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Abstract: Ecophysiological studies of remnant woodlands in saline environments are scarce. We investigated seasonal fluctuations in soil water and salinity together with leaf and branch traits (area-based maximum assimilation (Amax), foliar nitrogen, specific leaf area (SLA) and Huber value (Hv)) and sap velocities of Eucalyptus macrorhyncha at four semi-arid sites in south-eastern Australia. Summer and winter soil salinities (10 cm depth) were 15-35 dS m-1 and 8-10 dS m-1 respectively. Gravimetric soil water content in the upper 20 cm was 2-5% in summer and 7-23% in winter, resulting in a significant inverse correlation between soil water and soil salinity. We found significant correlations between soil conditions and plant traits and function across seasons. Soil water content was significantly correlated with foliar N, SLA, Hv and maximum sap velocity while soil salinity was significantly correlated with Amax, Hv and maximum sap velocity. Correlations indicate co-variation of soil conditions and plant physiology in response to environmental conditions such as solar radiation and vapour pressure deficit (D). E. macrorhyncha tolerates the dual stresses of high salinity and low soil water during summer. While the plants appeared unhealthy, our data show that remnant vegetation can remain functional even in close proximity to saline scalds.

If you have simultaneous readings of RH and temperature, that is enough to calculate VPD-and could modify Fig.1 and text accordingly. (If they are averages, should not calculate them, because it is a non-linear relationship)

The values of temperature and RH used to plot Fig. 1 were daily averages and these were then averaged to obtain monthly averages. As noted, using average values to calculate VPD will not provide accurate estimates of VPD so we have left Fig. 1 unchanged.

Seasonal variations in tree water use and physiology correlate with soil salinity and soil water content in remnant woodlands on saline soils

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Author contributions: CM and MZ conducted the fieldwork, analysed the data and prepared the manuscript, AP contributed to data analysis and manuscript preparation, DE contributed to experimental design, analysis and manuscript preparation.

Highlights

- Dryland salinity is a significant problem in many semi-arid regions of the world.
- Woodlands combat salinity by lowering groundwater, reducing surface salt.
- We measured seasonal carbon assimilation, water use, soil salinity and water content.
- Some plant functional measurements were correlated with some soil measurements.
- Seasonal changes in meteorology drove unexpected plant-soil relationships.

Abstract

Ecophysiological studies of remnant woodlands in saline environments are scarce. We investigated seasonal fluctuations in soil water and salinity together with leaf and branch traits (area-based maximum assimilation (A_{max}), foliar nitrogen, specific leaf area (SLA) and Huber value (H_v)) and sap velocities of *Eucalyptus macrorhyncha* at four semi-arid sites in south-eastern Australia. Summer and winter soil salinities (10 cm depth) were 15-35 dS m⁻¹ and 8-10 dS m⁻¹ respectively. Gravimetric soil water content in the upper 20 cm was 2-5% in summer and 7-23% in winter, resulting in a significant inverse correlation between soil water and soil salinity. We found significant correlations between soil conditions and plant traits and function across seasons. Soil water content was significantly correlated with foliar N, SLA, H_v and maximum sap velocity while soil salinity was significantly correlated with A_{max}, H_v and maximum sap velocity. Correlations indicate co-variation of soil conditions and plant physiology in response to environmental conditions such as solar radiation and vapour pressure deficit (D). *E. macrorhyncha* tolerates the dual stresses of high salinity and low soil water during summer. While the plants appeared unhealthy, our data show that remnant vegetation can remain functional even in close proximity to saline scalds.

Keywords: salinity, drought, tree water use, photosynthesis, sap flow, *Eucalyptus* macrorhyncha

1. Introduction

Dryland salinity is widespread across Australia and other semi-arid regions of the globe (Lambers, 2003). Rising water tables bring salts to the surface of the soil (Eberbach, 2003) after clearing of deep-rooted perennial native vegetation has resulted in increased groundwater recharge (Lambers, 2003; Rengasmy, 2010). This serious land-degradation restricts agricultural productivity as increased salt concentrations and waterlogging of soils associated with dryland salinity can reduce plant growth and threaten plant survival (Sun and Dickinson, 1995). Studies testing methods for rehabilitation of saline soils have identified native salt-tolerant species which can be planted on salt-damaged soils and potentially improve soil quality (Bell, 1999). However, little is known about the ecophysiology of remnant vegetation remaining on saline soil (Pannell and Ewing, 2006). Consequently, the extent to which different native species respond to soil salinity *in situ* is not clear (Cramer and Hobbs, 2002).

