

1 **Predicting range-shift success potential for tropical marine fishes**
2 **using external morphology**

3

4 Shannen M. Smith^{1*}, Rebecca J. Fox^{1,2}, Jennifer M. Donelson^{1,3}, Megan L. Head²,
5 David J. Booth¹

6

7 ¹ School of Life Sciences, University of Technology Sydney, Ultimo, NSW 2007,
8 Australia

9 ² Division of Evolution, Ecology and Genetics, Research School of Biology,
10 Australian National University, Canberra 2600, Australia

11 ³ ARC Centre of Excellence for Coral Reef Studies, James Cook University,
12 Townsville, QLD 4811, Australia

13

14 *Author for correspondence

15 E-mail: shannen.m.smith@outlook.com

16

17

18 **Abstract**

19 With global change accelerating the rate of species' range-shifts, predicting which are
20 most likely to establish viable populations in their new habitats is key to
21 understanding how biological systems will respond. Annually, in Australia, tropical
22 fish larvae from the Great Barrier Reef (GBR) are transported south via the East
23 Australian Current (EAC), settling into temperate coastal habitats for the summer
24 period, before experiencing near-100% mortality in winter. However within 10 years,
25 predicted winter ocean temperatures for the southeast coast of Australia will remain

26 high enough for more of these so-called “tropical vagrants” to survive over winter.
27 We utilised a method of morphological niche analysis, previously shown to be an
28 effective predictor of invasion success by fishes, to project which vagrants have the
29 greatest likelihood of undergoing successful range shifts under these new climatic
30 conditions. We find that species from the family of butterflyfishes (Chaetodontidae),
31 and the moorish idol, *Zanclus cornutus*, are most likely to be able to exploit new
32 niches within the ecosystem once physiological barriers to overwintering by tropical
33 vagrant species are removed. Overall, the position of vagrants within the
34 morphospace was strongly skewed, suggesting that impending competitive pressures
35 may impact disproportionately on particular parts of the native community.

36

37

38 **Keywords:** biological invasion, climate change, coral reef fishes, morphological
39 niche, tropicalisation

40

41

42 **1. Introduction**

43

44 Human-induced warming of the planet is driving shifts in the distributional ranges of
45 many terrestrial [1] and marine [2] organisms. Predicting which species are most
46 likely to be successful in establishing reproducing populations within their new
47 habitat range is a critical step in forecasting the potential changes in biodiversity and
48 ecosystem functioning that will occur as a result of continued global warming [3-5].

49 In the marine environment, poleward shifts in the distribution of tropical
50 species are already resulting in the “tropicalization” of temperate marine habitats [6].

51 The southeast coast of Australia lies at a known ocean warming hotspot [7-8] and is
52 therefore likely to be particularly susceptible to the process of tropicalization. Every
53 year tropical fish larvae are expatriated south from the GBR via the EAC, where they
54 arrive and settle in recruitment pulses from Jan-May [9]. Currently, juvenile mortality
55 reaches ~100% by July (Austral winter), driven by thermally induced reductions in
56 physiological capacity. However, sea surface temperatures in this region are expected
57 to rise by as much as 3°C by 2100, and coupled with the southward movement of the
58 EAC [10], more of these tropical species are predicted to soon be capable
59 overwintering at temperate latitudes. The winter ocean temperatures forecast to occur
60 in the Sydney region by 2020-30 [8] will be above the 17°C threshold currently
61 estimated to be the constraint on physiological survival of tropical vagrant species
62 [11], meaning that establishment of populations of tropical reef fishes in the waters
63 off Sydney and the southeast coast of Australia will be possible within the next 15
64 years. Predicting which of these vagrant species are most likely to establish breeding
65 populations, and therefore compete with native species for food and habitat resources,
66 is a critical issue for those concerned with managing these ecosystems into the future.

