

**Cross-generational impact of maternal exposure to low level of PM2.5 on kidney health**

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## Abstract

Inhaled fine and ultrafine particulate matter may affect organs other than the lung, including the kidney. Recent studies have consistently shown the possibility of air pollution in highly polluted countries to be nephrotoxic. However, in countries like Australia, where air quality generally adheres to or remains below the WHO standards, the subtle yet consequential impacts of chronic exposure to seemingly safe levels of traffic PM<sub>2.5</sub>, are a subject of increasing significance. However, how such exposures in the peri-pregnancy period affect kidney health in mothers and the offspring is unclear, which formed the aims of this study. Female Balb/c mice were exposed to PM<sub>2.5</sub> (5 µg/day) delivered nasally for 6 weeks prior to mating, during gestation and lactation (PM group). In a sub-group, PM<sub>2.5</sub> was switched to saline from mating until offspring were weaned to model mothers moving to areas with clean air. Kidneys were analysed in dams and adult offspring at 13 weeks of age. PM<sub>2.5</sub> induced oxidative stress without histological changes in the dam's kidney. However, male PM offspring displayed in-utero underdevelopment, characterised by reduced body weight and kidney-to-body weight at birth compared to control offspring, and lower glomerular numbers, with a marked increase in albuminuria, glomerulosclerosis, inflammation, oxidative stress, and mitochondrial injury. Female PM offspring had delayed postnatal development, lower glomerular numbers, increased glomerulosclerosis and oxidative stress injury markers. Removal of PM<sub>2.5</sub> from conception was overall protective to the offspring. In conclusion, there is no safe level of ambient PM<sub>2.5</sub> for kidney health when exposed in-utero. Maternal PM<sub>2.5</sub> exposure equally impacts the kidney health of male and female offspring.

**Keywords:** fetal programming, traffic-derived PM, oxidative stress, mitochondria, CKD

## 1. Introduction

Air pollution is a pervasive global concern, as over 90% of the world's population lives in areas with polluted air. Epidemiological and experimental studies have well characterised the close association between exposure to high levels of airborne particulate matter (PM) and adverse health outcomes in the respiratory and cardiovascular systems (World Health Organisation, 2018). PM<sub>2.5</sub> is the most hazardous component of air pollutants due to its small diameter. PM<sub>2.5</sub> is 2.5 micrometres or less and is, therefore, capable of crossing tissue barriers in the distal airway and alveoli, through which it can rapidly reach distal organs (Chen et al., 2021).

The kidney is one of the recently identified extrapulmonary targets of PM<sub>2.5</sub>. Although experimental findings for PM<sub>2.5</sub> exposure-related renal injury are scarce, studies have consistently shown nephrotoxic effects of urban air pollution where PM<sub>2.5</sub> levels are several times above the WHO standards (Bowe et al., 2018; Chan et al., 2019b; Nemmar et al., 2009; Nemmar et al., 2016). Chronic exposure to PM has been associated with reduced kidney function among adults in several populations (Bowe et al., 2019; Chan et al., 2018; Ran et al., 2020; Yang et al., 2017; Zhao et al., 2020). As mentioned above, these studies have predominantly focused on regions with highly polluted air, where PM<sub>2.5</sub> concentrations surpass the WHO's recommended standard levels. However, in countries like Australia, where air quality generally adheres to or remains below the WHO guidelines, the subtle yet consequential impacts of chronic exposure to seemingly safe levels of PM<sub>2.5</sub>, especially those derived from vehicles along busy roads, have become a subject of increasing significance.

Notably, our previous investigations using mouse models have unveiled concerning findings regarding the effects of exposure to low levels of PM<sub>2.5</sub>, even within the presumed safe range in both the short and long term (Chan et al., 2019a; Wang et al., 2021). Specifically, we observed lung emphysema-like changes in response to chronic exposure, highlighting the potential implications of seemingly innocuous air pollution levels (Wang et al., 2021). In addition, embryonic and fetal development is most sensitive to changes in the in-utero environment, e.g. environmental toxins, which can delay and impair the development of vital organs and increase the susceptibility to chronic diseases after birth (Chen et al., 2021; Chen et al., 2022a; Chen et al., 2022b; Wang et al., 2021). Indeed, we found an increased risk of asthma and liver dysfunction by in-utero exposure to a "safe level" of PM<sub>2.5</sub> (Wang et al., 2021). This discovery raises a significant research question regarding whether chronic exposure to low-level PM<sub>2.5</sub>, either direct or in-utero, can induce detrimental effects in the other vital organs, such as the kidneys. Given the size of PM<sub>2.5</sub>, systemic absorption is likely, which places fetuses at risk of systemic exposure.

Heavy metals (e.g. lead, cadmium, and arsenic) contained in PM<sub>2.5</sub> have been associated with renal tubular and interstitial damage (Kim, 2017; Möhner, Pohrt and Gellissen, 2017). Transition metals can attach to glycated proteins, enhancing free radical reactions and exacerbating oxidative stress (Shah et al., 2007). Acute exposure to environmental toxins causes direct renal injury to proximal tubules, while persistent or chronic exposure may result in hypertension, interstitial nephritis, and renal fibrosis (Kim, 2017; Navarro-Moreno et al., 2009; Soderland et al., 2010). Short-term exposure to traffic PM induces acute renal failure via oxidative stress (Waly, Ali and Nemmar, 2013). In rats with pre-existing renal injury, a single high dose (1 mg/kg) of diesel exhaust PM aggravated oxidative stress while reducing SOD activity in the renal cortex (Nemmar et al., 2010). Chronic PM<sub>2.5</sub> exposure also increases the risk of CKD (Bowe et al., 2017; Bowe et al., 2018). In particular, exposure to PM<sub>2.5</sub> significantly correlates with the progression to end-stage kidney disease and increased mortality due to renal failure (Bowe et al., 2019; Ran et al., 2020). These studies suggest exposure to ambient PM<sub>2.5</sub> as a possible risk factor for kidney diseases; however, the nephrotoxic effects, especially in the context of prenatal exposure, remain a knowledge gap. A poor intrauterine environment, such as sub-optimal nutritional status and environmental toxins inhaled by pregnant mothers like cigarette smoke and heavily polluted air, correlates with low birth weight (LBW) in the offspring (Rich et al., 2015; Wang et al., 2002; Wang et al., 2018), while individuals born with LBW had a 70% increased risk of CKD (White et al., 2009). We have shown that maternal smoking during

pregnancy caused low nephron number and progressive renal injury in mouse offspring in conjunction with LBW (Nguyen et al., 2015; Stangenberg et al., 2015a). Mitochondria are particularly susceptible to PM<sub>2.5</sub>-induced ROS and oxidative stress (Stangenberg et al., 2015a). Aberrant fetal programming of the kidney could be due to abnormal mitochondrial (mt)DNA (Nguyen et al., 2015). Hence, in-utero PM<sub>2.5</sub> exposure could provoke adverse mitochondrial biogenesis, resulting in impaired mitochondrial functional units (Chen et al., 2021; Nguyen et al., 2015).