Soil salinity affects plant function in a number of ways, for example, increased salinity reduces growth of stems, leaves and roots (Kayama et al., 2003; Myers et al., 1996; Rubinigg et al., 2004) in forest and greenhouse studies. The accumulation of salts in or below the root zone generally results in reduced tree growth and water use, and sometimes tree death (Macar et al., 1995; Sun and Dickinson, 1995; Feikema and Baker, 2011). High soil salinity can lead to changes in the soil-plant hydraulic conductance (Kayama et al., 2003; Rengasamy, 2006), leading to decreased tree water-use (Doody et al., 2009) which can end with death of non-halophytes (Akeroyd et al., 1998; Boland et al., 1996; Macar et al., 1995; Teobaldelli et al. 2004). This decreasing water-use may be a result of reductions in leaf area (Hatton et al., 1995; Hatton et al., 1998; Myers et al., 1998) or increased soil-to-leaf hydraulic resistance (Teobaldelli et al., 2004).

While the effects of high salinity on tree growth and tree water-use have been studied within many plantations (Benyon et al., 2001; Morris and Collopy, 1999; Hatton et al., 1998), few reports examine photosynthesis, growth and water-use of remnant vegetation on saline sites (Akeroyd et al., 1998; Barrett et al., 2005; Pannell and Ewing, 2006; Bann and Field, 2010; Bui, 2013). In one of the few studies of effects of salinity and waterlogging on native species *in situ*, Barrett et al. (2005) found very little measurable effect on pre-dawn leaf water potential, stomatal conductance and foliar ion concentrations, even when tree health was 3

visibly compromised. These results highlight the need to examine the ecophysiological response of native trees *in situ* because visible symptoms and impaired ecophysiological function may not always be correlated. Similarly, Marchesini et al. (2013) found predawn leaf water potentials at disturbed and undisturbed sites were unchanged despite lower gravimetric water content and lower electrical conductivity at disturbed semi-arid woodland sites. However, Bui (2013) identifies salt as a major ecological driver over geological and more recent time-scales in semi-arid areas. He suggests that salt has been largely ignored as an ecological driver and soil salinity should be considered together with other factors such as soil water and fire when exploring vegetation community composition (Bui, 2013). Pannell and Ewing (2006) advocate for cost effective management options to prevent expansion of salinised areas. Without adequate data on remnant vegetation, their value in the fight against salinity is not clear.

Seasonal variations in soil salinity are associated with seasonal rainfall (Tomar and Gupta, 1985; Srivastava and Jefferies, 1995; Rengasamy, 2010) and changes in watertable depth (Jackson et al., 1956; Peck, 1978). Heavy rains and rising watertables dilute salt in the soil while drier periods are associated with higher salinities due to evaporation and transpiration of soil water. Seasonal measurements are therefore required to capture the full range of combinations of environmental conditions which occur naturally. Trees in semi-arid regions affected by dryland salinity also have to tolerate winter frosts and high evaporative demands in summer. Seasonal patterns in function often reflect changing environmental conditions. For instance, the strong relationship between tree water-use and solar radiation (\mathbf{R}_n) , vapour pressure deficit (D) and soil water content (θ) is often exploited in simple empirical models of stand transpiration for well-coupled forest canopies (e.g. Whitley et al., 2008). In this study, we explored the patterns of θ and soil salinity associated with coarse seasonal patterns of water relations and productivity of remnant eucalyptus species in a region affected by dryland salinity. Specifically, we measured transpiration rates, hydraulic architecture, photosynthesis and leaf traits as well as θ and soil salinity across four sites in summer and winter and used correlation analysis to explore co-variation of plant function and soil conditions.