67 Many individual species-level traits have been proposed as predictors of
68 geographic range shift potential, (e.g. high propagule production, generalist diet).
69 However the predictive power of the trait-based approach across multiple ecosystems
70 and regions has been shown to be mixed, with some traits being identified as being
71 important in certain contexts but not in others [reviewed in 12]. Recently, Azzurro *et*
72 *al.* [13] demonstrated that external morphology alone can be an accurate predictor of
73 a species' ability to successfully colonise a novel habitat. By using morphology as a
74 proxy for a species' ecological niche, the method maps the existing resident
75 community in terms of its morphological niche space and examines the position of

76 potential invading species within this morphospace. Based on present-day abundances
77 of species that have previously invaded the Mediterranean Sea via the Suez Canal,
78 they showed that species which established successful populations were those that
79 either lay outside, or at the margins of, the resident morphospace. Those that were
80 morphologically similar to residents were unlikely to establish, presumably because
81 the ecological niche was already filled [14-15]. Here we use this method to examine
82 the position of seasonally-invading tropical reef fish from the GBR within the
83 morphospace of a native fish community of the southeast coast of Australia and to
84 predict which vagrant species might be most likely to establish residency and shift
85 their range as the physiological barriers to survival recede under global warming.

86

87 **2. Material and methods**

88

89 The morphological space of the native shoreline fish community of New South Wales
90 (NSW), Australia was defined based on the species assemblage of a representative
91 protected marine habitat (Cabbage Tree Bay, NSW, <https://ctbar.wordpress.com/>). A
92 database of 110 species of bony fishes was compiled (electronic supplementary
93 material, table S1) and an image of each species (left hand side of adult individual)
94 was sourced from online resources. Species' morphology was then characterised
95 using 27 anatomical landmarks with ecological significance [following 13,16,17],
96 (electronic supplementary material, table S1). Landmarking of images was done using
97 tpsDig v1.40 [18], and scaled via a generalised least squares procedure (generalised
98 Procrustes) to eliminate the effect of isometric body-size variation. Relative warp
99 analysis (tpsRelw v1.60) [19] was then used to obtain coordinates (relative warp
100 scores) for each species within the reduced shape space [20]. The first two warps

101 accounted for 47% of the observed morphological variation within the community
102 and factor scores from these two warps were used to plot the position of individual
103 species and define a morphospace of the native fish community in 2-dimensions. The
104 convex hull of the native community (the smallest set of points enclosing all 110
105 species, see [21]) was then calculated using Qhull [22] and the resulting convex hull
106 broken down into a series of Voronoi polygons to provide an estimate of the
107 ecological space around each species [23]. For species at the periphery, Voronoi
108 polygons are unbounded and use of the convex hull as a boundary therefore results in
109 an underestimation of the area surrounding peripheral and hull points.

110 The analyses were repeated for a fish community containing the 110 native
111 species plus the 11 tropical vagrant species most commonly observed as juveniles in
112 the Sydney region over at least two consecutive summers in the last 15 y [24],
113 (electronic supplementary information table S1). Relative warp analysis yielded the
114 coordinates of the vagrant species within the 2-dimensional community morphospace,
115 with Voronoi polygons giving conservative estimates of the amount of morphological
116 space occupied by the vagrant species in relation to their “closest” native neighbours
117 (polygons do not allow for overlapping niches). Based on the position of the tropical
118 species within the convex hull of the resident community (table 1), their distance from
119 neighbouring species, and Voronoi polygon area, the potential to establish successful
120 populations in the future was classified (table 2). Since the receiving community
121 could be precisely defined, and was based on a single specific habitat type, the
122 position of vagrant species was considered in relation to the whole native fish
123 community, rather than just in relation to pre-assigned guild members (see approach
124 in Azzurro et al. 2014). This meant that consideration of vagrants’ potential to
125 establish a niche was not *a priori* constrained alongside species with an assumed

126 similar functional role. Calculations of convex hulls, Voronoi polygons and nearest-
127 neighbour Euclidean distances were carried out in MATLAB version 9.0.0.

