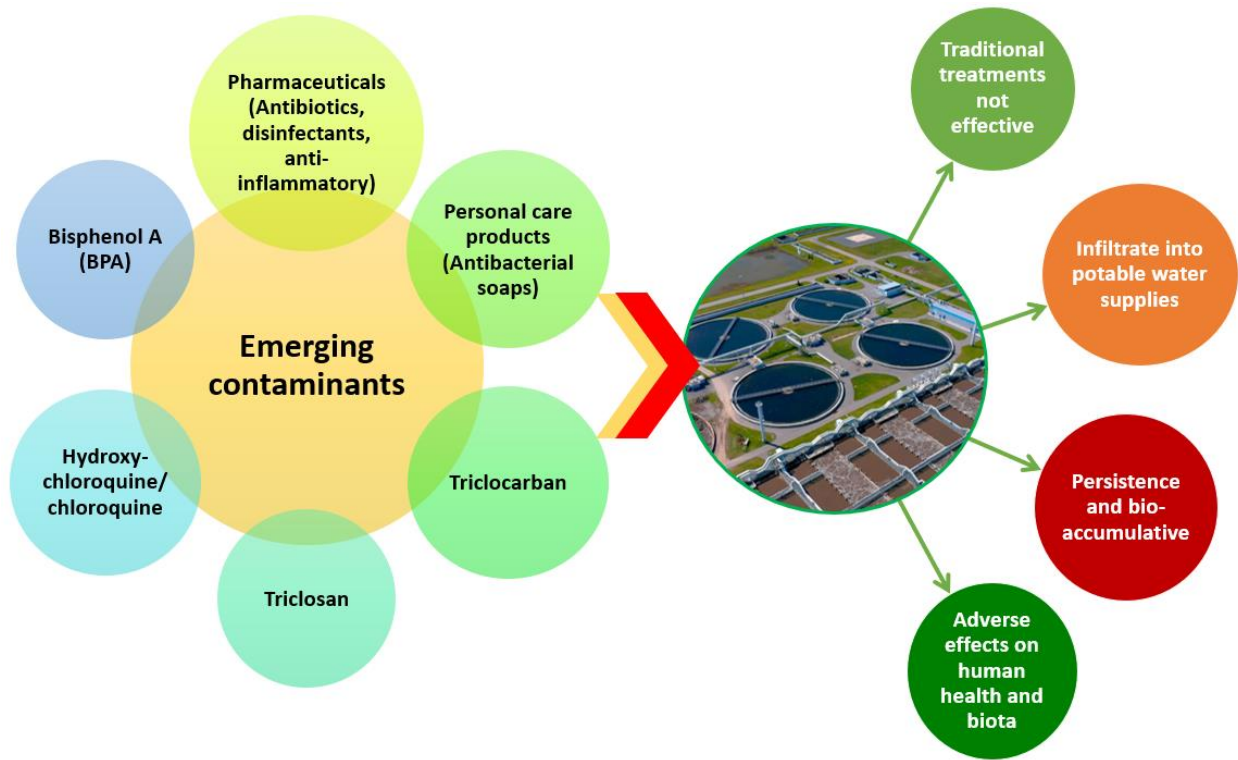
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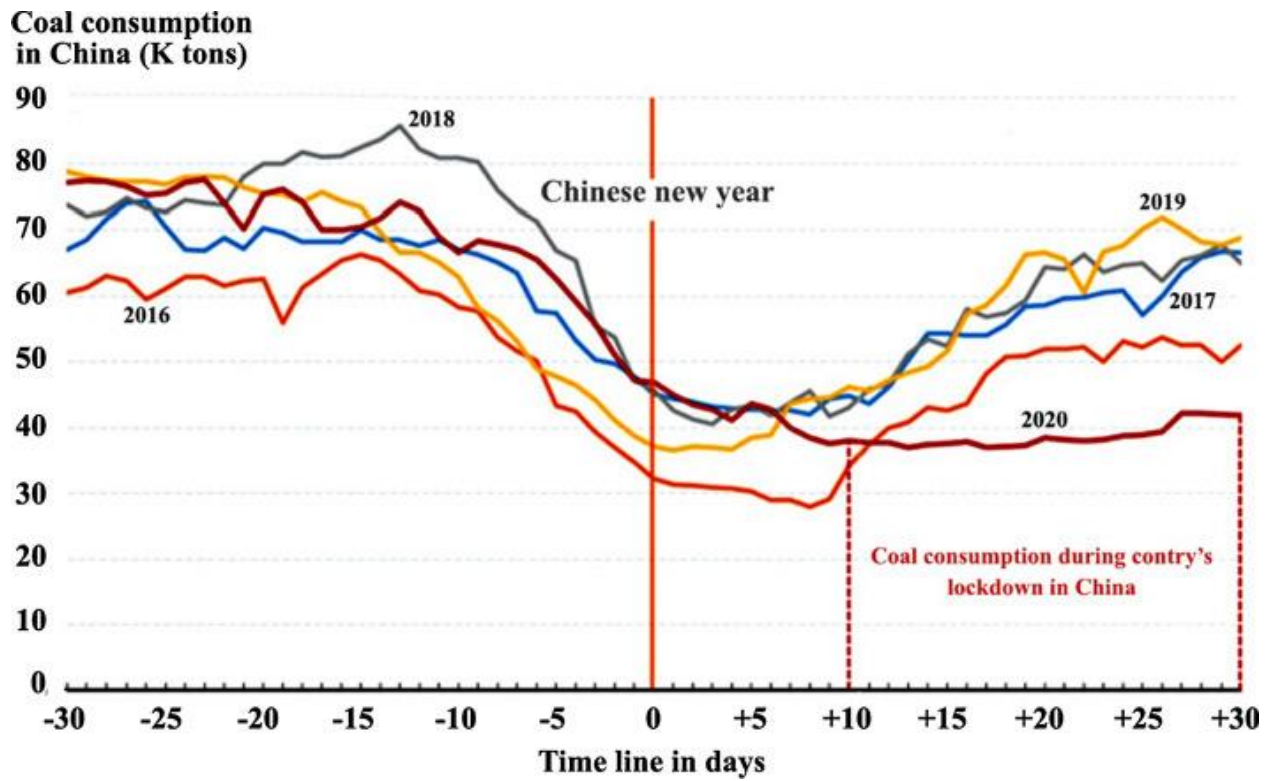
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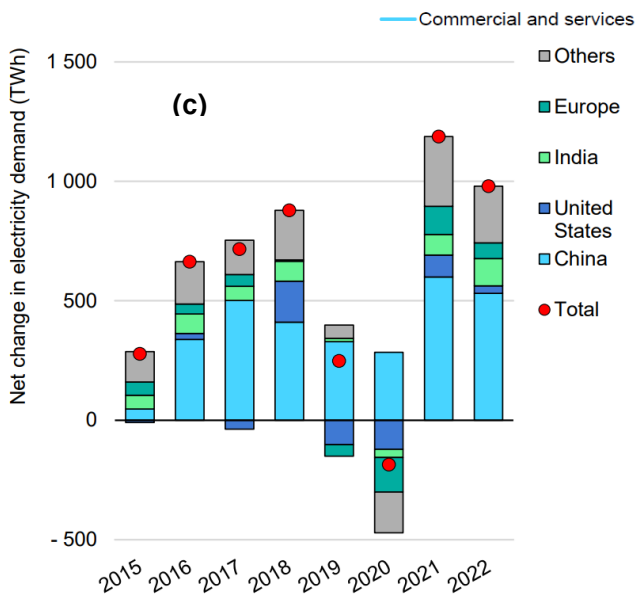