Few studies have investigated the effects of maternal PM exposure on the offspring and whether living in an environment with continuous low-level PM<sub>2.5</sub> exposure (akin to being close to a busy road) during pregnancy affects the offspring. This research aims to address this gap by investigating the relationship between maternal exposure to a low level of PM<sub>2.5</sub> and the subsequent renal health of offspring.

## **2. Materials and Methods**

### **2.1 Animal model**

The animal study was approved by the Animal Care and Ethics Committee at the University of Technology Sydney (ETH17-1998) and followed the Australian National Health and Medical Research Council Guide for the care and use of laboratory animals. Female Balb/c mice (8 weeks, Animal Resource Centre, WA, Australia) were divided into 3 groups. Animals were exposed to roadside PM<sub>2.5</sub> (5µg suspended in 40µl saline, delivered nasally (Chan et al., 2019a) or to saline (40µl of saline, delivered nasally) once per day for 6 weeks before mating and continued during gestation and lactation as we have previously published (Wang et al., 2021). The dose was determined following our previous studies (Chan et al., 2019a; Oliver et al., 2024). A subgroup of PM<sub>2.5</sub>-exposed mice was switched to saline during pregnancy and lactation (Pre-exposure group). The litter size was not significantly different among groups, as we published previously (Wang et al., 2021). The dams were harvested when the pups were weaned, and the pups were not subjected to direct PM<sub>2.5</sub> exposure. Birth weight and kidney weight were measured in the newborn. Kidneys from the dams and from offspring at 13 weeks of age (representing adulthood) and spot urine samples from adult offspring were analysed.

### **2.2 ELISA**

Spot urine samples were used to calculate the albumin:creatinine ratio (UACR) as a measure of glomeruli function. A mouse albumin ELISA kit (Crystal Chem, IL, United States) was used to determine the urine concentrations of albumin. Creatinine levels were determined using a colourimetric assay kit for creatinine (Cayman Chemical, MI, United States) as previously described (Larkin et al., 2021).

### **2.3 Histology and immunohistochemistry**

Kidneys from the dams and adult offspring were fixed in neutral buffered formalin (10%), embedded in paraffin and sectioned. The kidney structures were examined using hematoxylin and eosin (H&E). Periodic acid-Schiff (PAS) was used to detect polysaccharides suggestive of fibrotic changes in the kidney section (Hui et al., 2017). Paraffin-embedded sections were incubated in 1% w/v periodic acid (15 minutes), followed by Schiff's reagent (15 minutes), Mayer's haematoxylin (2 minutes), and finally Scott's blue (1 minute), with washes between each incubation. Then, the sections were dehydrated in increasing grades of ethanol (1x 95%, 2x 100%) for 3 minutes each, and xylene (2x 3 minutes) and then cover-slipped. The sections were analysed and quantified with ImageJ (National Institutes of Health, MD, USA).

For IHC staining, formalin-fixed paraffin-embedded sections were de-paraffinised and boiled for 20 min in 10 mM citrate buffer (pH 6.0) for epitope retrieval. Sections were washed in TBST buffer and exposed to 0.3% H<sub>2</sub>O<sub>2</sub> for 5 min to quench endogenous peroxidases, then blocked with Dako protein block (Dako, Carpinteria, CA, USA) for 10 min and incubated with a rabbit polyclonal antibody against 8-OHdG (1:100, BIOSS, Woburn, MA, USA) overnight as we have previously described (Stangenberg et al., 2015b). The tissues were then incubated with polymer secondary anti-rabbit antibodies (Dako Ref K4003), horseradish peroxidase enzyme, and DAB+ (liquid

152 DAB+substrate chromogen system, Dako Ref K3468). The sections were counterstained with  
153 haematoxylin. Negative controls were prepared by replacing the primary antibodies with rabbit IgG.  
154 Quantitation of the positive signals in the images was performed using Image J software (Image J,  
155 NIH, USA).

156 The glomerular number was estimated by counting the developed glomeruli in 8–10 different fields  
157 for the same kidney section and then averaged as we have previously demonstrated (Al-Odat et al.,  
158 2014). One section was used from each kidney (4-6 randomly selected mice from each group). The  
159 glomerular size was measured by assessing the glomerular perimeter using Image J (Image J, NIH,  
160 USA) in 8–10 non-overlapping images for the same kidney section and then averaged. Glomerular  
161 and tubular structure, in addition to glomerular number and size, were additionally assessed by an  
162 independent pathologist in a blinded manner for confirmation.

## 163 164 **2.4 Real-time PCR**

165 Total mRNA was extracted from frozen kidney tissue with TriZol reagent (Life Technologies, CA,  
166 USA), and the first strand cDNA was generated using M-MLV Reverse Transcriptase, RNase H,  
167 Point Mutant Kit (Promega, WI, USA).

168 Target gene expression was quantified with SYBR® primers (TNF  $\alpha$  , Forward-  
169 GGTGCCTATGTCTCAGCCTCTT, Reverse- GCCATAGAAGTATGAGAGGGAG; CD68,  
170 Forward GGCGGTGGAATACAATGTGTCC, Reverse AGCAGGTCAAGGTGAACAGCTG.  
171 Sigma-Aldrich, St. Louis, MO, US) and standardised to housekeeping 18s RNA (Forward  
172 GAATAATGGAATAGGACCGCGG, Reverse GGAAGTACGACGGTATCTGATC. Sigma-  
173 Aldrich). The probes of the target genes and 18s RNA were labelled with FAM® or SYBR®, using  
174 SensiFAST™ SYBR® Hi-ROX Kit (Bioline, NSW 2015). The average of the SHAM group was  
175 assigned the calibrator against which all other results were expressed as fold changes.

## 176 177 **2.5 Western blot**

178 Frozen kidney tissue was homogenised in cell lysis buffer in the Mitochondria Isolation Kit (Thermo  
179 Fisher Scientific. Cytosolic and mitochondrial fractions were separated following the manufacturer's  
180 protocol. Protein concentration was measured by Pierce™ BCA Protein Assay Kit (Thermo Fisher  
181 Scientific) and stored at -80°C for further analysis. The same amount of protein was loaded on SDS-  
182 PAGE electrophoresis gels (8-13%) and electroblotted to Hybond nitrocellulose membranes  
183 (Amersham Pharmacia Biotech). Membranes were incubated with primary antibodies against  $\beta$ -actin  
184 (Santa Cruz Biotechnology), Manganese Superoxide Dismutase (MnSOD) (1:2000, Millipore),  
185 translocase of the outer membrane 20 (1:2000, TOM20; Abcam), COXIV (1:4000, Cell Signalling),  
186 and OXPHOS complex I–V (1:4000, Abcam), followed by washing and incubation with secondary  
187 antibody. After washing the membrane, immunoblots were developed using Clarity ECL Substrates  
188 (Bio-rad) and visualised on ImageQuant LAS 4000 (Fujifilm, Tokyo, Japan). Membranes were  
189 restored by stripping buffer (Thermo Scientific) if necessary. ImageJ (National Institutes of Health)  
190 was used for densitometry, and  $\beta$ -actin was used as the housekeeping protein.

