

# Southern Ocean phytoplankton physiology in a changing climate

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29 **Summary**

30 The Southern Ocean (SO) is a major sink for anthropogenic atmospheric carbon dioxide  
31 (CO<sub>2</sub>), potentially harbouring even greater potential for additional sequestration of CO<sub>2</sub>  
32 through enhanced phytoplankton productivity. In the SO, primary productivity is primarily  
33 driven by bottom up processes (physical and chemical conditions) which are spatially and  
34 temporally heterogeneous. Due to a paucity of trace metals (such as iron) and high variability  
35 in light, much of the SO is characterised by an ecological paradox of high macronutrient  
36 concentrations yet uncharacteristically low chlorophyll concentrations. It is expected that  
37 with increased anthropogenic CO<sub>2</sub> emissions and the coincident warming, the major physical  
38 and chemical process that govern the SO will alter, influencing the biological capacity and  
39 functioning of the ecosystem. This review focuses on the SO primary producers and the  
40 bottom up processes that underpin their health and productivity. It looks at the major physico-  
41 chemical drivers of change in the SO, and based on current physiological knowledge,  
42 explores how these changes will likely manifest in phytoplankton, specifically, what are the  
43 physiological changes and floristic shifts that are likely to ensue and how this may translate  
44 into changes in the carbon sink capacity, net primary productivity and functionality of the  
45 SO.

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49 **Keywords:** Southern Ocean; Phytoplankton; Marine Primary Productivity; Climate Change

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55 **Abbreviations:**

56 SO; Southern Ocean

57 SSIZ; Seasonal Sea Ice Zone

58 POOZ; Permanently Open Ocean Zone

59 MIZ; Marginal Ice Zone

60 MLD; Mixed Layer Depth

61 HNLC; High Nutrient Low Chlorophyll

62 SAM; Summer Annual Mode

63 ACC; Antarctic Circumpolar Current

64 PAR; Photosynthetically Active Radiation

65 UVR; Ultra Violet Radiation

66 CCM; Carbon Concentrating Mechanism

67 CA; Carbonic Anhydrase

68 PSI; Photosystem I

69 PSII; Photosystem II

70 NPQ; Non-photochemical quenching

71 SST; Sea Surface Temperature

72 OA; Ocean Acidification

73 POC; Particulate Organic Matter

74 NPP; Net Primary Productivity

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## 80 **1. Southern Ocean Primary Productivity**

81 The Southern Ocean (SO), defined as waters south of the Subtropical Front, is one of the  
82 major ocean biomes on the planet, making up 20% of the global ocean (Boyd, 2002). It is a  
83 dynamic marine ecosystem and plays an important role in the regulation of the Earth's  
84 climate. The strong easterly flow of the Antarctic Circumpolar Current (ACC) allows for  
85 oceanic exchange, thermohaline circulation and global heat exchange, as well as contributing  
86 to global CO<sub>2</sub> drawdown (Rintoul and Bullister, 1999), where it is estimated to absorb ~20 Gt  
87 of atmospheric CO<sub>2</sub> annually (Takahashi et al., 2002; Sabine et al., 2004). Phytoplankton are  
88 the link that couples atmospheric and oceanic processes. They are a highly diverse group of  
89 microscopic photosynthesising protists that inhabit the sunlit surface waters of the ocean and  
90 are key to the health and productivity of the marine ecosystem, influencing nutrient cycling,  
91 food web dynamics and global biogeochemical cycling (Buesseler, 1998; Garibotti et al.,  
92 2003a, 2003b). Global carbon biomass of phytoplankton equates to less than 1% of the total  
93 photoautotrophic biomass on the planet (Bryant, 2003), yet accounts for 40-50% of global  
94 carbon fixation (Field et al., 1998), with more than 100 million tonnes of inorganic carbon  
95 fixed by phytoplankton on a daily basis (Behrenfeld et al., 2006). This organic carbon is  
96 transferred through the food web, re-released to the atmosphere or sequestered to ocean  
97 depths, through a process known as the biological pump. The efficacy of the biological pump  
98 is a function of phytoplankton physiology and community structure, which are in turn  
99 governed by the physical and chemical conditions of the ocean. Given its pivotal role in  
100 atmospheric, oceanic and trophic processes, knowledge of SO phytoplankton photosynthesis  
101 and photosynthetic processes are vital to understanding variability in marine primary  
102 production under a changing climate.

103 In the SO, primary production is mediated by a range of physical, chemical and  
104 biological factors, namely micronutrients (especially iron) and light, mixed layer depth, sea  
105 ice retreat, CO<sub>2</sub> concentration, grazing and senescence. The system is also influenced by loss  
106 processes such as sinking, aggregation and viral infection. Together these factors divide the  
107 SO into distinct provinces (Fig. 1) and generate functional assemblages of microbes that are  
108 predictable in space and time (Boyd, 2002; Constable et al., 2014). The principal dichotomy  
109 defining these functional assemblages is the presence of sea ice. Approximately 60% of the  
110 SO remains ice-free throughout the year (the permanently open ocean zone, POOZ). Primary  
111 productivity in these high nutrient, low chlorophyll (HNLC) waters is commonly co-limited  
112 by light and iron availability (Boyd et al., 2007). The deep well-mixed waters carry cells out  
113 of their optimal light environment, limiting rates of biomass accumulation (e.g. Nelson and  
114 Smith, 1991; Strutton et al., 2000), while iron limitation restricts phytoplankton productivity  
115 and favours a relatively stable community of nano- and picoplanktonic phytoplankton (e.g.  
116 Ishikawa et al., 2002), however, large, lightly silicified diatoms can also occur (Davidson et  
117 al., 2010). There are some exceptions in the POOZ, such as the Antarctic Polar Front (PF), in  
118 the lee of islands or where bottom topography enriches surface waters with iron, that are able  
119 to sustain larger, often heavily silicified phytoplankton in relatively high concentrations  
120 (Sokolov and Rintoul, 2007).

121 The productivity of the remaining 40% of the SO (~19 million km<sup>2</sup>) is strongly  
122 influenced by seasonal ice cover (Gloerson et al., 1992). The areal extent of the sea ice (Fig.  
123 1) makes it one of the largest biomes on Earth (Massom and Stammerjohn, 2010) and the  
124 seasonal sea ice zone (SSIZ) supports 5 to 30% of the annual primary production (Legendre  
125 et al., 1992; Arrigo et al., 1998; Arrigo et al., 2010). It provides an environment in which  
126 algae thrive, with high concentrations growing on the underside of the sea ice forming a vital  
127 source of winter-spring nutrition to higher trophic levels (Arrigo and Thomas, 2004; Massom

128 and Stammerjohn, 2010). In addition, incorporation of algae into and upon the ice effectively  
129 provides a “life boat”, enabling algal cells to be retained in a maximally sunlit environment  
130 over winter (Wright et al., 2010). As summer approaches, increased phytoplankton  
131 production, together with dilution of the sea ice brine and the dissolution of carbonate  
132 crystals, results in the sea ice region becoming a net sink for CO<sub>2</sub> that can contribute  
133 approximately 58% of the net uptake by the SO (Delille et al., 2014).

134         Melting sea ice releases fresh, buoyant, iron enriched (winter accumulation of aeolian  
135 input), water and phytoplankton into the water column, initiating a phytoplankton bloom  
136 which retreats southward in the wake of the receding ice edge over the austral summer  
137 (Lannuzel et al., 2007; Constable et al., 2014). The meltwater stabilises the water column,  
138 shallowing the mixed layer depth (MLD) and entraining cells in a high light, high nutrient  
139 environment. The simultaneous release of microalgae (some active and some dormant) into  
140 the water column, are said to seed blooms in the marginal ice zone (MIZ) (Mangoni et al.,  
141 2009), however, a large proportion of the population sediments rapidly from the photic zone  
142 (Wright and van den Enden, 2000; Wright et al., 2010). At its maximum extent in December,  
143 the MIZ covers ~6 million km<sup>2</sup>, or approximately 39% of the SSIZ (Fitch and Moore, 2007).  
144 The magnitude, extent and timing of SSIZ blooms is profoundly affected by the timing of sea  
145 ice disappearance, prevailing wind speed and wave action (Fitch and Moore, 2007; Massom  
146 and Stammerjohn, 2010; Constable et al., 2014). As a result, only a relatively small  
147 proportion of the entire MIZ (17 - 24% in December and February, respectively) actually  
148 supports the development of phytoplankton blooms (Fitch and Moore, 2007). Despite this  
149 patchiness, MIZ blooms can contribute 25-67% of all the planktonic production in the SO  
150 (Smith and Nelson, 1986). They can extend over thousands of kilometres and can increase the  
151 phytoplankton biomass more than two orders of magnitude during November and December,  
152 before declining again in January (Smith et al., 2003, 2004; Moore and Abbott, 2000; Fitch

153 and Moore, 2007). This seasonal cycle is driven by the physical changes associated with ice  
154 retreat and its amplitude increases southward, with maximum biomass and productivity in  
155 coastal and shelf areas (Smith and Nelson, 1985; Arrigo and van Dijken, 2003; Arrigo et al.,  
156 2008a, 2008b; Westwood et al., 2010; Wright et al., 2010). The Western Antarctic Peninsula  
157 (WAP) has chlorophyll concentrations that can reach up to  $50 \mu\text{g L}^{-1}$  in waters off Palmer  
158 Station (Tortell et al., 2014; Goldman et al., 2015; Kranz et al., 2015; Young et al., 2015;  
159 PALTER database), while in East Antarctic waters maximum chlorophyll concentrations are  
160 commonly more than an order of magnitude less (Wright and van den Enden, 2000; Wright et  
161 al., 2010). However, while the magnitude of the blooms may differ, the successional  
162 sequence of the phytoplankton in East and West Antarctic waters appear to be similar (e.g.  
163 Wright and van den Enden, 2000; Garibotti et al., 2003a, 2003b, Wright et al., 2010).