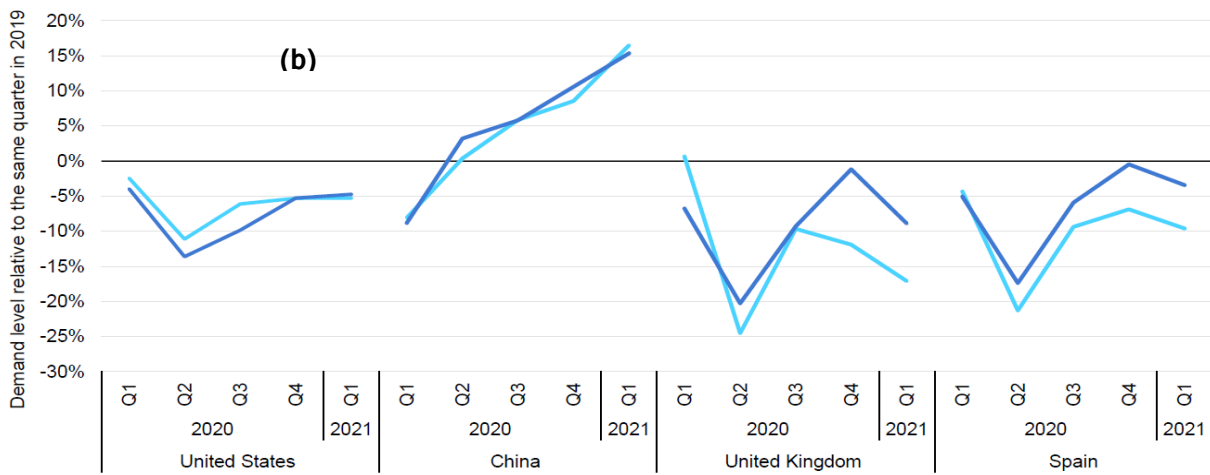
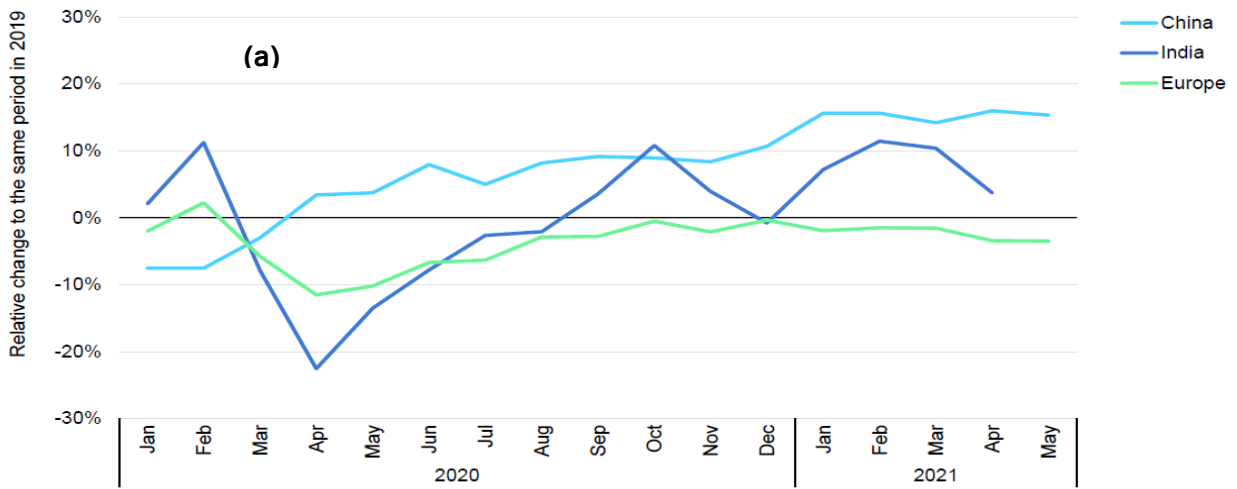
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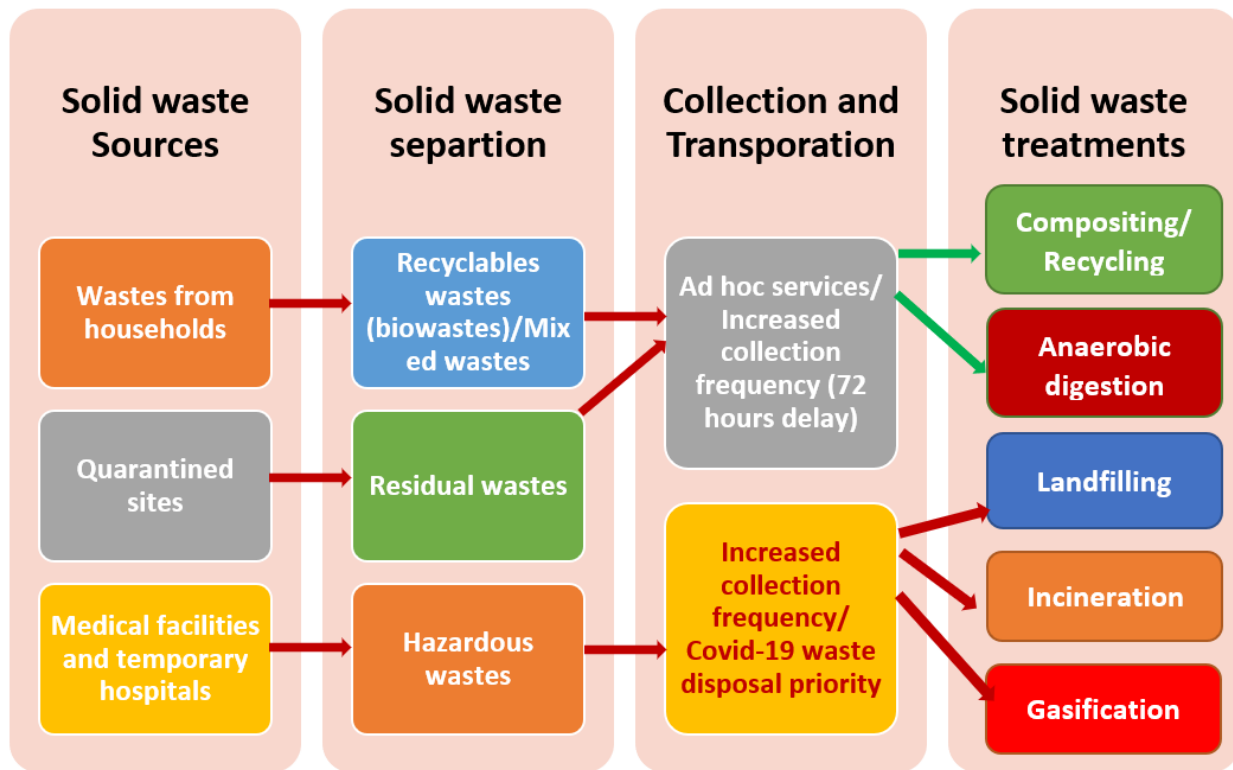