2. Materials and methods

2.1.Study site and species

The study sites were located near Crookwell, an agricultural and pastoral district in the southern Tablelands of New South Wales (34.4572 S, 149.4690 E, elevation 887 m). All sites

 had open forest with a grass understorey and very few shrubs and were within 3 km of each other. Site 1 at Laggan Corner was dominated by *Eucalyptus dives* Schauer (peppermint) directly adjacent to a saline scald, where salt crystals were forming on the soil surface and vegetation was absent due to the high concentrations of salt. Site 2 (also at Laggan Corner) was dominated by *Eucalyptus rossii* R. Baker & HG Smith (scribbly gum) and was 300 m from the nearest visible saline scald. *Eucalyptus macrorhyncha* F. Muell. ex Benth (red stringybark) occurred at both sites 1 and 2 as did *E. rossii*. Sites 3 and 4 at Laggan Dam were equally dominated by *Eucalyptus goniocalyx* F. Muell. ex Miq. (box), *E. macrorhyncha* and *E. dives*. Site 3 was directly adjacent to a saline scald appeared unhealthy with discoloured leaves.

Winter sampling took place in July 2003 and summer sampling occurred in January 2004. *E. macrorhyncha* was studied at all four sites. *E. rossii* and *E. goniocalyx* were also studied at Laggan corner and Laggan dam respectively but data are not shown because these data were very similar to those for *E. macrorhyncha*. Red stringybark is wide-spread in south-eastern NSW and locally dominant in dry schlerophyll forests and woodlands. It is hardy, tolerating frosts and occasional snow (Boland et al., 2006). Six sample trees were identified within two plots at each site (three trees per plot). All plant measurements were conducted on these trees. Leaf and branch samples were selected haphazardly from within 5 m of the ground.

This research was conducted during a prolonged and extreme drought lasting three years, with severe water deficiencies with rainfalls in the lowest 5% of historical totals (Bureau of Meteorology 2003). The region is characterised by a long-term mean annual rainfall of 861 mm but is located in a rain shadow so the sites commonly experience drought conditions with extended periods (months to years) with little rainfall. There is slightly more rain in winter than summer and average temperatures are 10 - 27 °C in summer and 0 - 10 °C in winter (Fig. 1). Climate data were obtained from the nearest Bureau of Meteorology station at Crookwell Post Office (approximately 8 km to the south of the experimental sites).

2.2.Soil salinity and water content

The soil is defined as coarsely cracking grey and brown clays (NSW Natural Resource Atlas, http://nratlas.nsw.gov.au, accessed on 22nd July 2013). The soil type is a sodosol (Isbell, 2002) and is classed as sodic saline soil according to Rengasamy's (2010) categories. Particle size analysis indicated the clay content was approximately 30%. Five soil samples were 5

collected from the base of each of the six sap flow sample trees to assess salt content of the soil at each of the four sites. Approximately 1000 cm³ were excavated at a depth of 10 cm and transported in a zip-lock bag to the laboratory. An aqueous saturated paste extract (of 50 mL water to 10 g air-dried soil) was produced by shaking the sample for 5 min then allowing to settle for an hour, following the methods of Rhoades (1982). The electrical conductivity (dS cm⁻¹) of this solution was measured with a conductivity meter (YSI Incorporated, Ohio) and divided by 100 to produce values of dS m⁻¹. The salinity of the saturated paste extract was multiplied by 8.6 (for a sandy clay loam) to determine the salinity experienced by roots (Taws, 2003). The remaining soil was dried to constant weight (110 °C) and weighed to determine soil water content. As our sites were privately owned, we did not have permission to dig deep holes so we were limited to non-invasive soil measurements in this project.

2.3.Photosynthesis

Photosynthesis was measured as maximum (light saturated) assimilation (A_{max}) with a HCM-1000 Portable Photosynthesis System (PPS) (Walz, Germany). Measurements were taken in the morning (between 0930 and 1130) and in the afternoon (between 1400 and 1600), to avoid any effects of midday down-regulation. The PPS is often used in the field to measure leaves still attached to plants (McPherson et al., 2004) but due to the height of leaves on the eucalypts in this study, small branches were cut off the trees with a long-armed pruner and placed in water. These branches were re-cut underwater and placed under three 50 watt halogen lamps delivering approximately 600 µmol photons m⁻² sec⁻¹ at leaf height, standardising light conditions. Preliminary investigations showed that removal of branches from the tree did not cause a significant change in the rate of photosynthesis over a 45 min period (data not shown).