164 The SO harbours a prolific and diverse protistan community dominated by diatoms,  
165 haptophytes and dinoflagellates, superimposed upon a background of pico- and  
166 nanophytoplankton (e.g. Wright et al., 2010; Wolf et al., 2013). The succession of  
167 environmental changes imposed on phytoplankton by recession of the sea ice in the SSIZ  
168 generates a predictable sequence of events from bloom formation, growth and production, to  
169 grazing, senescence and flux, and a predictable succession of phytoplankton functional  
170 assemblages and even taxa (e.g. Kang and Fryxell, 1993; Kang et al., 2001; Constable et al.,  
171 2014) (Fig. 2). During the winter, the sea ice hosts a diverse and metabolically active algal  
172 population dominated by psychrophilic diatoms (Stoecker et al., 2000; Kattner et al., 2004;  
173 Meiners et al., 2009; Petrou et al., 2010, 2011c). Satellite derived data, long-term databases  
174 and annual research cruises show that the retreat of sea ice initiates a bloom of diatoms in  
175 highly stratified waters and shelf areas (Garibotti et al., 2003a, 2003b; Alvain et al., 2008;  
176 Arrigo et al., 2008b). The spring diatom blooms deplete reserves of nutrients such as iron and  
177 silicate (e.g. Boyd, 2002) favouring blooms of the colonial, gelatinous *Phaeocystis*

178 *antarctica*, which are not so heavily grazed (Dennett et al., 2001; Wright et al., 2010). This  
179 leads to large expanses dominated by this haptophyte, particularly in the more deeply mixed  
180 and iron depleted areas. Senescence of this bloom, largely due to aggregation and sinking  
181 (DiTullio et al., 2000) leads to a community comprised of autotrophic nanoflagellates,  
182 including dinoflagellates, prasinophytes, cryptophytes, chlorophytes, chrysophytes and some  
183 diatoms (Fig. 2) (e.g. Peeken, 1997; Kang et al., 2001; Wright et al., 2010; Mills et al., 2012).

184 In general, due to the different in-water irradiance and nutrient dynamics, Antarctic  
185 coastal waters are more productive than the POOZ (Smith and Nelson, 1985; Arrigo et al.,  
186 2008a, 2008b; Wright et al., 2010). A spatial study using a mechanistic three-dimensional  
187 biogeochemical model of the Ross Sea, found Antarctic shelf waters to be a strong sink for  
188 CO<sub>2</sub>, due to high biological productivity, intense winds, high ventilation rates and extensive  
189 winter sea ice cover (Arrigo et al., 2008a). In contrast, a temporal study found substantial  
190 inter-annual variability in primary productivity across all regions of the SO between 1997 and  
191 2005, which appeared to be driven in large part by sea ice dynamics (Smith and Comiso,  
192 2008). Given the vast expanse and spatial and temporal variability in primary productivity  
193 across the entire SO, models will play an increasingly important role in helping predict future  
194 ocean conditions. However, for these models to formulate a complete understanding of SO  
195 primary productivity and changes to carbon sequestration, they will need to integrate the  
196 physiological complexity of the biota and the spatial dynamics of the annual sea ice cycle.

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## 198 2. **Climate-driven changes to the Southern Ocean**

199 Surface waters of the SO are expected to become warmer, fresher and more stratified with  
200 shallower MLD, reduced ice thickness, higher CO<sub>2</sub> concentrations, less sea ice and stronger  
201 westerly winds (Boyd et al., 2008; Constable et al., 2014). Together these spatially



202 heterogeneous changes will differentially affect phytoplankton productivity and carbon  
203 sequestration by altering ambient temperature, total irradiance and wavelength structure,  
204 nutrient availability and trophodynamics (Bopp et al., 2001; Boyd et al., 2008; Caron and  
205 Hutchins, 2013; Lewandowska et al., 2014). The effects of these changes on the distribution,  
206 composition and productivity of Antarctic marine microbes are poorly understood and  
207 difficult to predict (Marchant et al., 2001; Caron and Hutchins, 2013).

208         The SO is already experiencing the effects of climate change. Temperatures are  
209 increasing, ocean fronts are moving southward and there are region-specific changes in sea  
210 ice cover (Constable et al., 2014 and refs therein). Ozone depletion and greenhouse gasses  
211 have increased the strength of the Summer Annual Mode (SAM), which in its positive phase  
212 contracts the westerly wind belt southwards and increases their velocity. Ironically, these  
213 human-induced changes to the atmosphere have mitigated the southward penetration of  
214 global warming around Antarctica (Fyfe et al., 1999; Marshall, 2003; Turner et al., 2009;  
215 Swart and Fyfe, 2012). However, the continuing release of greenhouse gasses and the  
216 recovery of the ozone hole will inevitably cause future climate-induced warming and  
217 ecosystem change in Antarctic waters (Smetacek and Nicol, 2005; Boyd et al., 2008;  
218 Constable et al., 2014). Warming temperatures will unavoidably result in a decline in sea ice  
219 duration, extent and thickness (Constable et al., 2014 and refs therein). Sea ice plays a vital  
220 role in structuring the habitat and regulating the timing of biological production in the SO. It  
221 forms an essential feeding zone and breeding platform for higher trophic organisms,  
222 influencing recruitment, as well as reproductive and foraging behaviour for many species  
223 within the ecosystem and creates appropriate environmental conditions that foster  
224 phytoplankton blooms in the MIZ (Fig. 3). Decreased sea ice extent would restrict the  
225 productivity in the ice itself, reducing the drawdown of CO<sub>2</sub> and compromising the winter-  
226 spring nutrition for keystone species in the Antarctic ecosystem such as krill (Massom and

227 Stammerjohn, 2010; Delille et al., 2014). In addition, any reduction in sea ice area in the SO  
228 would limit its mediation of MLD, light climate, and micronutrient concentrations by sea ice  
229 melt; the very environmental changes that initiate the MIZ bloom and contribute much of the  
230 phytoplankton production that fuels the wealth of Antarctica life.

231 In the Arctic, the thinning and disappearance of sea ice has lengthened the  
232 phytoplankton growing period and has reportedly increased the net annual productivity  
233 (Arrigo et al., 2012; Bélanger et al., 2013) but this may not occur in Antarctic waters.  
234 Increasing temperature is predicted to strengthen stratification and shallow the MLD,  
235 reducing the flux of nutrients to surface waters (Bopp et al., 2001; Boyd and Doney, 2002;  
236 Boyd et al., 2008; Lewandowska et al., 2014). In the absence of enrichment of surface waters  
237 with iron from the melting ice, nutrient limitation is likely to constrain phytoplankton  
238 productivity (e.g. Boyd, 2002). Furthermore, nutrient limitation would foster phytoplankton  
239 communities typical of the POOZ, which are dominated by small flagellates (Ishikawa et al.,  
240 2002), grazed by microheterotrophs, which sink slowly and are likely to be remineralised in  
241 near-surface waters (Lewandowska et al., 2014), reducing carbon export. Conversely, Doney  
242 (2006) argued that increased stratification of the SO may increase phytoplankton production  
243 by reducing the deep mixing that often constrains their productivity. The net effect of these  
244 opposing processes in an increasingly ice-free SO is uncertain.

245 More than 90% of the extra heat energy from greenhouse warming is absorbed by the  
246 ocean (Stocker, 2015), raising the temperature of the ocean, particularly at high latitudes  
247 (Levitus et al., 2005; Domingues et al., 2008). Rising temperature could directly affect  
248 phytoplankton productivity by altering metabolic rates (Boyd et al., 2013; Lewandowska et  
249 al., 2014), as ambient water temperatures in the SO are commonly sub-optimal for the growth  
250 of Antarctic phytoplankton (Moisan et al., 2002). Thus, increasing sea surface temperatures  
251 are expected to increase photosynthesis and cause floristic shifts towards warm-water species,

252 while restricting cold-water species' ranges and reducing their present distribution and  
253 potentially the biodiversity of these cold water communities (Hays et al., 2005). Studies of  
254 Antarctic phytoplankton indicate that a rise in temperature alone causes only a modest  
255 increase in their growth rate (e.g. Rose et al., 2009; Boyd et al., 2013, 2015) and this increase  
256 is often overcompensated by the negative effects of increased temperature on abiotic and  
257 biotic stressors such as nutrient limitation and grazing (Caron and Hutchins, 2013;  
258 Lewandowska et al., 2014).

259         The increased velocity of the westerly winds as a result of the increase in positive  
260 SAM can enhance upwelling that enriches surface waters with nutrients. These nutrient-rich  
261 waters are driven northward by Ekman drift and promote productivity by diatoms south of the  
262 Polar Front (Lovenduski and Gruber, 2005). This agrees with studies using satellite  
263 observations to drive models of phytoplankton composition and abundance at synoptic spatial  
264 scales. These show that the increasingly positive SAM index correlated with an increase in  
265 diatom abundance (Alvain et al., 2013; Rousseaux and Gregg, 2015). Thus, SAM-induced  
266 upwelling of nutrient-rich waters may offset the effects of nutrient limitation due to the  
267 warming-induced increased stratification. Yet the increasing winds due to SAM have also  
268 increased eddy kinetic energy and the poleward movement of these eddies may mitigate the  
269 Ekman transport of nutrient rich water to the north (Lovenduski and Gruber, 2005). The net  
270 result of these opposing forces on phytoplankton in the SO is unclear, but the combined effect  
271 of increased upwelling and stratification may enhance phytoplankton production in waters  
272 between the Polar Front and SSIZ (Lovenduski and Gruber, 2005; Doney, 2006).

273         The concentration of hydrogen ions (protons) in the ocean has increased  
274 approximately 30% since the start of the industrial revolution and is predicted to increase by  
275 more than 100% by the end of this century (RCP 8.5; IPCC 2014). This is particularly  
276 concerning for the SO ecosystem, as the cold, high latitude waters of the SO are recognised

277 as the world's largest CO<sub>2</sub> sink (Sabine et al., 2004). Responses of marine microbes to  
278 elevated *p*CO<sub>2</sub> differ among studies. A number of studies conducted in polar waters indicate  
279 that moderate enhancement of CO<sub>2</sub> concentration promotes phytoplankton production and  
280 fosters the growth of large diatoms (Engel et al., 2008; Tortell et al., 2008a; Feng et al.,  
281 2010), while others find it promotes the growth of pico- and nanoplankton (Hare et al., 2007;  
282 Brussaard et al., 2013). Overall, differences in competitive fitness among phytoplankton  
283 functional groups indicate that exposure to elevated *p*CO<sub>2</sub> could alter the phytoplankton  
284 community in coming decades (Dutkiewicz et al., 2015). Responses by natural coastal  
285 communities of Antarctic marine microbes within and among seasons suggest that moderate  
286 increases in CO<sub>2</sub> concentrations may enhance phytoplankton productivity and growth, but by  
287 the end of this century CO<sub>2</sub> concentrations may have risen sufficiently that they could alter  
288 the species composition, reduce rates of biomass accumulation and enhance the relative  
289 abundance of small phytoplankton taxa (Davidson et al., 2016).