## 191 192 **2.6 Statistical methods**

193 Results are expressed as mean  $\pm$  SEM. Normality was tested. If the data were not normally distributed,  
194 then they were log-transformed. The results of the dam, male and female offspring were analysed  
195 independently, using one-way ANOVA, followed by Turkey post hoc tests (GraphPad Prism 10,  
196 GraphPad, CV, USA).  $P < 0.05$  was considered significant.

## 197 198 **3. Results**

### 199 **3.1 The effects of maternal PM exposure on anthropometric measurements and UACR**

200 Dams exposed to PM showed significant reductions in body weight ( $P < 0.05$ ) and kidney weight ( $P <$   
201  $0.05$ ), which were normalised by PM<sub>2.5</sub> cessation since mating (body weight  $P < 0.01$  PM<sub>2.5</sub> vs Pre-

202 exposure, Table 1). Kidney to body weight ratio was similar among the three groups. Maternal  
203 PM<sub>2.5</sub> exposure prior to mating for 6 weeks, during gestation and lactation induced reductions in body  
204 weight ( $P < 0.05$ ), kidney weight ( $P < 0.01$ ), and kidney/body weight ratio in male offspring at  
205 postnatal day 1. Such changes in kidney weight and kidney/body weight ratio were prevented if  
206 PM<sub>2.5</sub> exposure had been removed just before gestation ( $P < 0.05$ , Table 1). In sharp contrast to male  
207 offspring, female offspring showed no difference in body weight, kidney weight or kidney/body  
208 weight ratio at day 1; however, their body weight and kidney weight were significantly lower than  
209 the SHAM in adulthood ( $P < 0.05$ , Table 1). At week 13, male offspring showed no significant  
210 differences in anthropometric measurements. However, there was a significant increase in UACR in  
211 the PM group ( $P < 0.05$ ), which was significantly attenuated in the Pre-exposure group ( $P < 0.01$ ). In  
212 female offspring, there was also a trend of increased UACR, although this did not reach significance  
213 (Table 1).

### 214 215 **3.2 The effects of maternal PM exposure on glomerular histology**

216 Dams exposed to PM<sub>2.5</sub> showed similar number and size of glomeruli (Figure 1A and 1B). Conversely,  
217 glomerular numbers in both male and female offspring were significantly reduced by maternal  
218 PM<sub>2.5</sub> exposure (both sex  $P < 0.05$  vs SHAM), although kidney weights were normal at 13 weeks  
219 (Figure 1C and 1E), which were not reversed by the removal of PM<sub>2.5</sub> (male  $P < 0.05$  Pre-exposure vs  
220 SHAM). Glomerular size was not different among any offspring. With reduced glomerular number,  
221 maternal PM<sub>2.5</sub> exposure also led to significant increases in markers of glomerulosclerosis in both  
222 male and female offspring ( $P = 0.07$  male PM<sub>2.5</sub> vs SHAM;  $P < 0.05$  female PM<sub>2.5</sub> vs SHAM, Figure 2).  
223 The removal of PM<sub>2.5</sub> exposure at conception reversed glomerulosclerosis to the control levels in both  
224 male and female offspring (Figure 2).

### 225 226 **3.3. The effects of maternal PM exposure on kidney inflammation**

227 PM<sub>2.5</sub> exposure in the dams did not significantly increase inflammatory markers, including TNF $\alpha$  and  
228 macrophage marker CD68 mRNA expression (Figure 3A, D). At week 13, PM male offspring showed  
229 significantly increased TNF $\alpha$  and CD68 expression compared to the SHAM offspring (both  $P < 0.05$   
230 vs SHAM, Figure 3B, E). Maternal PM removal reduced TNF $\alpha$  expression and CD68 (TNF $\alpha$   $P < 0.05$   
231 Pre-exposure vs PM<sub>2.5</sub>, Figure 3B, E). In PM female offspring, both TNF $\alpha$  and CD68 expression was  
232 suppressed (CD68  $P < 0.05$  vs SHAM, Figure 3C, F). In the female Pre-exposure group, both TNF $\alpha$   
233 and CD68 levels were reversed to surpass the SHAM group levels (CD68  $P < 0.01$  vs PM<sub>2.5</sub>, Figure  
234 3C, F).

### 235 236 **3.4. The effects of maternal PM exposure on kidney oxidative stress markers and mitochondrial functional markers**

237 To examine oxidative stress in the kidney, we measured the protein levels of MnSOD in both the  
238 cytosolic and mitochondrial fractions. As shown in Figure 4, in the dams, MnSOD was significantly  
239 reduced by PM<sub>2.5</sub> exposure in the cytosolic and mitochondrial fractions in the kidneys (both  $P < 0.05$   
240 vs SHAM, Figure 4A, B); while PM removal only normalised cytosolic MnSOD level ( $P < 0.05$  vs  
241 PM<sub>2.5</sub>, Figure 4A). In line with this, the oxidative stress marker 8-OHdG was also increased in both  
242 PM and Pre-exposure dams ( $P < 0.05$ , Pre-exposure vs SHAM, Figure 4C). In the PM male offspring,  
243 MnSOD was reduced in the cytosolic fraction ( $P < 0.01$  vs SHAM), but increased in the mitochondrial  
244 fraction ( $P < 0.01$  vs SHAM), which was not affected by maternal PM removal at conception  
245 ( $P < 0.05$  and  $P < 0.01$  PM<sub>2.5</sub> vs SHAM for cytosolic and mitochondrial fractions, respectively, Figure  
246 4A, B). The level of 8-OHdG was also increased in male PM offspring, which was normalised in the  
247 Pre-exposure group ( $P < 0.05$  PM<sub>2.5</sub> vs SHAM, Figure 4C). In the PM female offspring, MnSOD levels  
248 in both cytosolic and mitochondrial fractions were similar among groups (Figure 4A, B). However,  
249 8-OHdG level was significantly increased in the female PM offspring and normalised in the Pre-  
250 exposure group ( $P < 0.05$  PM<sub>2.5</sub> vs SHAM,  $P < 0.01$  Pre-exposure vs PM<sub>2.5</sub>, Figure 4C).

251  
252 Mitochondria are highly susceptible to oxidative stress. Mitochondrial dysfunction is one of the key  
253 mechanisms in developmental programming. In the dams, kidney COXIV (Cytochrome c oxidase)  
254 and substrate transporter TOM20 were not affected by PM<sub>2.5</sub> exposure for any duration (Figure 5).  
255

However, COXIV was reduced in both male and female PM offspring (male  $P < 0.05$  PM<sub>2.5</sub> vs SHAM; female  $P = 0.25$  PM<sub>2.5</sub> vs SHAM), and only COXIV in female kidneys was reversed in the Pre-exposure group ( $P < 0.01$  vs SHAM,  $P < 0.001$  vs PM<sub>2.5</sub> Figure 5). TOM20 level was only reduced in male PM mice ( $P < 0.05$  vs SHAM), which was not affected by maternal PM removal since conception ( $P = 0.05$  vs SHAM, Figure 5).