Measurements were taken at ambient temperature, humidity and atmospheric CO_2 concentrations. The PPS was operated in differential mode with a 5 cm² closed cuvette, with a 2 mm fibre optic cable inserted through the cuvette lid for fluorescence measurements. Once the CO_2 differential had stabilised, A_{max} (defined as the rate of carbon fixation per unit of leaf under saturating light conditions) was measured. Assimilation rate was measured on three leaves on three plants in each plot, with two plots at each site (a total of 18 samples per site).

2.4. Foliar nitrogen content and specific leaf area

Leaves used in photosynthetic determination were collected for nitrogen analysis. Samples were placed in moistened snap-lock bags and placed on ice for transportation to the laboratory. Until analysis, samples were stored at -80°C. Leaf samples were prepared by maceration in a ball grinder and nitrogen content was measured with a carbon and nitrogen analyser (TruSpec CN 2000, Leco Corporation, Michigan).

Specific leaf area was calculated for three leaves from each sample tree (18 leaves at each site). Samples were transported on ice in a cool box to the laboratory. Five 1 cm² leaf discs were cut with a cork borer and dried at 70°C until constant weight (usually 5 days). Each disc was weighed to four decimal places and the specific leaf area (SLA) was calculated as the ratio of area to dry mass of the disc.

2.5. Huber value

Huber values (H_v) were calculated as the branch transverse area per leaf area distal to the cut such that $H_v = SA/LA$ where SA is the cross-sectional area of the sapwood of the branch and LA is the leaf area supported by that section of branch. The diameter of the basal end of each branch was measured using a micrometer and the length of each branch was measured with a ruler. Leaf area was determined using a WinDIAS (Type WDIGC-2 Delta-T Devices Ltd, U.K) leaf area meter. Sapwood cross-sectional area was determined by soaking thin sections of branch in 0.5% toluidine blue solution. The stained sapwood area was measured with a calibrated eyepiece micrometer.

2.6.Tree water-use

Tree water-use was measured on *E. macrorhynchya* using sapflow sensors following Zeppel et al. (2004). The sapwood depth of each tree was determined by inspecting samples collected with an increment borer. There was a visible colour change between conducting and non-conducting sapwood. The probe wound diameter was measured by taking wood samples back to the laboratory and measured using a binocular microscope. Mean wound width was estimated to be 2.67 mm. The radial profile of sap velocity through the sapwood depth was measured by moving probes gradually further into the profile over a number of days to ensure the probes were positioned in conducting sapwood. Heat-pulse method sapflow sensors (Greenspan Technology Pty. Ltd., Warwick, Australia) were inserted into the sapwood of six

tree at varying depths (5-10 mm below the cambium) to obtain a range of sap velocities. Sap velocities were recorded at hourly intervals and were used to estimate volumetric water use (Q) following Edwards and Warrick (1984) but for clarity, we used peak flow rates (cm h^{-1}) in our analysis. Peak (or maximum) flow was the highest flow rate measured across the day for each tree. Sensors were positioned in sapwood according to Zeppel et al. (2004) so as to capture the fastest sap flow rate.

At each site, tree water-use was measured for trees in two size classes, trees with a diameter at breast height (DBH) < 20 cm (small) and trees with a DBH > 30 cm (large). Trees were measured during a four week sampling period in July and August 2003, and in December to January 2004. A roaming sampling strategy was used. During the first two weeks, three or four trees were measured at each of the Laggan Dam sites, and one tree was measured at the Laggan Corner sites. During the next two week period one tree was measured in both Laggan Dam sites, while equipment was removed from the other two or three trees and set up in three or four trees at Laggan Corner. By establishing the relationship between the fixed and roaming sensors, sap flow rates could then be inferred for each tree for the whole period (Zeppel et al., 2004).

2.7. Statistical analysis

We explored the seasonal relationship between soil water and soil salinity as well as the correlation of these two factors with measured plant function and physiology. Correlation analysis was conducted with Spearman-rank correlations (two-tailed) using SPSS v. 14.0, 2005.

3. Results

3.1.Climate, soil salinity and water

Sampling occurred after a prolonged drought (Bureau of Meteorology, 2003) of over three years. Rainfall in the three months preceding and during the winter sampling period was lower than average (Fig. 1). The three months between sampling periods had higher than average rainfall, while the summer sampling period also had lower than average rainfall. The mean maximum daily temperature was 27.5 ° C in summer and 11.3 ° C in winter (Fig. 1). Minimum relative humidity was 57%, decreasing to 31% in summer, and there was a large degree of variation in humidity between days, in summer, compared with winter (Fig. 1).