290         The complex interactions and multiple feedback mechanisms involved in  
291 biogeochemical and biological processes, makes characterising an ecosystem and predicting  
292 its response to environmental change extremely challenging. Commonly, manipulative  
293 studies expose individual taxa to single stressors, while in nature, phytoplankton are  
294 simultaneously exposed to a range of climate-induced changes in their environment and the  
295 responses are potentially mediated by interactions within and among trophic levels (Boyd,  
296 2013). The complexity of performing experiments that capture all the physical, chemical and  
297 biological changes encompassed above is bewildering to contemplate. Instead, Boyd et al.  
298 (2015) identified the key climate-induced stressors for a subantarctic location and exposed a  
299 diatom from this region to combinations of these stressors to simulate different climate  
300 scenarios. Results showed that different combinations of stressors elicited markedly different  
301 response. In general, simultaneous exposure to all the predicted stressors was beneficial to

302 growth and net productivity (biomass) (Boyd et al., 2015). Specifically, when nutrient-  
303 replete, diatom growth rates doubled under future (2100) conditions, primarily as a result of  
304 Fe enrichment and warming (Boyd et al., 2015). Such composite studies, while complex and  
305 logistically demanding to perform, are necessary to help predict future changes in  
306 productivity and carbon flux in the SO.

307

### 308 **3. SO phytoplankton physiology: responses to global change**

309 Characterising phytoplankton physiological responses to environment change is complex, as  
310 measurements of primary productivity require an understanding of multiple physiological  
311 processes including light harvesting capacity, electron transport, carbon fixation, nutrient  
312 requirement and utilisation, and the sensitivity of each of these processes to environmental  
313 perturbation. Phytoplankton diversity adds an additional layer of complexity since species-  
314 specific responses can differ substantially in their magnitude, speed and direction.

315 Over geological time, SO phytoplankton diversification has been driven by extreme  
316 variation in physico-chemical conditions. Large variations in light, temperature, nutrients (N,  
317 P, Si, Fe and CO<sub>2</sub>) and salinity occur spatially (from the sea ice to the open ocean) and  
318 seasonally (Fig. 3). In the sea ice, phytoplankton are incorporated into the ice matrix, and are  
319 exposed to freezing, hypersaline conditions within the brine channels (Thomas and  
320 Dieckmann, 2002) as well as low light or complete darkness for several months of the year.  
321 In the MIZ, productivity is generally high, due to the stable, nutrient rich, sun-drenched  
322 conditions and lower salinity from the melting ice (Fig. 3). This contrast strongly with the  
323 deep, well-mixed, iron-poor waters of the POOZ (Fig. 3).

324 Projected physical and chemical changes to the future ocean are relatively well  
325 characterised, and while the biological responses to these changes are less clearly defined,

326 our understanding of bottom up influences on marine productivity is much greater than that  
327 of top down affects (Constable et al., 2014). Indeed, while there are still many gaps to fill, the  
328 most detailed knowledge on changes to ocean productivity come from phytoplankton studies.  
329 Much less is known about how grazing and higher trophic dynamics (top down processes)  
330 will be affected by climate change and in turn how these altered physiologies and ecologies  
331 will influence the lower trophic processes and ultimately the net productivity of the  
332 ecosystem. In this section we present the current state of knowledge on the physiological  
333 responses of SO phytoplankton to the changes in ocean physico-chemical conditions (light,  
334 temperature, macronutrients, pH and Fe availability) projected under future climate.

335

### 336 *Light*

337 Light underpins all primary productivity, and as such, any environmental stress (from  
338 alterations in light and nutrient availability, or sudden changes in temperature or salinity)  
339 experienced by phytoplankton can manifest in the photophysiology of the organism. Thus  
340 knowledge of primary productivity and how it may change in the future requires an  
341 understanding of the balance between incoming excitation energy and outgoing demand for  
342 electrons used in carbon fixation (Critchley, 1997) and other photophysiological processes,  
343 such as photoacclimation.

344         Photoacclimation is a complex light response that results from changes in cellular  
345 activity and adjustment of physiological properties to balance photosynthetic electron  
346 transport with metabolic demands, such as nutrient uptake (Falkowski and LaRoche, 1991).  
347 Strategies of photoacclimation vary from one species to another, having a strong influence on  
348 species distribution and abundance through specialised niche differentiation. Differences in  
349 photoacclimation strategy are due to the heterogeneity in photosynthetic structure and  
350 accessory pigment composition between species, influencing the rate and quantity of photons

351 received by the photosystem (Wilhelm, 1990). Some species acclimate by changing the size  
352 of the light harvesting antenna of the individual reaction centre, while others increase the  
353 total number of reaction centres, keeping antenna size constant (Falkowski and LaRoche,  
354 1991; Moore et al., 2006). As with temperate diatoms (Ruban et al., 2004; Dimier et al.,  
355 2007; Lavaud et al., 2007), SO- and sea ice diatoms are highly plastic to fluctuations in light  
356 with a strong dependence on rapid and reversible xanthophyll cycling (Moisan et al., 1998;  
357 Moisan and Mitchell, 1999; Kropuenske et al., 2009; Petrou et al., 2011a, 2011c), however,  
358 not all species utilise the same strategies, nor have the same level of physiological plasticity  
359 (Kropuenske et al., 2009; Mills et al., 2010; Petrou et al., 2011a). This physiological  
360 variability results in a wide range of light utilisation efficiency and photoprotective capacities  
361 from one phytoplankton species to another, making it important to investigate species-  
362 specific photosynthetic activity to understand what drives community composition and  
363 ultimately primary productivity.

364         There have been relatively few studies into the photoacclimation and photoprotection  
365 in SO phytoplankton species. At the community level, it has been shown that SO  
366 phytoplankton under iron-limitation and high light stress induce high non-photochemical  
367 quenching (NPQ), with a strong dependence on xanthophyll cycling (Petrou et al., 2011b).  
368 Similarly, diatom dominated sea ice algal communities are highly plastic (Lizotte and  
369 Sullivan, 1991, 1992; Robinson et al., 1997; Petrou et al., 2011c), with rapid induction of  
370 NPQ under increased irradiance (Petrou et al., 2011c) and a high intrinsic resistance to  
371 photoinhibition (Petrou et al., 2010). Although they only possess modest photosynthetic  
372 protein repair rates, possibly as a result of low temperatures, by rapidly activating  
373 xanthophyll cycling, their repair rates are sufficient to keep pace with rates of  
374 photoinactivation at irradiances up to  $200 \mu\text{mol photons m}^{-2} \text{ s}^{-1}$ , which is more than ten times  
375 *in situ* light levels (Petrou et al., 2010).

376 Antarctic species successions and distribution have been investigated previously  
377 (Lizotte, 2001; Kopczyńska et al., 2007; Almandoz et al., 2008; Beans et al., 2008), with  
378 some studies correlating observed distribution with physical and chemical oceanic parameters  
379 (Almandoz et al., 2008; Beans et al., 2008). Only a few however, have linked a species  
380 physiology and photosynthetic plasticity to its observed distribution and abundance  
381 (Kropuenske et al., 2009, 2010; Petrou et al., 2011a, Petrou and Ralph, 2011). Comparison of  
382 the photoacclimation strategies of *F. cylindrus* and *P. antarctica* provides an excellent  
383 example of how the photoprotective capacity of these two important SO species can be linked  
384 with their ecological niche occupancy. *F. cylindrus* has been shown to utilise rapid and  
385 dynamic NPQ under a range of environmental conditions (Mock and Hoch, 2005;  
386 Kropuenske et al., 2009, 2010; Petrou et al., 2011a; Petrou and Ralph, 2012), balancing  
387 growth with light acquisition through the modulation of their carotenoid pigments (Petrou et  
388 al., 2011a; Petrou and Ralph, 2012). This species is highly resilient to changes in irradiance,  
389 possessing a large chlorophyll *a* specific optical cross sectional area (useful in low light), yet  
390 able to keep excitation pressure low under high light by rapid induction of xanthophyll-driven  
391 NPQ and regulation of the functional cross sectional area of PSII ( $\sigma_{\text{PSII}}$ ) (Kropuenske et al.,  
392 2010; Petrou et al., 2011a, Petrou and Ralph, 2012). This plasticity is not common to all SO  
393 diatoms, instead other species exposed to the same stressors, have shown to be less plastic to  
394 changes in light, exhibiting limited NPQ capacity and no dynamic control over excitation  
395 pressure (Petrou et al., 2011a; Petrou et al. 2014), highlighting how photoacclimation strategy  
396 may be an important determinant of the success of *F. cylindrus* in polar marine environments  
397 (Kang and Fryxell, 1992; Lizotte, 2001; Kopczyńska et al., 2007). As with *F. cylindrus*, the  
398 photoacclimation strategy of the prymnesiophyte *Phaeocystis antarctica* provides us with  
399 some insight into the success of this species, including its fast photoacclimation capabilities  
400 (Moisan et al., 1998; Kropuenske et al., 2009, 2010) and high investment into carotenoid



401 synthesis under Fe limitation (van Leeuwe and Stefels, 2007). While it appears that these two  
402 species utilise similar strategies, studies directly comparing *P. antarctica* with *F. cylindrus*  
403 found that the former was able to acclimate to changes in irradiance much faster than the  
404 diatom, but was much slower to induce NPQ, resulting in more photoinhibitory quenching  
405 (Kropuenske et al., 2009; Mills et al., 2010) and a high dependence on photosynthetic protein  
406 repair processes (Kropuenske et al., 2009, 2010). Linking their photosynthetic strategies with  
407 niche occupancy, the high photoprotective capacity, low susceptibility to photoinhibition and  
408 slower photoacclimation rates of *F. cylindrus* (Kropuenske et al., 2009, 2010; Petrou et al.,  
409 2011a), are well suited to the uniform growth conditions of the shallow mixed layer and sea  
410 ice, whereas the rapid photoacclimation capacity of *P. antarctica* are consistent with growth  
411 in deeply mixed water, where photosynthesis can be maximised during exposure in the sunlit  
412 surface waters and any photoinhibition incurred at the surface has time to repair when at  
413 depth (Kropuenske et al., 2009, 2010).