To further examine mitochondrial functional units, we assessed the protein of mitochondrial oxidative phosphorylation complexes (OXPHOS). In the dams, PM exposure led to a significant reduction of complex I level ( $P < 0.05$  vs SHAM, Figure 6A), without any effects on the other four complexes. In male PM offspring, only complex IV was significantly reduced ( $P < 0.05$  vs SHAM, Figure 6B), which was not normalised in the Pre-exposure males; However, complex I level was doubled in the Pre-exposure group compared with the SHAM group ( $P < 0.05$ , Figure 6B). In female offspring, OXPHOS complexes were all at similar levels among experimental groups (Figure 6C).

#### 4. Discussion

Kidneys are vulnerable to both acute and chronic injury upon exposure to environmental toxins (Feng et al., 2023; Kim, 2017; Navarro-Moreno et al., 2009; Soderland et al., 2010). Low levels of PM have not previously been considered a concern with respect to kidney health. Here, we observed the impact of low-level PM<sub>2.5</sub> exposure on oxidative stress in the dams' kidneys without changes in inflammatory and histological markers. Strikingly, such a "mild" environmental factor has shown long-lasting adverse impacts on the offspring's kidneys. We also observed maternal PM<sub>2.5</sub> exposure induced in-utero underdevelopment in the male offspring but not female offspring. Interestingly, reduced glomerular numbers were observed in both male and female PM offspring, along with induced oxidative stress and mitochondrial injury in adulthood.

The advantage of using an animal model is that we can exclude the potential confounder of postnatal direct PM<sub>2.5</sub> exposure in the offspring and the genetic variations observed in the general population. As such, any effects observed in the offspring are mostly attributed to maternal effects. Dams exposed to PM<sub>2.5</sub> only during the preconception period (pre-exposure group) showed significant DNA oxidation in the kidneys, reflected by the increased 8-OHdG levels. There was also a trend of an increase in renal 8-OHdG levels in the dams when they were exposed to PM<sub>2.5</sub> during the gestation period. This was associated with a significant reduction in mitochondrial levels of the antioxidant marker MnSOD, suggesting increased renal oxidative stress. PM<sub>2.5</sub> exposure to dams during the preconception or the gestation period significantly reduced mitochondrial OXPHOS complex I levels. OXPHOS has a critical role in maintaining cellular homeostasis and is also known to support differentiation processes during embryonic development (Fernández-Vizarra et al., 2022). Although OXPHOS complex I was significantly regulated, other complexes, as well as the levels of COX4 and TOM20, were not altered in the dams, suggesting that mitochondrial function was not compromised. Although oxidative stress was increased in the dams, PM<sub>2.5</sub> exposure did not induce renal inflammation and histological changes. This can potentially be due to the placental ability to synthesise melatonin, which has antioxidant and anti-inflammatory effects, necessary to ensure a stable environment for both the mother and foetus (Joseph et al., 2024). The dams may also have other adaptive mechanisms and the mature kidneys are less vulnerable to oxidative stress than youngsters.

Maternal PM<sub>2.5</sub> exposure significantly reduced the offspring's birth weight in the male offspring only. A similar effect was observed in male offspring exposed to maternal cigarette smoke during gestation (Sukjamnong et al., 2017). Kidney-to-body weight was also significantly decreased in the male offspring at birth following exposure to PM<sub>2.5</sub> during gestation, and this was associated with a significant reduction of glomerular number and albuminuria. Removing PM<sub>2.5</sub> exposure during gestation prevented the development of albuminuria in male offspring, suggesting a protective effect on kidneys.

309 Previous studies have established sex bias in disease susceptibility, with females less likely to develop  
310 certain diseases in contrast to males. This has often been explained by the anti-inflammatory effects  
311 of estrogen as a defence mechanism (Chan et al., 2016a). In the kidney, epidemiological findings also  
312 support the male sex as a risk factor for the development of kidney diseases in adulthood (Li et al.,  
313 2008). A study showed significantly higher levels of renal profibrotic markers in male offspring in  
314 response to maternal high fat diet consumption (Nguyen et al., 2017). Our own study on maternal  
315 cigarette smoke exposure also showed a male-prominent risk of renal disorders in the offspring (Al-  
316 Odat et al., 2014; Chan et al., 2017; Chan et al., 2016a; Chan et al., 2016b). However, maternal  
317 exposure to a low level of PM<sub>2.5</sub> before pregnancy or during gestation seems to affect both sexes,  
318 although the changes in renal pathology and renal dysfunction are different between male and female  
319 offspring. Both male and female PM offspring displayed smaller glomerular numbers in adulthood;  
320 however, only males had in-utero renal underdevelopment, demonstrated by the reduced weight and  
321 kidney-to-body weight at birth. Normally, an increased glomerular perimeter is expected when the  
322 glomerular number is reduced to compensate for the filtration. However, this was not observed in PM  
323 offspring, suggesting impaired renal development without sex discrimination.

324  
325 Although the urinary albumin to creatinine ratio was not statistically significant in females, the  
326 physiological impact can be significant. This can result from glomerular endothelial dysfunction  
327 (Sato, 2012; Strutz, 2009), where proteins are filtered out and reabsorbed in the proximal tubules,  
328 causing tubular toxicity (Strutz, 2009). In animal models of high levels of PM exposure, sub-chronic  
329 exposure can impair renal function and structure, such as loss of glomerular integrity, glomerular  
330 atrophy, loss of epithelial cells, increased Bowman's space, oedema, and tubular dilation and  
331 vacuolation (Al Suleimani et al., 2017; Wardoyo, Juswono and Noor, 2018). This can also cause  
332 altered renal haemodynamics, haematuria, and albuminuria (Tavera Busso et al., 2018); while long-  
333 term PM exposure leads to increased renal inflammation, oxidative stress and DNA damage (Nemmar  
334 et al., 2016). PM and related toxicity can induce the release of cytokines and chemokines, contributing  
335 to the influx of mononuclear cells, e.g. macrophages (Strutz, 2009), which produce profibrotic  
336 cytokines that promote the accumulation of interstitial fibrosis (Black, Lever and Agarwal, 2019;  
337 Strutz, 2009). In the long term, this can lead to fibrotic scarring and irreversible chronic kidney  
338 disease (Chen et al., 2022a). However, the inflammatory response was only significant in the male  
339 offspring. Previous studies have demonstrated that the epigenetic effect of maternal cigarette smoking  
340 during gestation is more pronounced in males compared to females. Males are also more susceptible  
341 to environmental toxins and more vulnerable to abnormal methylation than females (Murphy et al.,  
342 2012). Such differences, as well as differences in early adaptation in males and females, can be  
343 responsible for the differences in inflammatory response observed in males and females in our study.  
344 Sexual dimorphism in adulthood is explained by either the effect of sex hormones, e.g. the potent  
345 anti-inflammatory effects of oestrogen in females, or epigenetic regulation that determines stronger  
346 anti-infection capacity of female immune cells (Camporez et al., 2019; Gal-Oz et al., 2019; Shepherd  
347 et al., 2021).