 Soil salinity ranged from 6.9 dS m⁻¹ to 29.5 dS m⁻¹ while soil water content ranged from 3-23% (Fig. 2). In both seasons, soil salinity and soil water contents (θ) were not significantly different between sites directly adjacent to and 2-300 m away from the saline scalds (statistical data not shown). These results indicated that elevated salinities within saline scalds were localised to the scald and not detectable within the patches of woodland. However, there was a significant negative correlation between soil water content (θ) and soil salinity across sites and seasons (Fig. 2). Drier summer soils were associated with saltier conditions, while wetter winter soils were less saline, consistent with more rainfall in winter (Fig. 1). On the whole, seasonal differences in salinity were larger than differences between sites (Fig. 2).

3.2. Soil properties and plant function

Several measures of leaf gas exchange and leaf traits were significantly correlated with soil salinity or water (Fig. 3). Leaf-area-based assimilation rate (A_{max}) was negatively correlated with soil salinity, and specific leaf area (SLA) was negatively correlated with soil water content, while the concentration of foliar nitrogen was positively correlated with soil water content (Fig. 3). Rates of assimilation ranged from 10-22 µmol m⁻² s⁻¹ (Fig. 3). A_{max} was higher in winter than summer. Specific leaf area ranged from 43-70 cm² g⁻¹ with higher values in summer than in winter, while foliar N concentration ranged from 2-3 g m⁻² with slightly higher values in winter than in summer (Fig. 3).

Huber values ranged from 2 to 4 x 10^4 across all sites, with larger values observed in winter than in summer (Fig 4). Maximum daily sap velocities generally occurred close to midday ranged between 8 and 37 cm h⁻¹ across all sites and seasons (Fig. 4). Sap velocities were higher in summer (18-37 cm h⁻¹) than winter (8-19 cm h⁻¹). Huber value and maximum sap velocity were both significantly correlated with soil salinity and water (Fig. 4). Huber value was negatively correlated with soil salinity and positively correlated soil water while maximum sap velocity was positively correlated with soil salinity and negatively correlated with soil salinity and negatively correlated with soil water (Fig. 4).

4. Discussion

Our results indicate that at these semi-arid sites, plant function, soil salinity and soil moisture (in the top 10 cm) are all seasonally variable. Assimilation rate (A_{max}), SLA, sap velocity and Huber value (H_v) were the most changeable between seasons, while foliar N was more stable 9

across the year but values recorded for all variables were indicative of healthy vegetation in other parts of Australia. We present a valuable dataset describing the value of remnant trees in an area suffering from dryland salinity. Despite the proximity of some of our sample trees to severe saline scalds, the trees remain fully functional with strong rates of photosynthesis and high sap flow rates. Visually assessing the health status of leaves (Seddon et al., 2007) may not be the best way of defining the severity of saline impacts. While we have identified some correlations between soil conditions and plant function across seasons, meteorological conditions (especially evaporative demand) are the likely key drivers of these soil and plant processes.

4.1.Soil salinity

Soil salinity concentrations are classified as 'moderate' for 2 to 6 dS m⁻¹, 'high' for 6 to 15 dS m⁻¹ and over 15 dS m⁻¹ is 'extreme' (Richards, 1954; Taws, 2003). Winter soil salinity concentrations were high at all sites at Crookwell, increasing to high and extreme levels in summer. In both winter and summer, there was a large degree of variation in soil salinity within sites due to variation in distance from the saline scald within each site (data not shown) but high spatial variability of soil salinity is common (Feikema and Baker, 2011). The abrupt change in soil salinities inside and outside of vegetated areas (data not shown) was also noted by Taws (2003) and is consistent with the observations of Ban and Field (2010) that salinised sites tend to be small and localised.