414         The extent to which ocean shoaling (reduced MLD) will influence SO primary  
415 productivity is dependent on the photophysiological strategy each organism employs. The  
416 predicted reduction in MLD means an increase in the integrated daily irradiance delivered to  
417 phytoplankton in the surface waters (Marinov et al., 2010). This trapping of phytoplankton in  
418 surface waters exposes them to higher irradiances of photosynthetic active radiation (PAR)  
419 and damaging, short wavelength ultraviolet radiation (UVR) (Marinov et al., 2010, Caron and  
420 Hutchins, 2013 and refs therein). High light and UVR are responsible for photoinhibition and  
421 cellular photodamage that can reduce rates of photosynthesis, growth and survival by  
422 phytoplankton (Davidson, 2006). Despite the recovery of the ozone layer, the effects of UVR  
423 exposure on SO phytoplankton as a result of ocean shoaling are likely to exceed that of  
424 previous ozone depletion (Williamson and Zagarese, 2003). However, this physiological  
425 stress may be countered by ocean warming, which has been shown to accelerate protein

426 synthesis (Boyd et al., 2015) or similarly, increased Fe bioavailability, which could enhance  
427 photosynthesis and quench reactive oxygen species. It is the interplay of different  
428 environmental conditions and the physiological responses they invoke, which makes  
429 predicting changes to primary productivity challenging. To increase knowledge of the  
430 generality of these responses, future research should focus on conducting photophysiological  
431 studies with an increased number of species. By comparing the responses of dominant species  
432 to environmental stressors, a more complete picture of ecosystem function in relation to  
433 species dominance and successions can be obtained. Furthermore, through the use of multi-  
434 species competition studies, a better understanding of potential community shifts and the  
435 implication for trophic dynamics and carbon cycling at a large scale can be obtained.

436

#### 437 *Temperature*

438 Temperature has recently been shown to be one of the strongest predictors of future SO  
439 diatom primary productivity (Boyd et al., 2015). However, the effect of warming on SO  
440 phytoplankton is complex and impossible to predict in isolation of other changing physico-  
441 chemical conditions. It is known that warming influences cellular processes such as growth,  
442 cell size, carbon and chlorophyll content as well as nutrient requirement in phytoplankton,  
443 however, in conjunction with other stressors, such as those that are likely to eventuate in a  
444 warmer world (nutrient availability, lowered pH and increased UVR stress), these responses  
445 could easily be steered in alternate directions.

446         Surface ocean temperatures control the distribution and abundance of phytoplankton  
447 in both space and time. Changes in the microbial community are difficult to quantify against  
448 a high background of temporal and spatial variability, but evidence indicates that  
449 dinoflagellate (Hallegraeff, 2010; McLeod et al., 2012) and *Emiliana huxleyi* (Cubillos et al.,  
450 2007) distributions are migrating poleward. Similarly, increased precipitation and glacial melt

451 from warmer temperatures reportedly favours dominance of cryptophytes over diatoms in  
452 Antarctic coastal waters (Moline and Prézelin, 1996; Moline et al., 2004). In addition, the  
453 timing, duration and magnitude of phytoplankton blooms is defined by the impact of  
454 temperature on the physical environment (e.g. MLD, light, nutrients) but the nature and  
455 magnitude of this effect differs among taxa (Sommer and Lengfellner, 2008; Sommer and  
456 Lewandowska, 2011). Temperature also affects the timing of phytoplankton blooms by  
457 triggering the formation and germination of phytoplankton resting stages (McQuoid and  
458 Hobson, 1995). Such changes could result in a temporal mismatch between phytoplankton  
459 and higher trophic levels with ramifications for trophodynamics, ecosystem productivity and  
460 carbon sequestration in the SO (e.g. Ji et al., 2010).

461 Antarctic phytoplankton species are well adapted to sub-zero temperatures (Morgan-  
462 Kiss et al., 2006), having evolved cryoprotectants (Raymond and Knight, 2003) and anti-  
463 freeze proteins (Janech et al., 2006). In general, they show higher growth rates and  
464 physiological activity than would be expected (Feller and Gerday, 2003; Morgan-Kiss et al.,  
465 2006; Young et al., 2015), which may be a form of cold adaptation, where the cell is able to  
466 overcome the slow enzymatic rates usually enforced by cold temperatures (Feller and  
467 Gerday, 2003; Morgan-Kiss et al., 2006).

468 Temperature response functions in computational models allow for reasonably  
469 accurate productivity measurements to be made based on satellite chlorophyll and sea surface  
470 temperature (SST) measurements (Behrenfeld and Falkowski, 1997). This phenomenon can  
471 be explained at the biochemical level, since higher temperatures can increase enzymatic  
472 turnover rates (and thus productivity) through increasing the activity of thermally sensitive  
473 enzymes (see Wagner et al., 2016) – at least until a maximum rate is reached. This behaviour  
474 is defined by the temperature coefficient  $Q_{10}$ , which describes a rate change per change of  $10^{\circ}$   
475 C in temperature. An average value of 2-3 for the  $Q_{10}$  of the carboxylating enzyme Ribulose-

476 1,5-bisphosphate carboxylase/oxygenase (RubisCO) has been measured in phytoplankton in  
477 general and recently verified for SO species (Descolas-Gros and de Billy, 1997; Devos et al.,  
478 1998; Sage, 2002; Young et al., 2015). Using this coefficient, if SST of the SO were to rise  
479 by 6°C, it would result in an approximate doubling of physiological activity (growth).  
480 However, while higher temperatures can have positive effects on metabolism, species have  
481 temperature maxima, beyond which a sharp drop in metabolic efficiency occurs (Boyd et al.,  
482 2013). In one study, a temperature increase of 3°C lead to a 25% increase in growth by a polar  
483 diatom, yet further increases in temperature resulted in a rapid decline (Boyd et al., 2013),  
484 clearly defining a growth rate thermal maximum.

485         Increased temperature can alter the balance between the rates of phytoplankton  
486 growth and grazing mortality, where the growth and grazing rates of microzooplankton  
487 increase more rapidly with temperature than growth of their phytoplanktonic prey (Rose and  
488 Caron, 2007; Chen et al., 2012; Caron and Hutchins, 2013). In addition, increases in ocean  
489 temperature coincide with a decline in the size of diatom frustules (Falkowski and Oliver,  
490 2007) making them too small to be efficiently captured by krill and favouring their  
491 consumption by microheterotrophs such as ciliates (Boyd et al., 1984; Kawaguchi et al.,  
492 1999; Caron and Hutchins, 2013 and refs therein). The resultant shift in the SO  
493 trophodynamics toward grazing by microzooplankton would reduce the transfer of  
494 phytoplankton productivity to higher trophic levels, again favouring the respiration of carbon  
495 substrates in surface waters (Wohlers et al., 2009). Furthermore, the resulting decline in the  
496 efficiency of the biological pump would have a positive feedback on global climate change.  
497 However, grazing and recycling of material in surface waters by the microbial loop may also  
498 limit the hypothesised stripping of iron from the upper water column by krill grazing (Wright  
499 et al., 2010) allowing the persistence of this vital nutrient to support recycled production.

500 As one of the strongest predictors of SO diatom productivity (Boyd et al., 2015),  
501 understanding the physiological and biochemical responses of key phytoplankton species and  
502 mixed communities to temperature is a key priority. However, as discussed in Wagner et al.  
503 (2016, this issue), one of the major challenges remaining is to find that key functional trait or  
504 traits that provide an invariable temperature-dependent fingerprint.

505

### 506 *Nutrients and macromolecular composition*

507 The SO exhibits strong heterogeneity in the distribution and concentration of nutrients. While  
508 the SO has some of the highest macronutrients (N, P, Si) concentrations of any marine  
509 province across the globe, Si and Fe are spatially and temporally variable (Martin et al., 1990;  
510 Coale et al., 2003, 2004; Hiscock et al. 2003; de Baar et al., 2005; Maldonado et al., 2005).  
511 This irregularity in nutrient availability greatly influences the spatial dynamics of biological  
512 communities and more specifically, the regulation of marine primary production (Hiscock et  
513 al., 2003). In phytoplankton, variations in nutrient ratios (elemental stoichiometry) will  
514 reflect their underlying allocation of energy into major molecules (RNA, cellulose, lipids,  
515 proteins etc) and are closely associated with key traits such as growth rate, and size  
516 (Klausmeier et al., 2008; Finkel et al., 2009). Therefore, alterations in elemental ratios can  
517 have large effects on rates and efficiencies of energy transfer and elemental cycling in the  
518 ocean. Furthermore, through their influence on cell metabolism, nutrient dynamics can alter  
519 competitive advantages and thus species dominance and community structure (Klausmeier et  
520 al., 2008).

521 The effect of nutrients on phytoplankton communities extends beyond changes in  
522 carbon productivity, as changes in nutrient availability, especially Fe, also influence cellular  
523 metabolism, which can lead to substantial shifts in carbon partitioning between  
524 macromolecular (proteins, lipid and carbohydrates) stores (Milligan and Harrison, 2000; Van

525 Oijen et al., 2004). Phytoplankton have been shown to alter their macromolecular  
526 composition to compensate for reduced RubisCO activity, energy storage and capture of  
527 transient nutrient pulses (Arbrust, 2009; Marchetti et al., 2009; Sackett et al., 2013, 2015;  
528 Young et al., 2015). Changes in the production of macromolecules by phytoplankton are  
529 fundamental to the health and functioning of the marine ecosystem, since the nutritional  
530 quality of the phytoplankton affects the efficiency of trophic energy transfer (Hessen et al.,  
531 2004). The cellular energy partitioning and macromolecular composition (lipids: proteins:  
532 carbohydrates) will determine the nutritional value of the phytoplankton and ultimately,  
533 through trophic transfer, the productivity of the ecosystem. Under sea ice conditions (sub-  
534 zero temperature, low light and high salinity), microalgae preferentially produce lipid  
535 macromolecules for energy storage (Mock & Kroon 2002; Sackett et al 2013). Lipids are the  
536 most energy-rich macromolecules, with approximately double the caloric value of proteins  
537 and carbohydrates (Whyte, 1987). Organisms such as juvenile krill depend on these rich  
538 sources of lipids for growth and productivity, particularly during the long, winter months  
539 (Falk-Petersen et al., 1998; Mock and Kroon, 2002; Lee et al., 2008). However, with  
540 predicted reductions in the duration, extent and thickness of sea ice expected with future  
541 warming, the physico-chemical conditions that currently drive this large allocation of energy  
542 into lipids stores, may alter, causing a shift in cellular resource allocation. A recent study  
543 conducted in the naturally Fe-rich waters near the Kerguelen Islands revealed that under high  
544 Fe availability, the dominant diatom species *Fragillariopsis kerguelensis* favoured the  
545 production of carbohydrate over lipid or protein stores (Sackett et al., 2014). This was in  
546 contrast to the less abundant diatom *Pseudo-nitzschia subcurvata*, which exhibited a more  
547 subtle response, whereby carbon partitioning remained relatively independent of iron  
548 availability (Sackett et al., 2014). Plasticity in macromolecular composition may be

549 advantageous to *F. kerguelensis* by allowing the species to switch from producing one type of  
550 molecule to another, depending on resource availability.