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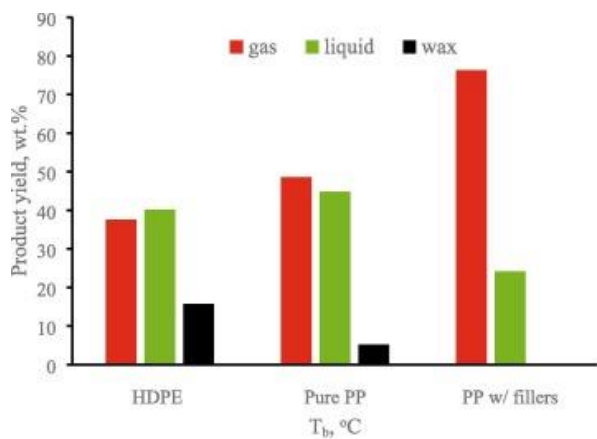
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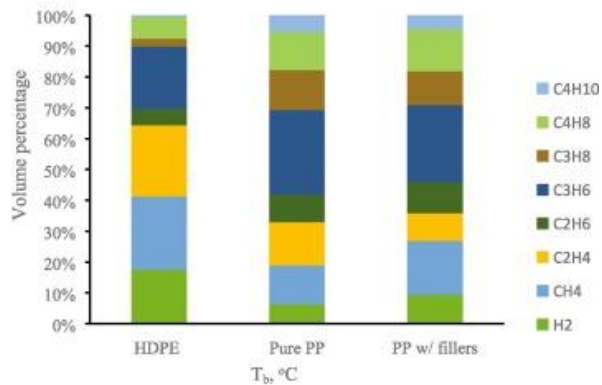