Soil salinity may be highly variable in space and time but salinity of saturated extracts of soils at Crookwell (8-35 dS m⁻¹) were within the range of other saline sites in Australia (Thorburn et al., 1993; Benyon et al., 1999; Morris and Collopy, 1999; Taws, 2003) so our results are likely to be representative of other patches of remnant vegetation. Within each site, there was some variation in soil salinity within a season as reported by Taws (2003). Similarly, an increase in salinity from winter to summer has been reported in similar vegetation in NSW (Benyon et al. 1999; Taws, 2003). Transient salinity is associated with patterns in rainfall, surface evaporation, transpiration and leaching of the clay layer (Rengasamy, 2010). It is common to find increased soil salinity during summer due to increased evapotransiration removing moisture from the soil and concentrating salts (Rengasamy, 2006). However, our soil salinity method of drying samples before making the saturated soil extract reduced the influence of seasonal variation in soil water content on salt concentrations. Therefore, the significant inverse correlation in Fig. 2 does not merely show the effect of soil water content changes on soil salinity, it indicates possible seasonal mobility of salts unrelated to soil water 10

content at the time of sampling. While most changes in salinity are likely to involve water as a driver, it is difficult to tease apart the contributions of the different factors. We expect that soil salinity would decline through the soil profile (Wong et al., 2009) so our values may overestimate the salinity deeper roots are exposed to. What we can say for certain is that our sample trees tolerated a large fluctuations of soil salinities within the moderate to extreme ranges. These patches of vegetation are therefore very hardy and resistant to salt impacts.

4.2. Tree physiology

Rates of carbon assimilation, foliar N and SLA were similar to those of other Australian native plants, including eucalypts (e.g. Eamus et al., 1999; Warren et al., 2000; Wright et al., 2001; MacPherson et al., 2004; Lewis et al., 2011; Koerber et al., 2012) indicating that E. macrorhyncha was healthy despite the recent prolonged drought and saline soils. In order to understand the broader meaning of the correlations reported here, it is important to consider the drivers of seasonal variation in soil salinity and water and plant traits and function. Seasonal variation in assimilation rates is common across a variety of ecosystems. Changes in temperature, solar radiation and water availability are common factors which influence photosynthesis rates, often with higher rates in summer (King and Ball, 1998; MacPherson et al., 2004). However, the A_{max} measurements in this study were higher in winter than in summer (Fig. 3). The inverse correlation between soil salinity and Amax lead to higher salinities being associated with lower Amax values. While high salinity in the soil causes impairment of the photosynthetic apparatus (Lawlor, 2002) or degradation of photosynthetic pigments and the thylakoid membrane (Ashraf and Harris, 2013), it is more likely that the reduction in A_{max} during summer was caused by stomatal limitation rather than biochemical limitation of photosynthesis because assimilation rates were within the range of healthy eucalypts (Lewis et al., 2011) and Doody et al. (2009) found very little change in stress of trees as watertables freshened and lowered. Higher concentrations of salt in the soil were correlated with lower soil water content (Fig. 2) which would cause decreasing leaf water potentials during summer. Dry conditions in air and soil result in declining stomatal conductance which causes a decline in assimilation rate (Lawlor, 2002). We attribute the inverse correlation between soil salinity and Amax to reduction in stomatal conductance during summer, when soil water content was reduced and soil salinity increased. The lack of a significant correlation between soil water and A_{max} is likely due to other complicating factors such as the patchiness of soil water and water uptake from deeper soil. Unfortunately, we do not have stomatal conductance measurements to confirm this but the mechanistic framework

we describe is supported results from other systems (Lawlor, 2002; Flexas et al., 2004; Koerber et al., 2012). Further data on depth of soil water uptake is also required to better understand the salinity and importance of deep soil moisture.

Changes in foliar N and specific leaf area (SLA) may have also contributed to seasonal variation in A_{max}. There was a slight (but not significant) decline in foliar N during summer which may have contributed to declining assimilation rates (Fig. 3; Reich et al., 1998), seasonal patterns in SLA were associated with the onset of observed new leaf growth (leaf flushing) in response to soil water. In the three months prior to summer sampling, rainfall was above average (Fig. 1). This additional soil water prompted growth of new leaves, probably as a response to rising water potentials of soil and consequently of leaves (Williams et al., 1997). Our sample leaves were fully expanded but they were clearly younger than the leaves in winter and this caused higher SLA values in summer (Fig. 4). Aging of eucalypt leaves is associated with increased sclerophylly and therefore decreased SLA (Wright and Cannon, 2001). The SLA values measured at Crookwell were low on an international scale, across multiple biomes (e.g. Reich et al., 1995) but consistent with other Australian studies (e.g. Specht and Rundel, 1990; Eamus et al., 1999; Koerber et al., 2012). Foliar N concentrations are also low but consistent with nutrient depleted soils of Australia (Specht and Rundel, 1990; Warren et al., 2000).