551         Phytoplankton are intimately involved in a number of biological feedbacks including  
552 the regulation of nutrient stoichiometry in the global ocean (Falkowski, 1998). Their  
553 influence over these elemental ratios is due in large part to the proportion of sinking cells that  
554 are broken down (re-mineralised) releasing their nutrients into the water. As a result, the  
555 nutrient ratio of the ocean often reflects the average nutrient ratio (i.e. the elemental  
556 stoichiometry) of the entire phytoplankton community (Arrigo, 2005). Given that each  
557 species can have a different stoichiometry, any shifts in community composition in response  
558 to climate change, could alter the nutrient budget in various oceanographic regions.  
559 Furthermore, since consumers are most successful (i.e. productive and healthy) when  
560 consuming phytoplankton of similar stoichiometry to themselves, it is likely that shifts in  
561 phytoplankton species (and thus stoichiometry) could result in changes in trophic transfer  
562 efficiency (Andersen et al., 2004). Nutrients are a key determinant in phytoplankton growth  
563 and carbon allocation and emerging research (see Wagner et al., 2016; Sackett et al., 2015)  
564 has shown that the macromolecular fingerprint of a cell can provide a ‘snapshot’ prediction  
565 of carbon efficiency (Sackett et al., 2015). This new insight and methodology delivers a  
566 promising biomarker for predicting growth and productivity, and is worthy of further  
567 investigation.

568

### 569 ***Ocean acidification – pH and CO<sub>2</sub>***

570 Photosynthetic carbon fixation by marine phytoplankton is constrained by low CO<sub>2</sub>  
571 concentrations present in ocean surface waters (<10-30 μM), caused by the low equilibrium  
572 concentrations and slow diffusion rates into the cell. Further restrictions to CO<sub>2</sub> uptake are  
573 imposed by the slow turnover rate and low affinity to CO<sub>2</sub> of RubisCO, which is central to

574 the Calvin cycle. To overcome these limitations and maintain high rates of photosynthesis,  
575 marine phytoplankton use a carbon concentrating mechanism (CCM, Reinfelder, 2011). The  
576 induction of a CCM involves the active uptake of CO<sub>2</sub> and/or bicarbonate (HCO<sub>3</sub><sup>-</sup>) and the  
577 activity of extra- and/or intracellular carbonic anhydrase (CA), an enzyme that catalyses the  
578 reversible dehydration of HCO<sub>3</sub><sup>-</sup> to CO<sub>2</sub> and vice versa. Extracellular CA is thought to sustain  
579 inorganic carbon uptake by generating CO<sub>2</sub> at the cell surface, a process that is of particular  
580 importance under low CO<sub>2</sub> conditions (Badger and Price, 1994; Hopkinson et al., 2013). As  
581 active uptake of HCO<sub>3</sub><sup>-</sup> and CO<sub>2</sub> results in the concentration of inorganic carbon inside cells  
582 in excess of that in seawater, the ability for phytoplankton to minimise the CO<sub>2</sub> loss out of the  
583 cell also represents an important component of the CCM (Rost et al., 2006).

584         The extent to which lowered pH and elevated CO<sub>2</sub> concentrations in seawater will  
585 alter carbon acquisition and photosynthetic carbon fixation in phytoplankton strongly  
586 depends on the physiological mechanisms of inorganic carbon uptake and intracellular  
587 assimilation. In response to lowered pH, studies have shown a down-regulation of the CCM,  
588 e.g. decreased uptake rates of CO<sub>2</sub> and HCO<sub>3</sub><sup>-</sup> as well as lowered external CA activities (see  
589 Reinfelder, 2011). Natural phytoplankton communities of the SO (Weddell Sea, Drake  
590 Passage, Western Antarctic Peninsula, Amundsen and Ross Sea) were found to actively take  
591 up CO<sub>2</sub> and HCO<sub>3</sub><sup>-</sup> (Cassar et al., 2004; Tortell et al., 2008a, 2008b, 2010; Neven et al., 2011;  
592 Tortell et al., 2013; Kranz et al., 2015; Trimborn et al., 2015), thus indicating the presence of  
593 a CCM. Also, laboratory studies have demonstrated the operation of CCMs in various SO  
594 diatoms (*Chaetoceros debilis*, *Fragilariopsis kerguelensis*, *F. cylindrus*, *Nitzschia frigida*,  
595 *Pseudo-nitzschia subcurvata*) and a prymnesiophyte (*Phaeocystis antarctica*), although  
596 important differences between species were seen in the ratios of HCO<sub>3</sub><sup>-</sup> and CO<sub>2</sub> uptake and  
597 external CA activity (Mitchell and Beardall, 1996; Trimborn et al., 2013; Kranz et al., 2015;  
598 Young et al., 2015).



599 Lowered pH and elevated CO<sub>2</sub> concentrations in seawater are considered to favour the  
600 growth of species that thrive under higher diffusive CO<sub>2</sub> uptake and have lower CO<sub>2</sub> leakage  
601 from the cell. Even though carbon fixation rates of SO phytoplankton species with highly  
602 efficient CCMs are already close to saturation under present-day CO<sub>2</sub> concentrations  
603 (Trimborn et al., 2013; Kranz et al., 2015; Young et al., 2015), lower energy expenditures, as  
604 well as optimised resource allocation resulting from CCM down-regulation, may enable  
605 enhanced carbon fixation rates and/or growth at lowered pH levels. Under these conditions,  
606 temperate diatom species, being characterised to operate highly efficient CCMs (Burkhardt et  
607 al., 2001; Roberts et al., 2007; Trimborn et al., 2008, 2009), were estimated to save about  
608 20% of the CCM-related energy expenditure and 3-6% of the energy expended on carbon  
609 fixation (Hopkinson et al., 2011). Among tested SO phytoplankton species, CO<sub>2</sub>-dependent  
610 responses on growth strongly vary. For one Antarctic sea ice diatom (*Nitzschia lecointei*),  
611 growth was slightly stimulated from ambient to low pH levels in short-term experiments  
612 (Torstensson et al., 2013), but long-term acclimation (~200 days) under these conditions  
613 resulted in a reduction in growth by 3-4% (Torstensson et al., 2015). However, growth and/or  
614 carbon fixation of other diatoms (*C. brevis*, *C. debilis*, *Rhizosolenia* cf. *alata*, *P. subcurvata*  
615 and *P. alata*) and *P. antarctica* remained unaffected by lowered pH (Riebesell et al., 1993;  
616 Boelen et al., 2011; Hoogstraten et al., 2012; Trimborn et al., 2013; Hoppe et al., 2015)  
617 suggesting that overall, the SO phytoplankton species investigated do not seem to benefit  
618 from lowered pH levels.

619 Natural phytoplankton assemblages of the SO can be susceptible to lowered pH, both in  
620 terms of community structure and productivity (Tortell et al., 2008a; Feng et al., 2010; Hoppe  
621 et al., 2013). Lowered pH promoted the growth of diatoms and a floristic shift from pennate  
622 *Pseudo-nitzschia* (Tortell et al., 2008a; Hoppe et al., 2013) or *Cylindrotheca* (Feng et al.,  
623 2010) dominated assemblages to communities dominated by *Chaetoceros* (Tortell et al.,

624 2008a; Feng et al., 2010) or *Fragilariopsis* (Hoppe et al., 2013). Although *Chaetoceros* did  
625 not dominate the assemblage in one of the studies (Hoppe et al., 2013), a 50% increase in its  
626 relative abundance was observed, pointing towards a pH-dependent stimulation of growth for  
627 this genus. These floristic shifts were also accompanied by an increase in carbon fixation  
628 (Tortell et al., 2008a; Hoppe et al., 2013), which could either be related to higher diffusive  
629 carbon supply or lowered energetic costs of carbon acquisition under lowered pH (Rost et al.,  
630 2008).

631 Energetic benefits resulting from lowered pH may be especially important under growth  
632 limiting conditions such as iron- or light-limitation as frequently observed in the SO  
633 (Timmermans et al., 2001; Feng et al., 2010; Petrou et al., 2011b; Strzepek et al., 2012). At  
634 present, only two studies on SO species have examined the physiological influence of pH and  
635 light in combination, finding that neither the exposure to limiting nor to saturating irradiance  
636 levels modulated pH-specific growth responses of the Antarctic diatoms *P. alata* (Hoogstraten  
637 et al., 2012) and *C. brevis* (Boelen et al., 2011). Similarly, under both constant and fluctuating  
638 light no significant pH-effects on growth were observed in *C. brevis* (Boelen et al., 2011) and  
639 *C. debilis* (Hoppe et al., 2015). In the latter study, however, dynamic light and lowered pH  
640 synergistically lowered carbon fixation, potentially due to higher metabolic costs for  
641 photoacclimation and PSII repair. Information on interactive effects of iron (Fe) and lowered  
642 pH is still very limited. In line with earlier field studies (Tortell et al., 2008a; Feng et al.,  
643 2010), Hoppe et al. (2013) reported that both Fe enrichment and lowered pH increased  
644 primary production and triggered a shift away from weakly to heavily silicified SO diatoms  
645 (*Pseudo-nitzschia* sp. and *Fragilariopsis* sp. at pH levels of 8.03 and 7.74, respectively).  
646 Under Fe-limitation, however, primary production remained unaltered at lowered pH while  
647 thinly silicified diatoms dominated irrespective of the applied pH level (*Pseudo-nitzschia* sp.  
648 and *Synedropsis* sp. at pH levels of 8.03 and 7.74, respectively, Hoppe et al., 2013).