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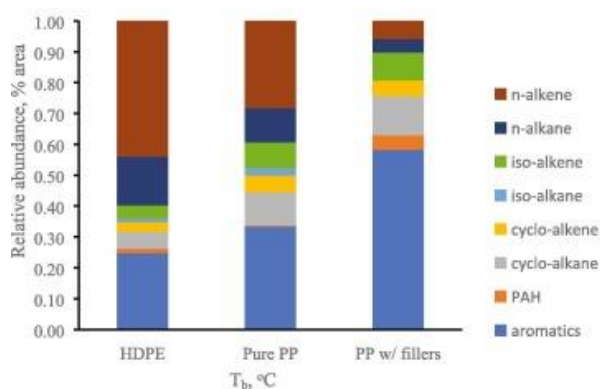
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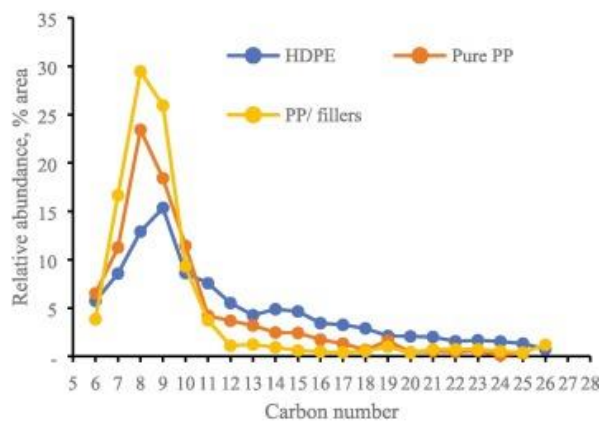