348 The composition of PM<sub>2.5</sub> includes different types and amounts of heavy metals such as iron (which  
349 is mostly abundant), copper, potassium, calcium, zinc, nickel, sodium, manganese, magnesium,  
350 chromium and cadmium (Kim et al., 2010). Previous studies have demonstrated the role of fine heavy  
351 metal particulates in inducing oxidative stress, DNA damage and inflammatory responses (Karlsson et  
352 al., 2006; Karlsson, Nilsson and Möller, 2005; Seaton et al., 2005). Here, we showed increased  
353 oxidative stress / oxidative damage in both male and female offspring.

354 PM<sub>2.5</sub> has been shown to demonstrate a strong oxidative potency that reaches the fetal circulation  
355 (Bové et al., 2019; Leni, Kunzi and Geiser, 2020; Li, Xia and Nel, 2008). Intracellular ROS are by-  
356 products of ATP synthesis (Nguyen et al., 2015), which are normally scavenged by the endogenous  
357 antioxidant system, such as MnSOD, to maintain a balance between antioxidants and prooxidants  
358 (Che et al., 2014; Daenen et al., 2019). MnSOD is one of the key antioxidative enzymes involved in  
359 oxidative stress in CKD (Nguyen et al., AJP, 2015). Male PM offspring displayed reduced cytosolic  
360 MnSOD but increased mitochondrial MnSOD and oxidative stress injury marker 8-OHdG, suggesting

361 an overall reduction in antioxidant capacity and an adaptive/stress response in the mitochondria. The  
362 Tom20 and OXPHOS complex IV were also reduced in male kidneys, suggestive of reduced substrate  
363 transportation into mitochondria for ATP synthesis. This may explain the increased mitochondrial  
364 MnSOD that normally scavenges ROS that are generated during ATP production mediated by  
365 COXIV that was also reduced in male offspring. At the same time, reduced cytosolic MnSOD may  
366 suggest oxidative stress induced by unknown mechanisms, such as protein misfolding observed in  
367 other organs (Onoda et al., 2020), which is beyond the scope of the current study and requires further  
368 investigation. The fact that increased 8-OHdG was observed in female PM offspring despite  
369 unchanged levels of MnSOD suggests that other elements of the antioxidant defence may be impaired.  
370 Further studies are needed to confirm this hypothesis. Moreover, increased proinflammatory  
371 cytokines and oxidative stress are critical contributors to renal diseases (Mihai et al., 2018).  
372 Therefore, maternal PM<sub>2.5</sub> exposure-induced inflammation and oxidative stress may also be  
373 accountable for the profibrotic changes in the offspring's kidney, reflected by PAS staining.

374 The mitochondrion is the powerhouse of the cell, the principal site for ATP synthesis via the  
375 OXPHOS complexes, ensuring sufficient vitality to cells and tissues throughout the body. Prolonged  
376 PM<sub>2.5</sub> exposure-associated ROS overproduction can cause mitochondrial depolarisation (Yang et al.,  
377 2018). Similar to MnSOD, other mitochondrial markers, COXIV and TOM20, were most affected in  
378 the male offspring. These two proteins were still suppressed in the male pre-exposure groups,  
379 suggesting permanent impacts of maternal PM<sub>2.5</sub> exposure on mitochondrial functional units. Such  
380 an effect may affect cellular energy suppliers in the offspring, as discussed above. Interestingly,  
381 COXIV expression was markedly upregulated in the female pre-exposure group, likely due to an  
382 adaptive response. It is unclear if such adaptation would affect the cellular function when the  
383 offspring age. As such, it would be interesting to follow these offspring to aged stages. Overall,  
384 mitochondrial OXPHOS complexes in male and female offspring, were relatively stable among the  
385 three groups.

386 Although PM<sub>2.5</sub> exposure during the pre-gestation period did not improve MnSOD levels in the males  
387 and induced macrophage infiltration in the females, it mitigated DNA oxidation, kidney function and  
388 pathology in both offspring. This study hence demonstrated that continuous exposure to low level of  
389 PM<sub>2.5</sub> from the pre-gestation period and during gestation has detrimental effects on the offspring's  
390 kidney but this can be reversed if PM<sub>2.5</sub> exposure ceased during gestation. Metals present in polluted  
391 air or PM can cross the placental membrane and directly affect the foetus (Liu et al., 2021; Reichrtová,  
392 Dorociak and Palkovicová, 1998), which might explain the reason for the adverse effect of low  
393 PM<sub>2.5</sub> levels. Additional studies are required to confirm this.

394  
395  
396

## 397 **Conclusion**

398 Low levels of PM<sub>2.5</sub> exposure had a marginal impact on the dams' kidney health, albeit with some  
399 increase in inflammatory markers. However, such a "mild" environmental factor exerted long-lasting  
400 adverse impacts on the kidney histological integrity in both male and female offspring if exposure  
401 occurs in their dams.

402

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408

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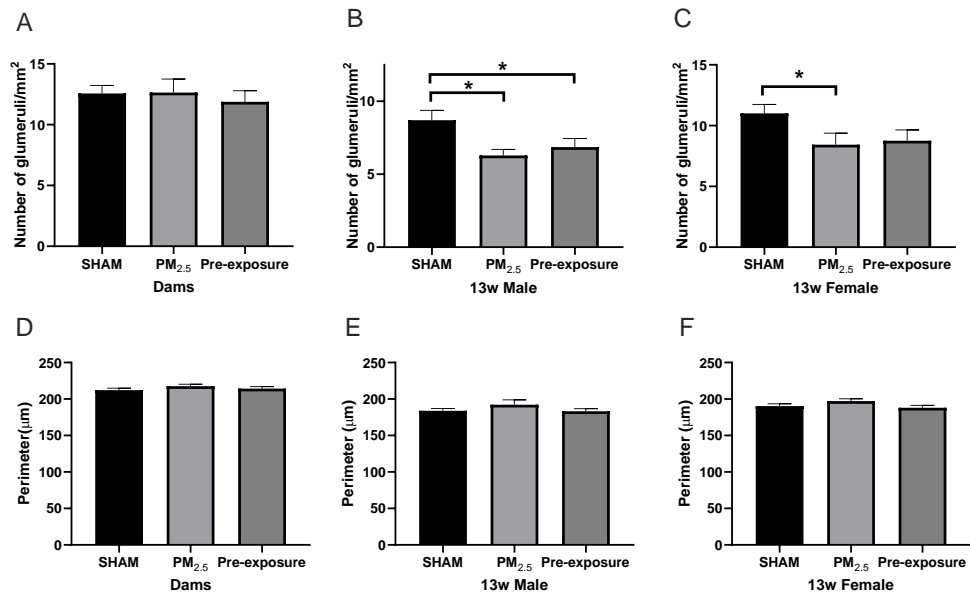
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547