649 Low Fe combined with high light conditions strongly affect Antarctic phytoplankton  
650 physiology (e.g. decreased growth, photosynthesis and carbon fixation (Timmermans et al.,  
651 2001; Feng et al., 2010; Petrou et al., 2011b; Strzepek et al., 2012), combined with lowered  
652 pH, these conditions could impose even stronger stresses for phytoplankton. In temperate  
653 phytoplankton, it has been shown that elevated pH potentially increases the degree of Fe-  
654 limitation (Shi et al., 2010), where the pH-dependent changes in Fe chemistry presumably  
655 influenced the observed shift in species composition (Hoppe et al., 2013). Natural  
656 phytoplankton assemblages from the Ross Sea responded strongly to changing irradiance and  
657 Fe and to a much lesser degree to elevated pH (Feng et al., 2010). However, a recent study on  
658 the interactive effects of pH, light, temperature and Fe on growth and primary productivity of  
659 *P. antarctica* and *F. cylindrus*, found that these species may actually increase their Fe  
660 requirement relative to carbon fixation under elevated pH and increased temperature  
661 conditions (Xu et al., 2014).

662 The last decade of research has resulted in considerable advances in our understanding  
663 of the physiological responses of phytoplankton to changes in seawater carbonate chemistry,  
664 including a number of mesocosm studies with natural marine communities. The varied results  
665 from these studies, however, highlight the complexity involved in understanding biological  
666 responses to OA. While logistically challenging, future research needs to extend the  
667 complexity of the studies and assess community changes in response to combined stressors, as  
668 low pH will not occur in isolation of the multiple changes predicted for the marine  
669 environment, nor will species respond in isolation of other species.

670

### 671 ***Iron (Fe) limitation***

672 In the modern ocean, Fe is rapidly oxidised from the soluble Ferrous iron (Fe<sub>(II)</sub>) to the highly  
673 insoluble Fe<sub>(III)</sub> and ferric hydroxide species, which are rapidly exported to depth by

674 aggregation and scavenging. In phytoplankton, Fe is an essential micronutrient for  
675 photosynthesis, required in Chlorophyll *a*, photosystem I (PSI), photosystem II (PSII),  
676 cytochrome *b<sub>6</sub>-f* complex, cytochrome *c<sub>6</sub>*, ferredoxin and Nicotinamide Adenine Dinucleotide  
677 Phosphate (NAD(P)H) dehydrogenase (Raven et al., 1999; Behrenfeld and Milligan, 2013).  
678 Additionally, Fe is involved in other key cellular processes such as respiration, macronutrient  
679 assimilation and detoxification of reactive oxygen species (Sunda, 1989; Morel et al., 1991;  
680 Sunda and Hutschmann, 1995). Due to their high demand for Fe, primary producers have  
681 developed specialised mechanisms to satisfy their needs; resulting in a decoupling between  
682 intracellular and dissolved Fe stoichiometry (Morel and Price, 2003; Moore et al. 2013), as  
683 well as complex interactions and feedbacks between Fe biology and its chemistry (Hassler et  
684 al., 2011a).

685 Over 30 years of research has demonstrated that Fe limitation controls the efficiency  
686 of the biological pump and the structure of phytoplankton communities in the SO (e.g.,  
687 Martin, 1990; Coale et al., 2003; de Baar et al., 2005; Maldonado et al., 2005; Blain et al.,  
688 2007; Boyd et al., 2007; Smetacek et al., 2012). However, only the biologically active pool  
689 (defined here as bioavailable) is the fraction that can be effectively taken up by  
690 microorganisms to support their nutrient-dependent metabolic processes and growth (Hassler  
691 et al., 2012). Enrichment of Fe can drive shifts in the community structure from a  
692 nanoplankton (<10 µm) to a microplankton (>10µm) dominated assemblage (de Baar et al.,  
693 2005), increasing the sinking of particulate organic carbon (POC) and thus export to the deep  
694 ocean (Boyd et al., 2000). As such, it is the bioavailable pool that shapes SO phytoplankton  
695 communities and caps the efficiency of the biological pump.

696 Across the SO, low bioavailability in Fe means that microorganisms are in  
697 competition to satisfy their biological requirement for growth. For that purpose,  
698 phytoplankton have evolved various strategies such as a reduction in cell size, induction of

699 high affinity transporters and overexpression of surface proteins or siderophores (Maldonado  
700 and Price, 1999; Trick and Wilhelm, 1995; Mioni et al., 2005) to acquire Fe and modulate  
701 their requirement. Microorganisms can also modulate their Fe biological requirement using  
702 enzyme replacement and modification of the photosynthetic antenna (Behrenfeld and  
703 Milligan, 2013; Petrou et al., 2014). Diatoms exhibit the highest sensitivity to Fe limitation  
704 (Miller et al., 1991; Morel et al., 1991) with shifts to larger sizes ( $> 10 \mu\text{m}$ ) in response to Fe  
705 fertilisation (de Baar et al., 2005). Experiments carried out with cultured diatoms in the  
706 laboratory showed a relationship between Fe, the surface/volume ratio (S/V) and the iron  
707 biological requirement for growth (de Baar et al., 2005; Timmermans et al., 2004, Sarthou et  
708 al., 2005), with larger diatoms being associated with greater iron requirement. Therefore, Fe  
709 bioavailability is not only influenced by its chemical forms, but also by the different uptake  
710 strategies, biological requirements and interactions of the phyto- and bacterio-plankton  
711 communities (e.g., Barbeau et al., 1996; Hutchins et al., 1999).

712         The amount of inorganic, bioavailable Fe in the SO is often insufficient to solely  
713 sustain phytoplankton growth (Hassler et al., 2012), as more than 99% of the dissolved Fe is  
714 associated with dissolved organic ligands (e.g., Boyd and Ellwood, 2010; Hassler et al.,  
715 2012). Fe-binding organic ligands are critical for Fe biogeochemistry, improving its solubility  
716 and affecting its reactivity to support phytoplankton growth (Maldonado et al., 2005; Hassler  
717 et al., 2011a, 2012). The distribution of Fe-binding organic ligands in the open ocean is  
718 compatible with multiple biological sources associated with Fe-stress and its  
719 recycling/remineralisation (Hunter and Boyd, 2007). However, despite their recognised  
720 importance, the production pathways, nature and binding mechanisms of in-situ Fe-binding  
721 ligands are mostly unknown (Gledhill and Buck, 2012).

722         The recycling of Fe (and iron-binding ligands) is rapid (hours to days) and occurs  
723 through grazing, phytoplankton lysis, bacterial and viral infections (Barbeau et al., 1996;

724 Poorvin et al., 2004; Strzepek et al., 2005), as well as active excretion and transformation of  
725 organic compounds (e.g., Trick and Wilhelm, 1995; Ogawa et al., 2001; Hassler et al.,  
726 2011b). In the remote SO where external sources of Fe are limited, Fe recycling is efficient,  
727 providing 20 to 100% of the Fe required to sustain phytoplankton growth (Hutchins et al.,  
728 1993; Poorvin et al., 2004; Strzepek et al., 2005; Sarthou et al., 2005). Remineralisation of  
729 particulate Fe at depth mediated by heterotrophic bacteria can also release Fe and ligands  
730 (Boyd et al., 2010), thus potentially affecting Fe bioavailability to primary producers.  
731 Moreover, microorganisms can affect Fe redox chemistry, either by direct reduction of an Fe  
732 organic complex (often resulting in Fe<sub>(II)</sub> complexes easier to dissociate; e.g., Shaked et al.,  
733 2005) or by the excretion of compounds such as superoxide (Kutska et al., 2005). Despite  
734 numerous studies, the parameters controlling the bioavailability of Fe to SO primary  
735 producers remains poorly understood (Boyd and Ellwood, 2010; Hassler et al., 2011a).

736         Considering the complexity of the feedback between Fe chemistry, biological  
737 responses to Fe limitation, the specificity of Fe biological requirement, and biological  
738 interactions at play, predicting changes to Fe bioavailability under future climate remains a  
739 challenging task. First, changes in Fe input rates in the SO could be related to changes in  
740 important current sources, including precipitation, atmospheric dust deposition and sea ice  
741 melting (e.g., Lizotte, 2001; Lannuzel et al., 2008; Moore and Braucher, 2008; Boyd and  
742 Ellwood, 2010). Secondly, warmer temperatures and ocean acidification (OA) will directly  
743 affect Fe chemistry, namely its solubility and strength of complexation with organic ligands  
744 (Millero, 2009; Breitbarth et al., 2010; Hassler et al., 2013; Gledhill et al., 2015). In artificial  
745 seawater (pH 8.1) Fe solubility decreased from 0.5 nM at 5 °C to 0.03 nM at 25 °C (Liu and  
746 Millero, 1999). In contrast, the ratio between the concentrations of inorganic Fe to Fe<sub>(III)</sub>  
747 increased by 1.16-fold for each 1 °C increment (Byrne et al., 1988). Therefore, it is unclear  
748 how an increase in temperature will affect Fe bioavailability. As oceans become more acidic,

749 both the hydroxide ( $\text{OH}^-$ ) and carbonate ( $\text{CO}_3^{2-}$ ) ions that form strong complexes with Fe, are  
750 expected to decrease, dropping by up to 80% by the end of the millennium (Millero, 2009).  
751 Therefore, a decrease in pH would contribute not only to an increase in the concentration of  
752 dissolved inorganic Fe,  $\text{Fe}_{(\text{II})}$  concentrations, and  $\text{Fe}_{(\text{II})}$  half-life times, but also to a weakening  
753 of Fe binding with ligands (Millero, 2009; Breitbarth et al., 2010, Gledhill et al., 2015), all of  
754 which should result in increased Fe bioavailability. However, to date, studies using diatom  
755 cultures and natural assemblages from the SO have shown that Fe bioavailability was  
756 decreased under the OA scenario (Shi et al., 2010; Hoppe et al., 2013; Sugie and Yoshimua,  
757 2013; Sugie et al., 2013). This indicates that, while changes to physico-chemical conditions  
758 in the future are reasonably well characterised for inorganic Fe species, it is not the case for  
759 in-situ organic species (Gledhil et al., 2015). Given that more than 99% of Fe is associated  
760 with loosely characterised organic ligands, many uncertainties remain with respect to  
761 understanding changes to Fe bioavailability based on its chemistry. Furthermore, both OA  
762 and temperature can affect biological growth and generation of organic ligands, potentially  
763 favouring growth of specific phytoplankton species and inducing a shift in the community  
764 structure unrelated to Fe (e.g., Tortell et al., 2008a; Trimborn et al., 2014). As different  
765 biological communities excrete different iron-binding ligands (Trick and Wilhelm, 1995;  
766 Maldonado et al., 2002; Hassler et al., 2015; Norman et al., 2015), a shift in community  
767 structure or changes in Fe-stress could have a greater impact on Fe chemistry and  
768 bioavailability than those associated with changes in temperature and pH (Gledhill et al.,  
769 2015). However, the effect that the modulation of the excretion of organic ligands associated  
770 with basal biological activity or the change in the nature of organic ligands associated with a  
771 shift in community will have on Fe chemistry and bioavailability is yet to be investigated.