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(b)



(c)



(d)

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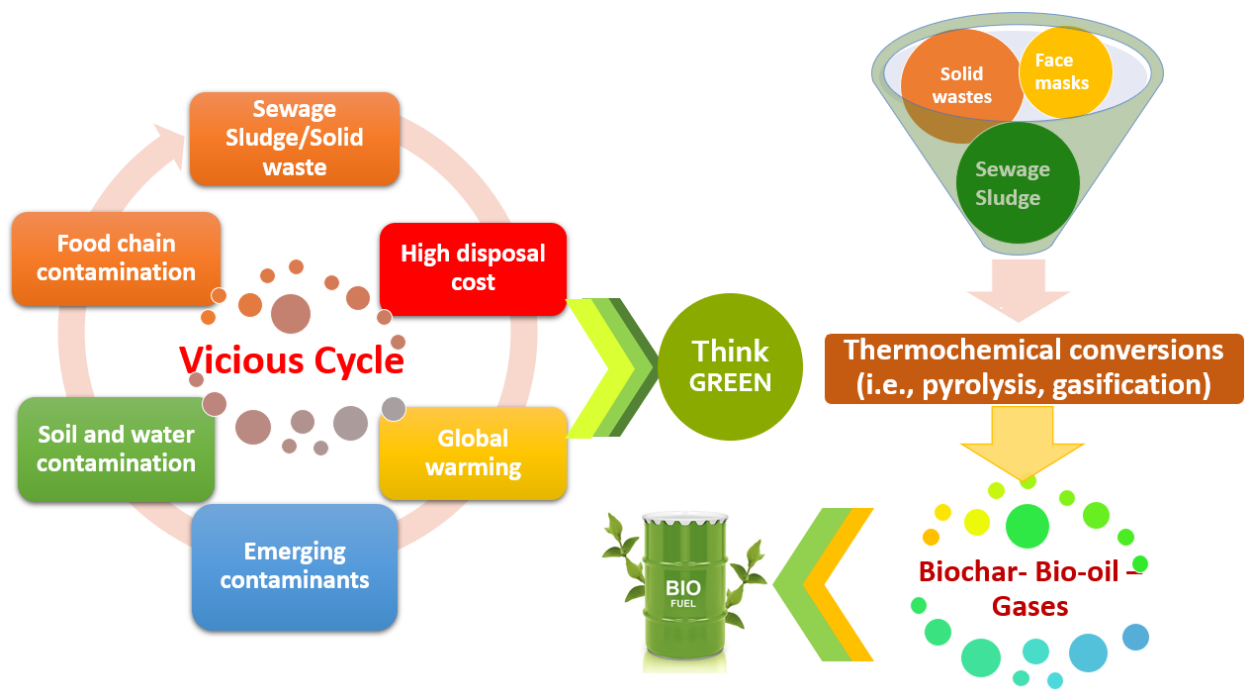
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