548 Table 1. Anthropometric parameters in the dams and offspring of both sexes

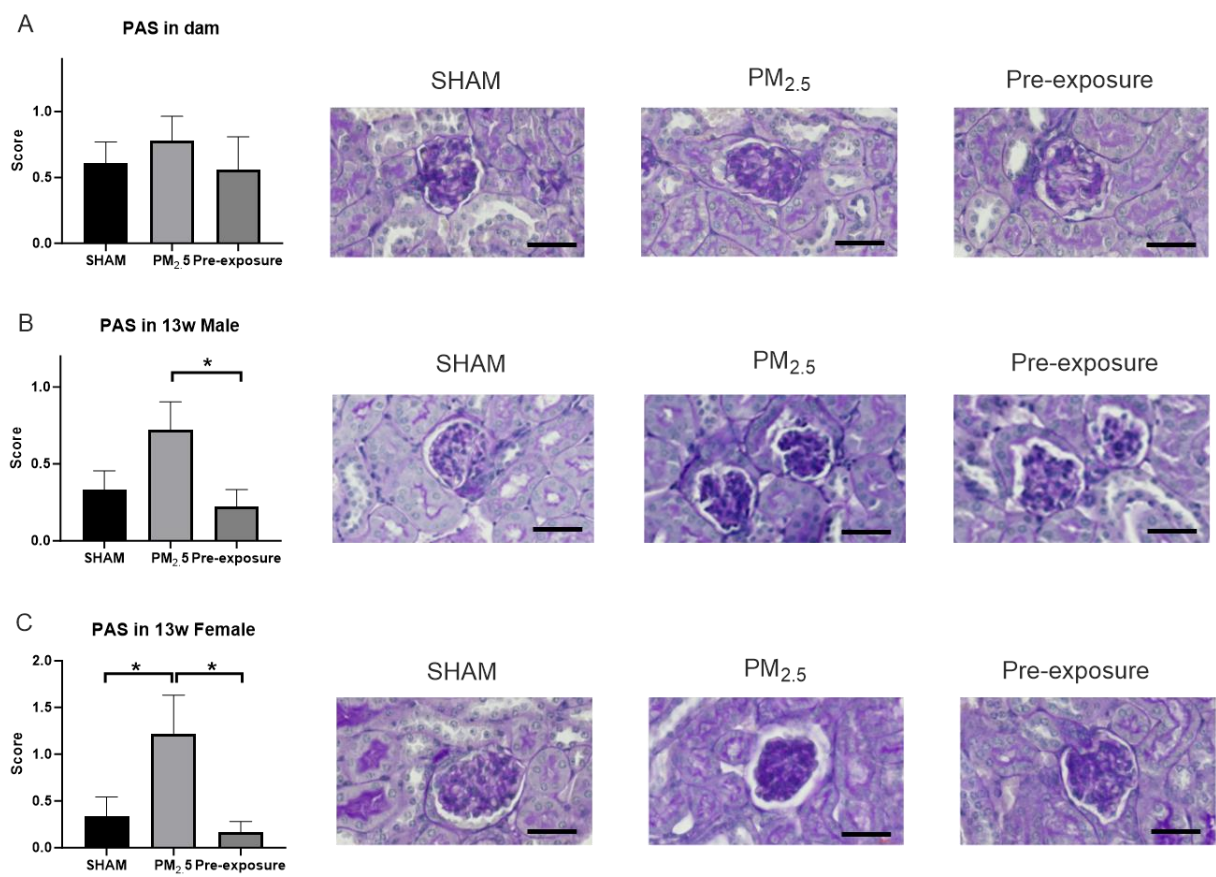
	Anthropometry	SHAM	PM <sub>2.5</sub>	Pre-exposure
Dams	Body weight (g)	27.2 ± 1.44	24.9 ± 0.53*	28.2 ± 2.46 <sup>##</sup>
	Kidney weight (g)	0.22 ± 0.019	0.19 ± 0.014*	0.21 ± 0.016
	Kidney weight/body weight	0.78 ± 0.05	0.78 ± 0.069	0.76 ± 0.07
Day 1 male	Body weight (g)	1.76±0.226	1.55±0.221*	1.56±0.303*
	Kidney weight (g)	0.0104±0.00057	0.0074±0.00064**	0.0098±0.00108 <sup>#</sup>
	Kidney weight/body weight (%)	0.5929± 0.0318	0.4710±0.0335*	0.6119±0.0415 <sup>#</sup>
13-week male	Body weight (g)	25.7± 1.23	26.6±1.9	26.3±1.5
	Kidney weight (g)	0.23 ± 0.005	0.23±0.007	0.24 ± 0.009
	Kidney weight/body weight (%)	0.90± 0.02	0.88±0.02	0.91±0.02
	Urine albumin / creatinine (µg/mg)	23.9 ± 2.50	64.3 ± 16.7*	14.6 ± 3.9 <sup>##</sup>
Day 1 female	Body weight (g)	1.55±0.208	1.62±0.256	1.49±0.308
	Kidney weight (g)	0.0082±0.00049	0.0081±0.00074	0.0093±0.00072
	Kidney weight/body weight (%)	0.5305±0.0316	0.4957±0.0427	0.6186±0.0388
13-week female	Body weight (g)	21.7±1.91	20.1±1.88*	20.8 ±0.593
	Kidney weight (g)	0.15±0.005	0.14±0.003*	0.15±0.003
	Kidney weight/body weight (%)	0.707±0.020	0.708±0.026	0.735±0.013
	Urine albumin / creatinine (µg/mg)	10 ± 3.39	16 ± 6.46	12 ± 7.02

549 Results are expressed as mean ± SEM and were analysed by one-way ANOVA followed by Turkey  
550 post hoc tests. N=9-14. \*P< 0.05 SHAM vs PM<sub>2.5</sub>, SHAM vs Pre-exposure. #P< 0.05, ##P< 0.01 PM<sub>2.5</sub>  
551 vs Pre-exposure. P: postnatal day. PM<sub>2.5</sub>: maternal PM<sub>2.5</sub> exposure prior to mating for 6 weeks, during  
552 gestation and lactation. Pre-exposure: maternal exposure to PM<sub>2.5</sub> for only 6 weeks prior to mating.  
553