Warmer and drier sites are associated with increased Huber values (Mencuccini, 2003) as the water relations of these sites reduce photosynthetic revenue (Westoby et al., 2012). However, seasonal variation in H_v at our saline sites showed declining H_v values as soil became drier during summer because like SLA, H_v was also influenced by summer leaf flush (Fig. 4; Macinnis-Ng et al., 2004). The significant correlations for H_v with salinity and soil water content represent coordinated seasonal variation of these variables rather than a direct response relationship. Coordination of H_v and tree water-use in response to changes in soil water content and atmospheric aridity has been observed in eucalypts previously (Eamus et al., 2001; Carter and White, 2009).

Maximum sap velocity was significantly correlated with both soil salinity and soil water content (Fig. 4). Values were similar to those of Cook and O'Grady (2006) and the seasonal pattern involved higher maximum flow rates in summer similar to that observed by Zeppel et al. (2004). When the difference between soil salinity at different sites is large, other studies

report that increasing soil salinity decreases tree water-use (Akeroyd et al., 1998; Benyon et al., 1999; Benyon et al., 2001; Boland et al., 1996; Kozlowski, 1997; Teobaldelli et al., 2004; Doody et al., 2009). Reduction in sap velocity is attributed to increases in soil-to-leaf hydraulic resistance (Teobaldelli et al., 2004) and the correlations reported here are likely to have been driven by physiological responses to seasonal factors (described below) other than soil conditions. Zeppel et al. (2004) identified three factors which led to increased sap flow during summer: peak solar radiation, increased evaporative demand due to elevated air temperatures and D and increased soil water. In the present study, the first two factors were present but soil water content was lower in summer than winter. Therefore, we conclude that solar radiation and evaporative demand were more important in controlling maximum rates of sap flow than soil water content at our Crookwell sites. Despite indications that stomatal conductance was limiting Amax (discussed above), stomatal closure was insufficient to prevent higher maximum sap flow in summer. Increased evaporative demand combined with leaf flushing (which increased leaf area) were overriding any water-savings through stomatal closure. The combined seasonal pattern of Amax and maximum sap velocity implies water-useefficiency declined during summer, possibly due to new leaf growth. Seasonal variation in tree water-use was a potential driver of decreasing soil water during summer (when tree water-use was enhanced relative to the winter period), causing increased concentrations of salt in the soil, particularly in the root zone (Feikema et al., 2010).

Overall, we found very little indication that physiology of *E. macrorhyncha* was compromised by seasonal fluctuations of soil water and salinity because we did not find significant differences between seasons for the measured variables (ANOVA data not shown). A number of *Eucalyptus* species survive and even thrive in saline soils (Macar et al., 1995) so it is not surprising that our study species remains productive and functioning at these sites. In one of the few studies investigating physiological condition of native vegetation exposed the saline groundwater, Barrett et al. (2005) found that predawn leaf water potential, stomatal conductance and foliar salt concentrations were similar in trees exposed to varying depths and salinities of ground water. Foliar concentrations of Na and K were high and varied seasonally at all sites, independent of groundwater salinity and these authors suggested root hydraulic redistribution of water between different soil layers may account for the lack of clear physiological impacts of soil salinity. Similarly, lowering and freshening of watertables did not have a detectable effect on tree stress (Doody et al., 2009). As woody thickening (the increasing density of shrubs and trees) is enhanced by climate change in semi-arid regions due

to boosted water use efficiency of C3 plants (Macinnis-Ng et al., 2010), the water table will lowered if sufficient vegetation establishes. Our results suggest *E. macrorhyncha* is a promising candidate for phytoremediation under a changing climate (Jesus et al., 2015).