128

129 **3. Results and Discussion**

130 Based on morphometric analysis, the species of tropical reef fishes with the greatest
131 probability of exploiting an available niche within the existing resident assemblage
132 are the butterflyfishes *Chaetodon auriga*, *C. flavirostris* and *C. vagabundus* and the
133 zanolid *Zanclus cornutus* (figure 1, table 2). The damselfishes *Abudefduf bengalensis*,
134 *A. sexfasciatus*, and *A. vaigiensis* are predicted to have medium probability of
135 establishing within the resident community, based on morphology and available niche
136 space (figure 1, table 2). Overall, the morphological position of vagrant species were
137 strongly skewed to the upper and lower left-hand corners of the convex hull.

138 Azzuro *et al.* [13] also found that species belonging to the family
139 Chaetodontidae had a high invasion probability in the Mediterranean, but argued that
140 habitat constraints (obligate association with coral of the species concerned) not
141 accounted for in the methodology meant that they would remain rare. The
142 butterflyfishes identified by our study as having strong range-shift potential actually
143 show versatility in their habitat associations, even within reef environments. *C.*
144 *auriga*, for example, can be observed in mangroves or rubble habitats, suggesting that
145 they may be able to exploit the rocky shoreline habitats of southeast Australia, as
146 predicted here. Additionally, they are not obligate coral feeders, exhibiting a more
147 generalist diet including algae and invertebrates. The potential impact of climate
148 change on the Australian temperate fish community could therefore be to expand
149 niche opportunities for algal crevice feeders (*Z. cornutus*), algal and invertebrate

150 feeders (*Chaetodon* spp.).

151 The most abundant tropical vagrant species currently found as juveniles along
152 the NSW coast is the damselfish *A. vaigiensis*. In our analysis this species is predicted
153 to have a moderate likelihood of establishing abundant populations within the resident
154 community, being located within, but in close proximity to, the boundary of the
155 convex hull of the resident assemblage. Its position lies in close proximity to the
156 resident species *Microcanthus strigatus* and its addition to the assemblage divides the
157 niche space of *M. strigatus* into a smaller polygon, suggesting potential competition.
158 Juvenile *A. vaigiensis* are commonly observed in close proximity to *M. strigatus*,
159 using similar habitat following the summer recruitment event (Smith pers. obs.)
160 suggesting strong overlap in ecological niche and highlighting the predictive power of
161 the morphological approach used here. Furthermore, the fact that the position of
162 tropical vagrant species considered in the current study is strongly skewed to the
163 upper and lower left-hand corners of the convex hull suggests that, in ecological
164 terms, the impending escalation in competition for niche space may impact
165 disproportionately on particular parts of the native community.

166 Geometric morphometrics, while a useful predictive tool, is based on the
167 assumption that competitive interactions between residents and invaders are driving
168 the likelihood of population establishment success. For our study system, the model
169 assumes that all commonly observed tropical fish that can be transported to Sydney
170 via the EAC have the potential to interact with temperate species and have equal
171 opportunity to use resources. The model then assesses this probability relative to
172 competitive exclusion due to morphological differences. It does not take into account
173 environmental drivers of range-shift success. Model predictions are also critically

174 related to the morphologies used to construct the morphospace. As Azzurro *et al.*
175 [13] point out, morphological peculiarities not captured by the analysis can prove to
176 be critical ecological novelties. For example the lionfish (*Pterois* spp.), predicted only
177 to have a moderate chance of invasion success within the defined morphospace [13],
178 has unfortunately become a highly successful invader in the Western Atlantic [25]
179 and Mediterranean [26].

180 The value of the geometric morphometric approach, however, lies in the
181 ability to generate hypotheses of range-expansion potential, which can be tested once
182 ocean temperatures reach over-wintering levels on a consistent basis. The growth of
183 citizen science projects such as the Range Extension Database and Mapping Project
184 (Redmap, <http://www.redmap.org.au>) and ongoing data collection by Reef Life
185 Survey (<http://reeflifesurvey.com/>) will be an avenue to ground-truthing the
186 predictions made here. For the temperate marine ecosystems of Australia's southeast
187 coast, Redmap's database of year-on-year numbers of adult-sized tropical vagrant
188 fishes will provide a means by which the predictions made by the current study can be
189 tested, once over-wintering by all vagrant species becomes a physiological reality.