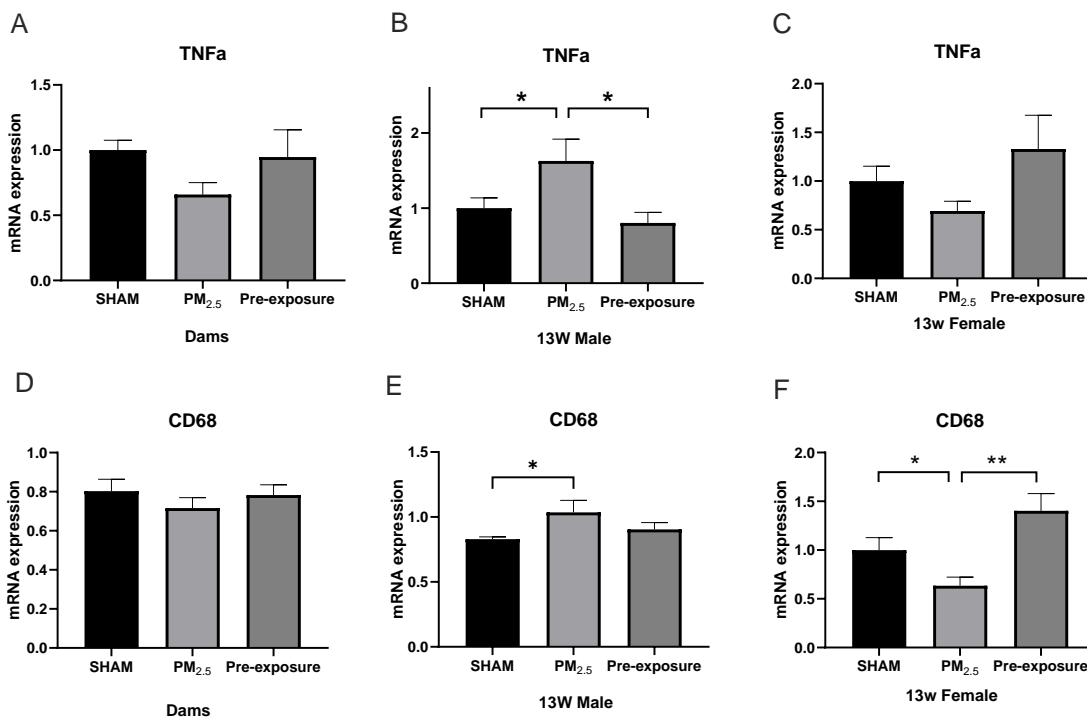


554 Figure 1. Glomerular number (A, B, C) and Glomerulus size (D, E, F) in the dams, 13 weeks old male  
555 and female offspring (n=5). Data were analysed by one-way ANOVA followed by Turkey post hoc  
556 tests. \*P < 0.05. PM<sub>2.5</sub>: dams exposed to PM<sub>2.5</sub> (5µg/day) prior to mating for 6 weeks, during gestation  
557

558 and lactation. Pre-exposure: dams exposed to PM<sub>2.5</sub> for 6 weeks prior to mating only.



559  
560 Figure 2. PAS staining of the glomeruli and representative images (mag 10x) in the dam (A), 13  
561 weeks old male (B) and female (C) offspring (n=5). Data were analysed by one-way ANOVA  
562 followed by Turkey post hoc tests. \*P<0.05. PM<sub>2.5</sub>: dams exposed to PM<sub>2.5</sub> (5µg/day) prior to mating  
563 for 6 weeks, during gestation and lactation. Pre-exposure: dams exposed to PM<sub>2.5</sub> for 6 weeks prior  
564 to mating only. Scale bar is equivalent to 100µm



565  
 566 Figure 3. Renal levels of TNF $\alpha$  mRNA (A, B, C) and CD68 mRNA (D, E, F) in dams, 13 weeks old  
 567 male and female offspring (n=7-8). Data were analysed by one-way ANOVA followed by Turkey  
 568 post hoc tests. \*P < 0.05, \*\*P < 0.01. PM<sub>2.5</sub>: dams exposed to PM<sub>2.5</sub> (5 $\mu$ g/day) prior to mating for 6  
 569 weeks, during gestation and lactation. Pre-exposure: dams exposed to PM<sub>2.5</sub> for 6 weeks prior to  
 570 mating only.