772         The implication of changes in Fe bioavailability, (floristic shifts and primary  
773 productivity) is highly dependent on the changes in co-varying environmental parameters, all

774 of which will likely influence oceanic carbon recycling and export (Shi et al., 2010; Hoppe et  
775 al., 2013). Fe limitation and OA can alter the extent of diatom silicification (Frank et al.,  
776 2000) and thus their contribution to carbon export (Hoppe et al., 2013). Similarly, light is also  
777 known to affect Fe limitation (Timmermans et al., 2001; Petrou et al., 2011b, 2014) and  
778 different taxa could be controlled by other trace elements (e.g., Zn and Co for *Phaeocystis*  
779 *antarctica*; Saito and Goepfert, 2008), suggesting that multiple “stressors” need to be  
780 simultaneously accounted for in order to predict future changes in Fe bioavailability and  
781 primary productivity, something that has only recently started to be investigated (Boyd et al.,  
782 2015).

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#### 784 **4. Future of SO primary productivity?**

785 There is ample evidence that climate change will shape the function and ecology of  
786 phytoplankton of the future, influencing their composition, abundance, distribution,  
787 phenology and physiology (Falkowski et al., 2004; Strzepek and Harrison, 2004) and that this  
788 process has already begun (Polovina et al., 2008; Doney et al., 2009; Constable et al., 2014).  
789 Yet the magnitude, and even the nature, of those changes remain unclear. In the SO,  
790 projected changes in nutrient supply and light climate, as a result of stratification, will likely  
791 result in regional changes in net primary productivity (NPP), where previously light-limited  
792 areas could see an increase in NPP (Doney, 2006; Arrigo and Thomas, 2004), whereas in  
793 nutrient-limited areas, NPP may decline (Arrigo and Thomas, 2004). Of course, these  
794 simplistic projections are made more complex by the biota and its physiological response to  
795 altered conditions, as any resultant shift in species composition will not only influence  
796 productivity, but also trophodynamics and biogeochemistry. If communities shift from  
797 diatom dominated to non-diatom dominated populations, an increase in primary production  
798 rates may occur, as nutrient utilisation efficiency and carbon fixation is generally lower in



799 diatoms (Arrigo et al., 1999). However, smaller non-silicified cells may have lower sinking  
800 rates and therefore weaken the biological pump and reduce carbon sequestration. In addition,  
801 the effect of OA brings an added complexity to understanding the way communities will shift  
802 and respond and how this will influence food webs and carbon sequestration.

803         Ocean warming and the subsequent diminishing of the MLD will influence SO  
804 primary productivity in the future, but the extent and direction of this change is still unclear.  
805 Based on the existing knowledge of predicted climate forcing and phytoplankton physiology,  
806 two opposing scenarios can be derived (Fig. 4): In the first scenario, the shoaling of the ocean  
807 leads to a reduction in macronutrient input from the deep and there is a diminishing supply of  
808 bioavailable Fe, either from the loss in sea ice as a source or from the effects of warming on  
809 Fe solubility (Liu and Millero, 1999). Under these conditions, phytoplankton are trapped in  
810 warm, macro- and micronutrient-poor surface waters (Fig. 4a) where they are exposed to high  
811 light and damaging UV radiation (Marinov et al., 2010), likely reducing rates of  
812 phytoplankton photosynthesis, growth and NPP. These future conditions also favour  
813 phytoplankton communities typical of the POOZ, which contribute little to POC export,  
814 weakening the biological pump (Fig. 4a). In the second scenario (Fig. 4b), climate-induced  
815 physical and chemical changes to the SO increase the bioavailability of Fe due to warming  
816 (Byrne et al., 1988) and lowered pH (Millero, 2009; Breitbarth et al., 2010, Gledhill et al.,  
817 2015) and favour the increase in the prevalence of large diatoms. In addition, the  
818 intensification of westerly winds, due to the increasing positive SAM index (Fig. 4b), drive  
819 upwelling and thus nutrient supply to the surface waters, alleviating nutrient-limitation and  
820 increasing diatom abundance (Alvain et al., 2013; Rousseaux and Gregg, 2015). In  
821 combination, such conditions enhance NPP in the SO, through faster growth rates and  
822 minimal floristic shifts, maintaining high POC export and efficient ocean biogeochemical  
823 cycling.

824 To date, nearly all phytoplankton physiological studies have constrained their  
825 manipulations to one or two variables, providing information on a species' response to  
826 individual or dual environmental stressors. These studies, while valuable, lack the validity  
827 and relevance to assess global change at the ecological scale, as they do not test the effects of  
828 multiple stressors, nor indeed, multiple species. The inherent spatial heterogeneity in climate-  
829 mediated changes to physico-chemical conditions of the ocean and the specificity of species'  
830 physiological responses to individual and combined stressors, makes understanding how  
831 climate change will influence SO phytoplankton physiology, community structure,  
832 biogeochemistry, NPP and ultimately the efficacy of the SO as a carbon sink, an ongoing  
833 challenge.

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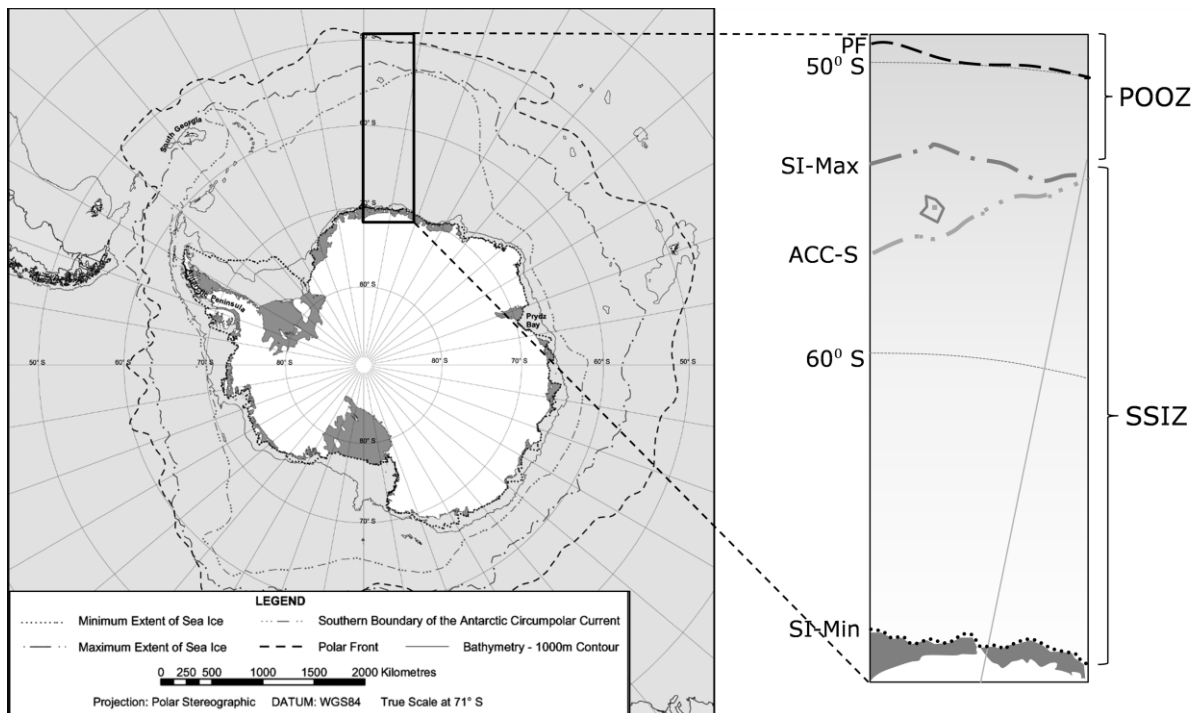
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1524 **Figures:**

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1528 **Figure 1:** The Southern Ocean oceanographic zones, south of the Polar Front (PF; ---).  
1529 Stippled lines indicate the maximum extent of sea ice (SI-Max; · — · — ·), the minimum  
1530 extent of sea ice (SI-Min; ·····) and the southern boundary of the Antarctic Circumpolar  
1531 Current (ACC; ··· — ···). Magnification of marked area (RHS), shows the Seasonal Sea Ice  
1532 Zone (SSIZ) and Permanently Open Ocean Zone (POOZ) for the Atlantic sector in relation to  
1533 the oceanographic zones and latitude. Map (LHS) courtesy of the Australian Antarctic  
1534 Division. © Commonwealth of Australia 2004.