Further research is needed to disaggregate plant responses to seasonal fluctuations in soil water and salinity and the role of the drought prior to our measurements. Identifying the depth of soil water extraction by roots will improve our understanding of the impact of soil salinity on tree physiology, particularly plant water use and the need for salinity adaptations. If the trees are able to use deeper, cleaner water then surface fluctuations may not be an issue. Using isotopic analysis and water balance calculations to identify sources of transpired water (Feikema et al., 2010) along with more detailed physiological measurements will indicate possible physiological adjustments these trees are using. *E. macrorhyncha* is commonly found in saline and non-saline areas of NSW (Taws, 2003) but there is little information available on the salinity tolerance of this species. Our data clearly indicate that *E. macrorhyncha* can be resistant to high and extreme levels of soil salinity in the upper soil profile and is therefore an excellent candidate for growth in saline areas. Furthermore, *E. macrorhyncha* is capable of surviving fluctuations in soil salinity and water across seasons and is therefore suited to the changing climate of south-eastern Australia.

4.3. Conclusions

On the whole, our sample species was very tolerant to field-based seasonal variations in topsoil salinity. Remnant woodlands of *E. macrorhyncha* and associated species provide a carbon store and pathway for transpiration. Such patches of native vegetation are also likely to provide habitat and food for native fauna and are therefore a rich resource for conservation and biodiversity. Visual inspection of our trees suggested those closest to saline scalds may be physiologically compromised. However, our measurements showed no difference between trees with or without leaf discolouration. We emphasise the importance of conserving existing patches of remnant vegetation. Our data demonstrate the role of remnant woodlands in ameliorating saline scalds as the trees adjacent to scalds were able to maintain performance comparable to that of trees further from the scalds. Correlations between soil water content, salinity and tree traits do not reflect causation but rather indicate that seasonal variation in meteorological conditions is driving plant function, soil water and salinity in a coordinated way. A complete suite of environmental variables (including meteorological conditions) must

be collected when measuring plant performance on saline soils to properly understand drivers of plant productivity and water use and seasonal variation in these values must be considered.

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Figure captions

Fig. 1. Mean annual temperatures, rainfall and relative humidity (shown as 100 year mean) and actual rainfall for April 2003- March 2004 for Crookwell meteorological station. (Data source: Australian Bureau of Meteorology).

Fig. 2. Relationship between soil water content and soil salinity of the saturated paste extract across seasons for the four sites. Each point represents the site mean \pm standard error of the mean in either summer or winter. Points joined with a line are summer and winter values for one site.

Fig. 3. Relationships between saturated soil salinity and soil water content and area-based assimilation under saturating irradiance (top panels), foliar N (middle panels) and specific leaf area (lower panels). Each point represents the mean \pm standard error of the mean (n = 6 trees) for each of the four sites in summer or winter. Summer values fall to the right of salinity plots and to the left of soil water content plots. Correlation values are Spearman-rank coefficients.

Fig. 4. Relationships between saturated soil salinity and soil water content and Huber value (top panels) and maximum sap velocity (lower panels). Each point represents the mean \pm standard error of the mean (n = 6 trees) for each of the four sites in summer or winter. Summer values fall to the right of salinity plots and to the left of soil water content plots. Correlation values are Spearman-rank coefficients.

Highlights

- Dryland salinity is a significant problem in semi-arid regions of the world.
- Woodlands combat salinity by lowering groundwater, reducing surface salt.
- We measured seasonal carbon assimilation, water use, soil salinity and water content.
- Some plant functional measurements were correlated with some soil measurements.
- Seasonal changes in meteorology drove unexpected plant-soil relationships.

Figures and Tables

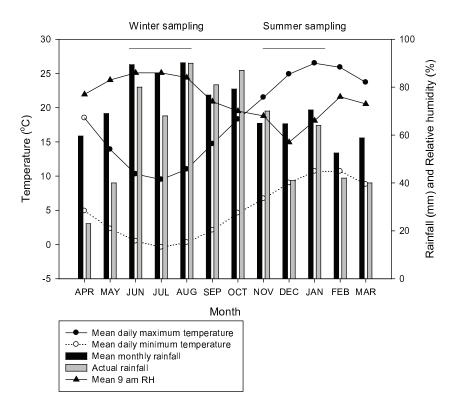


Fig. 1. Mean annual temperatures, rainfall and relative humidity for Crookwell meteorological station. Data represent the 100 year mean. In addition to actual rainfall for April 2003- March 2004. (data source: Australian Bureau of Meteorology).