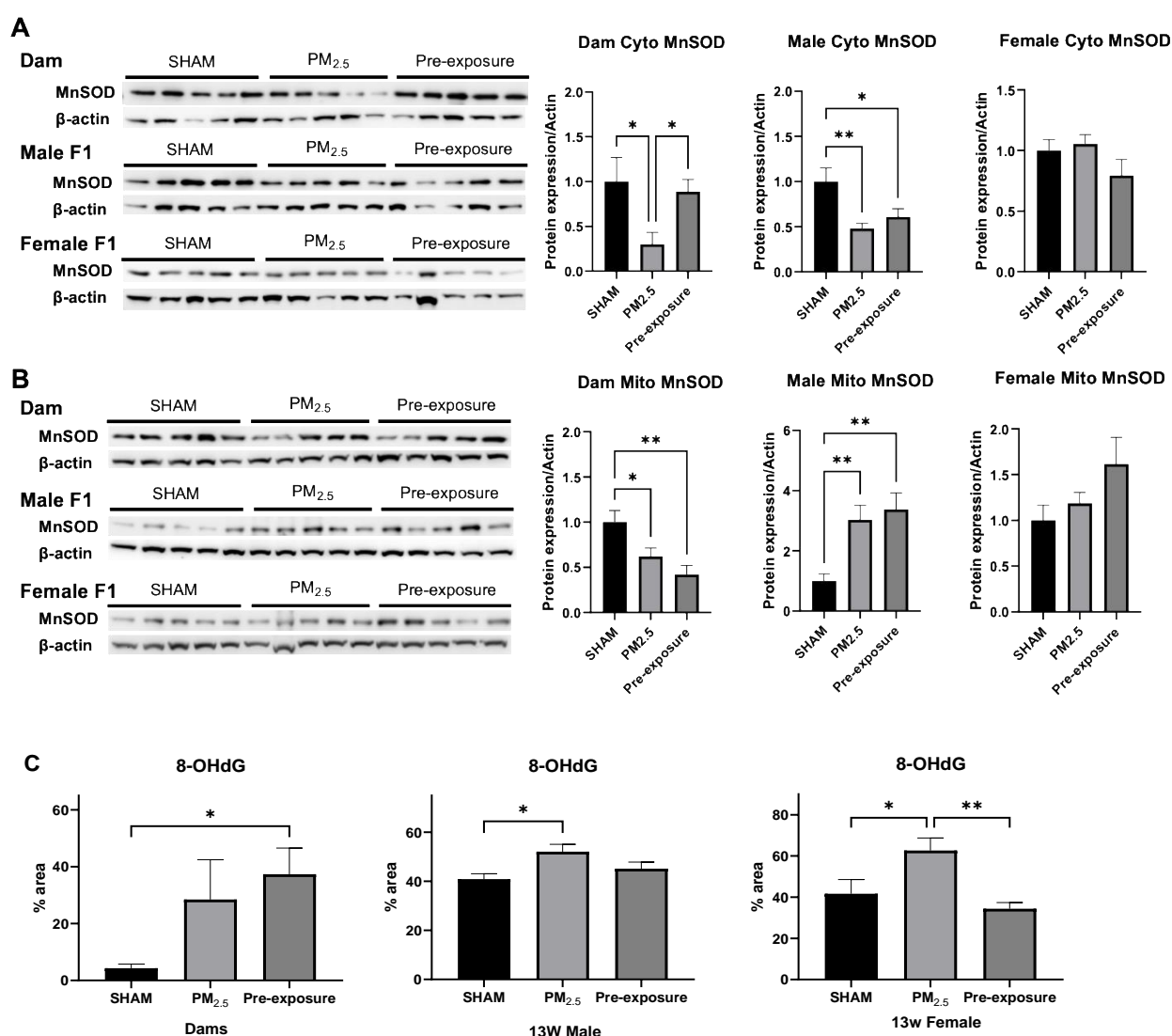
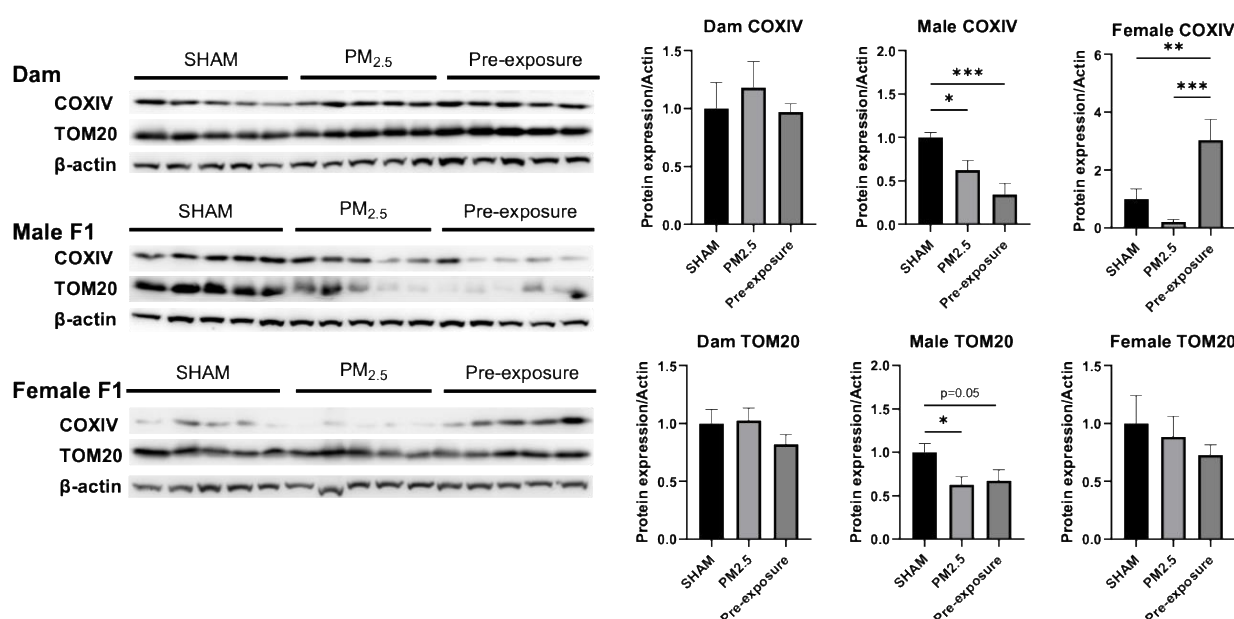
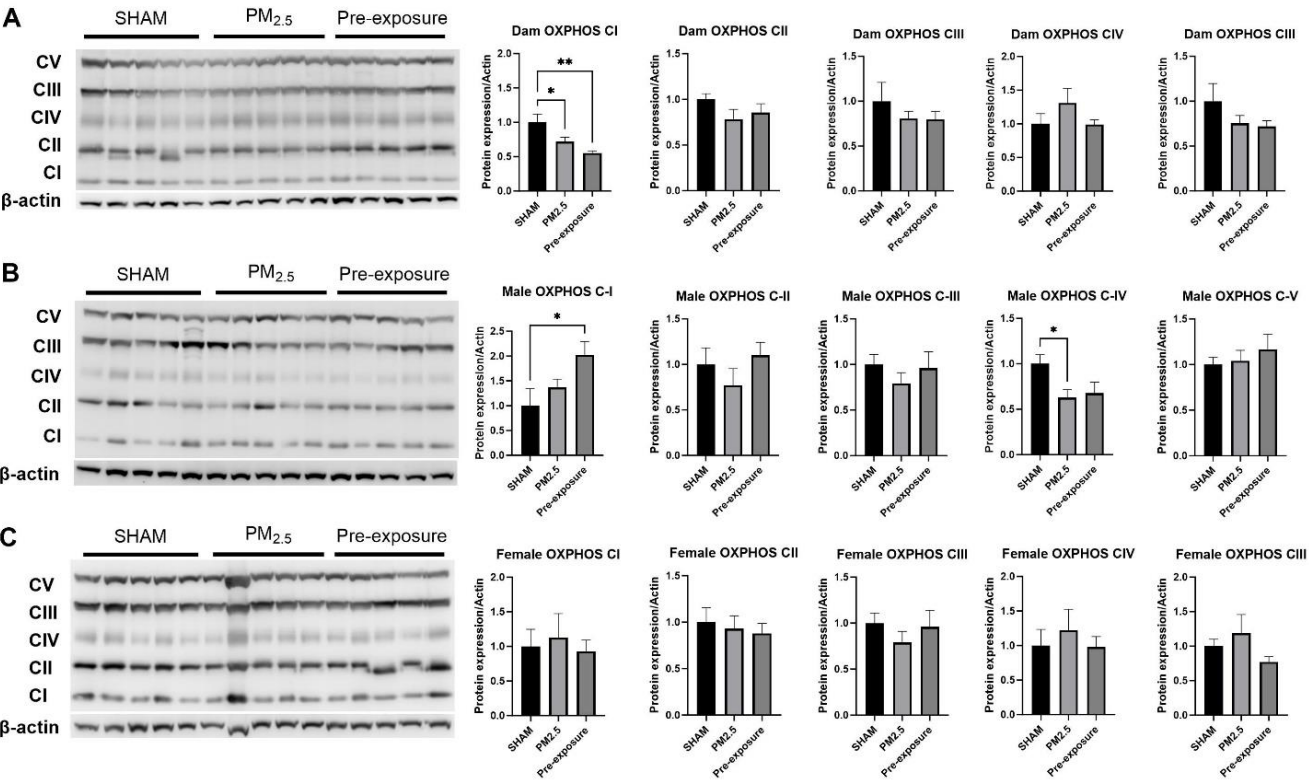


Figure 4. Cytosolic (A) and mitochondrial (B) MnSOD protein levels, as well as 8-OHdG staining in the kidneys in the dam, 13 weeks old male and female offspring (n=5). Data were analysed by one-way ANOVA followed by Turkey post hoc tests. \*P<0.05, \*\*P<0.01.



577 Figure 5. COXIV and TOM20 protein levels in the kidneys in the dam, 13 weeks old male and female  
 578 offspring (n=5). Data were analysed by one-way ANOVA followed by Turkey post hoc tests. \*P<0.05,  
 579 \*\*P<0.01, \*\*\*P<0.001.  
 580



581 Figure 6. OXPHOS complexes I-V protein levels in the kidneys in the dam (A), 13 weeks old male  
 582 (B) and female offspring (C) (n=5). Data were analysed by one-way ANOVA followed by Turkey  
 583 post hoc tests. \*P<0.05, \*\*P<0.01.  
 584