190

191

192 **Ethics:** No ethics approvals were required for this study.

193 **Data accessibility:** Data deposited in Dryad: <http://dx.doi.org/10.5061/dryad.q0g60>

194 **Author contributions:** SS, RF, JD, DB designed the study. SS, MH and RF collected
195 data and performed the analyses, SS, RF, JD and DB interpreted the data. All authors
196 contributed to drafting of the manuscript, and subsequent edits made in response to

197 reviewer comments. All authors approved the final version for publication and agree
198 to be held accountable for the content therein.

199 **Acknowledgements:** We thank our two anonymous reviewers for their considered
200 and helpful comments.

201 **Competing interests:** The authors declare no competing interests.

202 **Funding:** This research was funded by the Ian Potter Foundation (JMD) and UTS
203 School of Life Sciences (SS).

204

205

206 **References**

207

208 1. Chen I-C, Hill JK, Ohlemüller R, Roy DB, Thomas CD. 2011 Rapid range shifts of
209 species associated with high levels of climate warming. *Science* **333**, 1024-1026.

210

211 2. Perry AL, Low PJ, Ellis JR, Reynolds JD. 2005 Climate change and distribution
212 shifts in marine fishes. *Science* **308**, 1912-1915

213

214 3. Pearman PB, Guisan A, Broennimann O, Radin CF. 2008 Niche dynamics in space
215 and time. *Trends Ecol. Evol.* **23**, 149-158.

216

217 4. Ehrlén J, Morris WF. 2015 Predicting changes in the distribution and abundance of
218 species under environmental change. *Ecol. Lett.* **18**, 303-314.

219

- 220 5. Parravicini V, Azzuro E, Kulbicki M, Belmaker J. 2015 Niche shift can impair the
221 ability to predict invasion risk in the marine realm: an illustration using
222 Mediterranean fish invaders. *Ecol. Lett.* **18**, 246-253.
223
- 224 6. Vergés A, Steinberg PD, Hay ME, Poore AGB, Campbell AH, Ballesteros E, Heck
225 KL, Booth DJ, Coleman MA, Feary DA et al. 2014 The tropicalization of temperate
226 marine ecosystems: climate-mediated changes in herbivory and community phase
227 shifts. *Proc. R. Soc. B* **281**, 20140846.
228
- 229 7. Hobday AJ, Pecl GT. 2014 Identification of global marine hotspots: sentinels for
230 change and vanguards for adaptation. *Rev. Fish Biol. Fish.* **24**, 415-425.
231
- 232 8. Sen Gupta A, Brown JN, Jourdain NC, van Sebille E, Ganachaud A, Vergés A.
233 2015 Episodic and non-uniform shifts of thermal habitats in a warming ocean. *Deep*
234 *Sea Res. II*, **113**, 59-72
235
- 236 9. Feary DA, Pratchett MS, Emslie M, Fowler AM, Figueira WF, Luiz OJ, Nakamura
237 Y, Booth DJ. 2014 Latitudinal shifts in coral reef fishes: why some species do and
238 others do not shift. *Fish Fish.* **15**, 593–615.
239
- 240 10. Ridgeway KR. 2007 Long-term trend and decadal variability of the southward
241 penetration of the East Australian Current. *Geophys Res Lett.* **34**, L13613.
242
- 243 11. Figueira WF, Booth DJ. 2010 Increasing ocean temperatures allow tropical fishes
244 to survive overwinter in temperate waters. *Glob. Change Biol.* **16**, 506-516.

245

246 12. Sunday JM, Pecl GT, Frusher S, Hobday AJ, Hill N, Holbrook NJ, Edgar GJ,
247 Stuart-Smith R, Barrett N, Wernberg T et al. 2015 Species traits and climate velocity
248 explain geographic range shifts in an ocean-warming hotspot. *Ecol. Lett.* **18**, 944-953.