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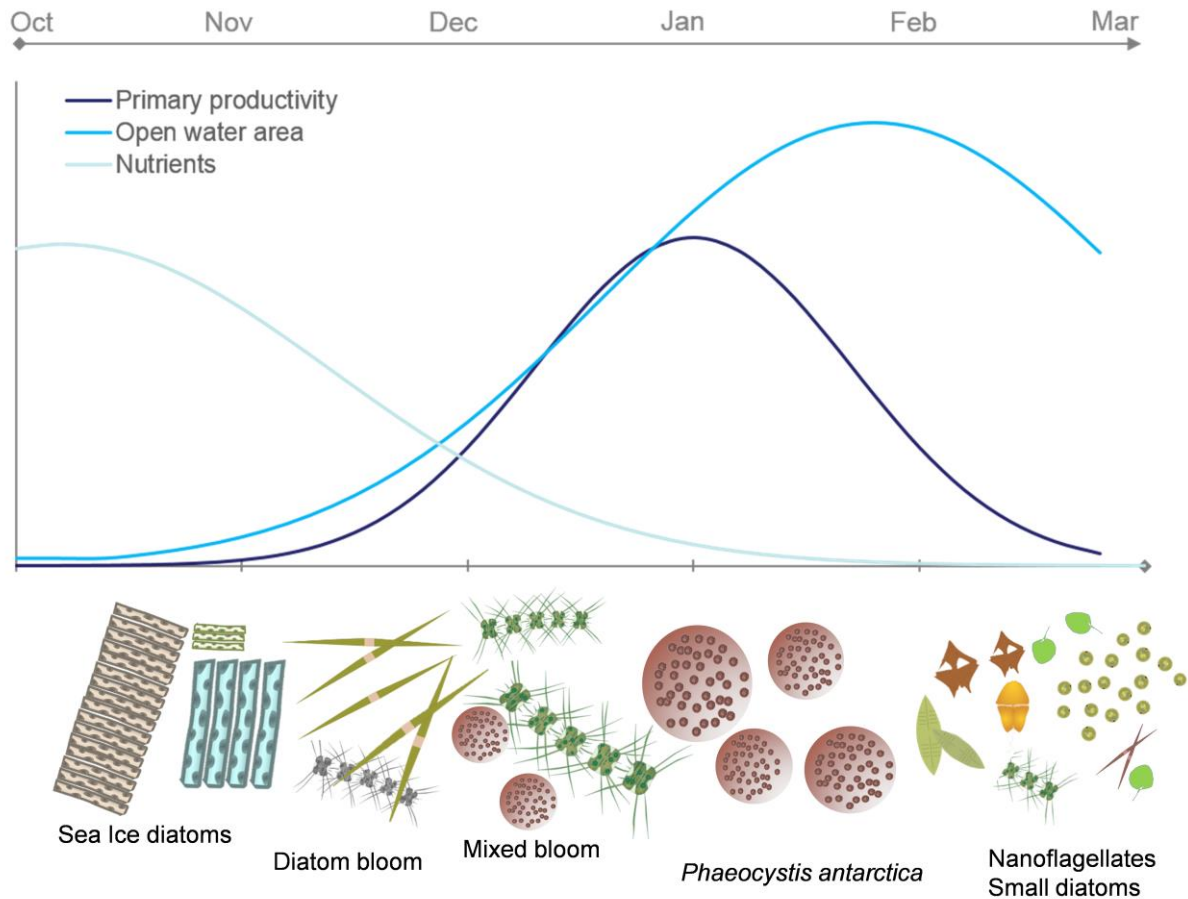
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1543 **Figure 2:** Seasonal changes in open water area, phytoplankton primary productivity and  
 1544 nutrient concentration in the SSIZ from October to March, where the retreat of the sea ice and  
 1545 rapid nutrient drawdown drive the observed succession of Antarctic phytoplankton  
 1546 communities. Timing and magnitude of changes in SSIZ are adapted from Arrigo et al.,  
 1547 (2008b) and species succession information taken from Kang et al. (2001), Garibotti et al.  
 1548 (2003a, 2003b) and Wright et al. (2010).

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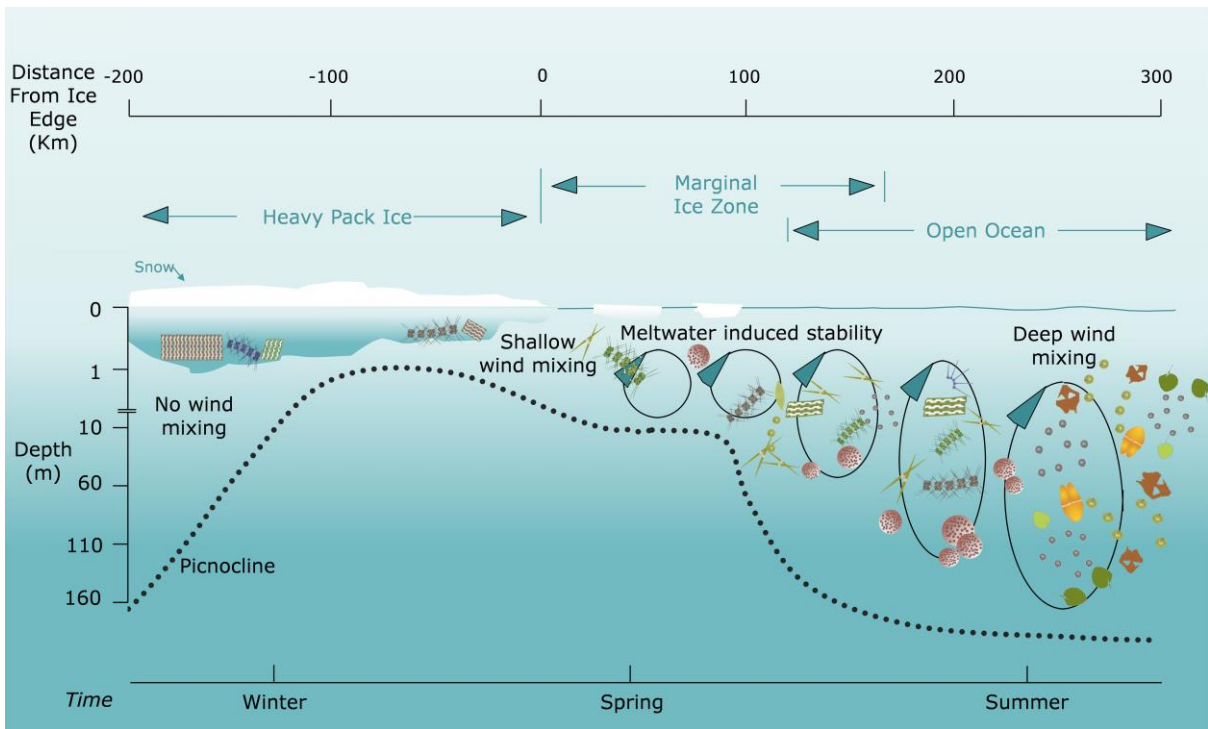
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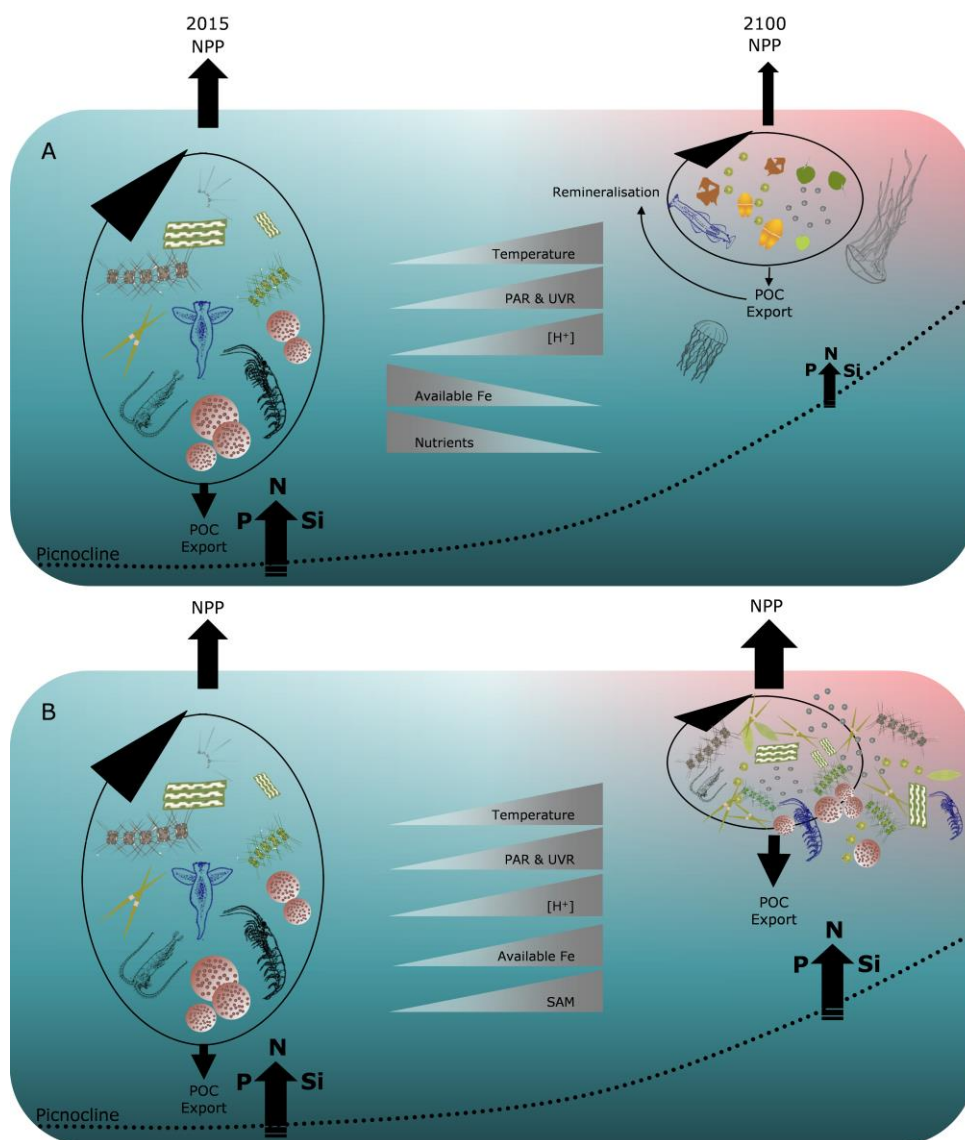
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1556 **Figure 3:** Schematic of the depth integrated environmental conditions found in the Pack Ice,  
 1557 Marginal Ice Zone and Open Ocean. The horizontal axes shows how conditions change  
 1558 spatially (top X-axis) and temporally (bottom X-axis) in the Antarctic marine ecosystem and  
 1559 its influence on phytoplankton communities.

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1564 **Figure 4:** Cartoon illustrating possible scenarios for SO net primary productivity (NPP)  
 1565 today and at the end of the twenty-first century (2100). A) Ocean shoaling leads to reduced  
 1566 nutrient input, changing species composition and trapping phytoplankton in warm, nutrient-  
 1567 poor, high UVR waters. These conditions result in large floristic shifts to smaller, flagellate  
 1568 species, reducing POC export and high rates of damage to photosynthesis from UVR  
 1569 exposure, reducing overall NPP. B) Ocean shoaling is diminished by increased positive SAM  
 1570 index, enhancing upwelling and delivering nutrients to well-lit, surface waters, promoting  
 1571 phytoplankton growth and boosting NPP. Increased Fe bioavailability maintains a high  
 1572 diatom diversity (minimal floristic shift) and silicification, resulting in high POC export to  
 1573 the deep ocean. In both scenarios the increase in hydrogen ions [H<sup>+</sup>] results in a loss of  
 1574 calcifying organisms.

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