249

250 13. Azzurro E, Tuset VM, Lombarte A, Maynou F, Simberloff D, Rodríguez-Pérez A,
251 Solé RV. 2014. External morphology explains the success of biological invasions
252 *Ecol. Lett.* **17**, 1455–1463.

253

254 14. MacArthur R, Levins R. 1967 The limiting similarity, convergence, and
255 divergence of coexisting species. *Am. Nat.* **101**, 377-385.

256

257 15. Pianka ER. 1974 Niche overlap and diffuse competition. *Proc. Nat. Acad. Sci.*
258 *USA* **71**, 2141-2145.

259

260 16. Farré M, Tuset VM, Maynou F, Recansens L, Lombarte A. 2013 Geometric
261 morphology as an alternative for measuring the diversity of fish assemblages. *Ecol*
262 *Indic.* **29**, 159-166.

263

264 17. Recasens L, Lombarte A, Sánchez P. 2006 Teleostean fish composition and
265 structure of an artificial reef and a natural rocky reef area in Catalonia (North Western
266 Mediterranean). *Bull. Mar. Sci.* **78**, 71-82.

267

268 18. Rohlf FJ. 2004 Department of Ecology and Evolution, State University of New
269 York, Stony Brook, NY. tpsDig 1.4 and tpsUtil.

270

271 19. Rohlf FJ. 2015. Department of Ecology and Evolution, State University of New
272 York, Stony Brook, NY. tps Relative Warps v1.60.

273

274 20. Smith S, Fox R, Donelson J, Head M, Booth D. 2016 Data from: Predicting range-
275 shift success potential for tropical marine fishes using external morphology. Dryad
276 Data Repository <http://dx.doi.org/10.5061/dryad.q0g60>.

277

278 21. Cornwell WK, Schwikl DW, Ackerly DD. 2006 A trait-based test for habitat
279 filtering: Convex hull volume. *Ecology* **87**, 1465-1471s

280

281 22. Barber CB, Dobkin DP, Huhdanpaa HT. 1996 The Quickhull algorithm for
282 convex hulls. *ACM Trans Mathematic Software* **22**, 469-483. <http://www.qhull.org>

283

284 23. Du F, Zhang XC, Sui Y, Shao M, Hu L, Shan L. 2012 The relationship between
285 aboveground biomass and Voronoi area of coexisting species in an old-field
286 community. *Pol. J. Ecol.* **60**, 479-489

287

288 24. Booth DJ, Figueira WF, Gregson MA, Brown L, Beretta G. 2007 Occurrence of
289 tropical fishes in temperate southeastern Australia: role of the East Australian
290 Current. *Est. Coastal. Shelf Sci.* **72**, 102-114.

291

292 25. Albins MA, Hixon MA. 2013 Worst case scenario: potential long-term effects of
293 invasive predatory lionfish (*Pterois volitans*) on Atlantic and Caribbean coral-reef
294 communities. *Environ. Biol. Fish.* **96**, 1151-1157.

295

296 26. Kletou D, Hall-Spencer JM, Kleitou P. 2016 A lionfish (*Pterois miles*) invasion

297 has begun in the Mediterranean Sea. *Mar. Biodiv. Rec.* **2016** 9:46.

298 **Table caption**

299 Table 1: Conditions of invader position within resident morphospace corresponding to
300 probability of invasion success.

301

302 Table 2: Predicted likelihood of range expansion by tropical vagrant fish species into
303 temperate marine communities of the southeast coast of Australia, based on
304 morphology alone.

305

306

307 **Figure Legend**

308 Figure 1: (a) Morphological position of species within the overall convex hull of a
309 representative present-day assemblage of bony fishes from the southeast coast of
310 Australia. Each dot represents a single species and outlines represent Voronoi
311 polygons defining ecological niche space surrounding each species. (b) Convex hull
312 of the present-day resident assemblage (dashed line) overlaid with assemblage
313 including seasonally-invading tropical vagrant fish species (solid line). Numbers
314 inside Voronoi polygons refer to species listing in table 2.