

1 **Land use transition and ecological consequences: a spatiotemporal analysis**  
2 **in south-eastern Bangladesh**

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## 30 **Abstract**

31 Urban and economic growth around the world have rapidly degraded the ecosystem and  
32 environmental quality. Chattogram, a renowned city in Bangladesh known for its scenic  
33 environment, has experienced significant urban growth, which has degraded the environmental  
34 and ecological quality to the greatest extent. To investigate this issue, we assessed the  
35 Chattogram Metropolitan Area (CMA)'s ecological quality using the remote sensing-based  
36 ecological index (RSEI), incorporating four ecological indices such as heat index (LST),  
37 wetness (WET), dryness (NDBSI), and vegetation index (NDVI) through principal component  
38 analysis (PCA) and random forest (RF) model to analyze land use transitions (LUTs) using an  
39 advanced google earth engine (GEE). The PCA1 accounted for 80% of the dataset, delving into  
40 the highly significant features. Land use and land cover classes (LULCs) showed a gradual  
41 expansion of urban areas over the last two decades, indicating socio-economic progress. The  
42 urban area expanded from 34.08 km<sup>2</sup> in 2003 to 101.99 km<sup>2</sup> in 2023, an increase of 67.91 km<sup>2</sup>  
43 over 20 years, with an annual expansion rate of 2.34 km<sup>2</sup>/year. The RSEI values remained  
44 consistently >0.6, indicating good ecological levels, primarily due to sustainable urban  
45 planning. RSEI negatively correlated with LST and NDBSI, suggesting a detrimental  
46 influence, particularly in urban and economic zones. Conversely, the NDVI and WET indices  
47 showed a positive correlation, indicating a constructive impact on the local ecological  
48 landscape. This study will assist policymakers in formulating strategies for ecological  
49 conservation to achieve sustainable development goals (SDGs) and implement the Chattogram  
50 Metropolitan Master Plan (2020-2041).

51 **Keywords:** *Bangladesh, degradation, ecology, GIS, modelling, RSEI, urban growth.*

52

## 53 **1. Introduction**

54 Land is the fundamental component of Earth, playing a crucial role in maintaining ecological  
55 processes and developing terrestrial ecosystems, which encompass complex interactions  
56 between living organisms and their environment (Ma and Zhang 2023; McCourt *et al.*, 2023;  
57 Wei *et al.*, 2023). Ecosystem services play a significant role in the well-being, socio-economic  
58 and environmental progress of humans (Hariram *et al.*, 2023; Sangha *et al.*, 2022). However,  
59 human activities have drastically destroyed the suitable land use patterns, known as land use  
60 transitions (LUTs), resulting in significant impacts on ecosystem services (Yin *et al.*, 2020). As  
61 a result, critical ecological zones, such as vegetation and water bodies, have experienced

62 progressive depletion and fragmentation (Tolessa *et al.*, 2017). In recent years, urban expansion  
63 has changed land use patterns and structures (Long *et al.*, 2021) and led to various ecological  
64 issues, such as air pollution, deteriorating soil quality, an increased urban heat island, and a loss  
65 of biodiversity (Ullah *et al.*, 2022). From 1997 to 2011, estimates of the projected annual loss  
66 in ecosystem service value (ESV) due to changes in land use ranged from 4.3 to 20.2 trillion  
67 USD globally (Costanza *et al.*, 2014; Yin *et al.*, 2020). Investigating, analyzing, and  
68 comprehending the processes of LULC is imperative for devising solutions to achieve a  
69 sustainable ecosystem (Airiken *et al.*, 2022; Mao *et al.*, 2023). Numerous researchers have  
70 employed various remote sensing indices to evaluate diverse ecosystems, including urban  
71 areas, forests, grasslands, and watersheds. The normalized difference vegetation index (NDVI)  
72 evaluates changes in forest ecosystems (Prävālie *et al.*, 2022) while the normalized difference  
73 impervious surface index (NDISI) assesses urban ecological environments (Li *et al.*, 2021).  
74 The chlorophyll-a and suspended-solid indices analyze water quality, whereas land surface  
75 temperature (LST) assesses the influence of urban heat islands (Sekertekin and Zadbagher,  
76 2021). Two primary ecological indicators, the ecological index (EI) (Karbalaei Saleh *et al.*,  
77 2021) and the RSEI (Yuan *et al.*, 2021), are utilized to evaluate regional ecological conditions,  
78 whereas RSEI is superior to EI, demonstrating greater effectiveness in monitoring eco-  
79 environmental quality (EEQ) (Mao *et al.*, 2023). RSEI, which is based solely on remote  
80 sensing, incorporates four key ecological indicators: heat index, wetness, vegetation index, and  
81 dryness. Numerous studies currently assess the ecological environment at various levels,  
82 ranging from global to intercontinental (Lu *et al.*, 2019), national (Liao and Jiang, 2020), urban  
83 agglomeration (Ji *et al.*, 2020), and city levels (Lu *et al.*, 2023).

84 Bangladesh is a developing country with a substantial population (Rahman *et al.*, 2023; Soma  
85 *et al.*, 2023), which often leads to alterations in the natural landscape or land use (Rahaman *et al.*,  
86 2023). Chattogram, situated in southeastern coastal Bangladesh, stands as one of the  
87 nation's most renowned and industrially productive cities, often referred to as the commercial  
88 capital and the gateway to Bangladesh (Imran *et al.*, 2022). The rapid urbanization aimed at  
89 boosting economic conditions has impacted the local ecology to varying extents, rendering it  
90 more vulnerable to land degradation (Hasan *et al.*, 2020). There is a significant expansion in  
91 urban areas as more individuals move to metropolitan regions in search of employment  
92 (Chakma *et al.*, 2023; Roy *et al.*, 2020). Several studies have explored the correlation between  
93 LUTs and Summer LST in Chattogram City (Miah *et al.*, 2023). Some studies have focused on  
94 exploring several heat indices to understand the impact of LUTs on LST in the study area (Kafy

95 *et al.*, 2020; Roy *et al.*, 2020) and predicted the LULC and seasonal LST for Chattogram in  
96 2029 and 2039. Furthermore, the heat-vulnerable areas in CMA have also been mapped in  
97 recent years (Raja *et al.*, 2021).

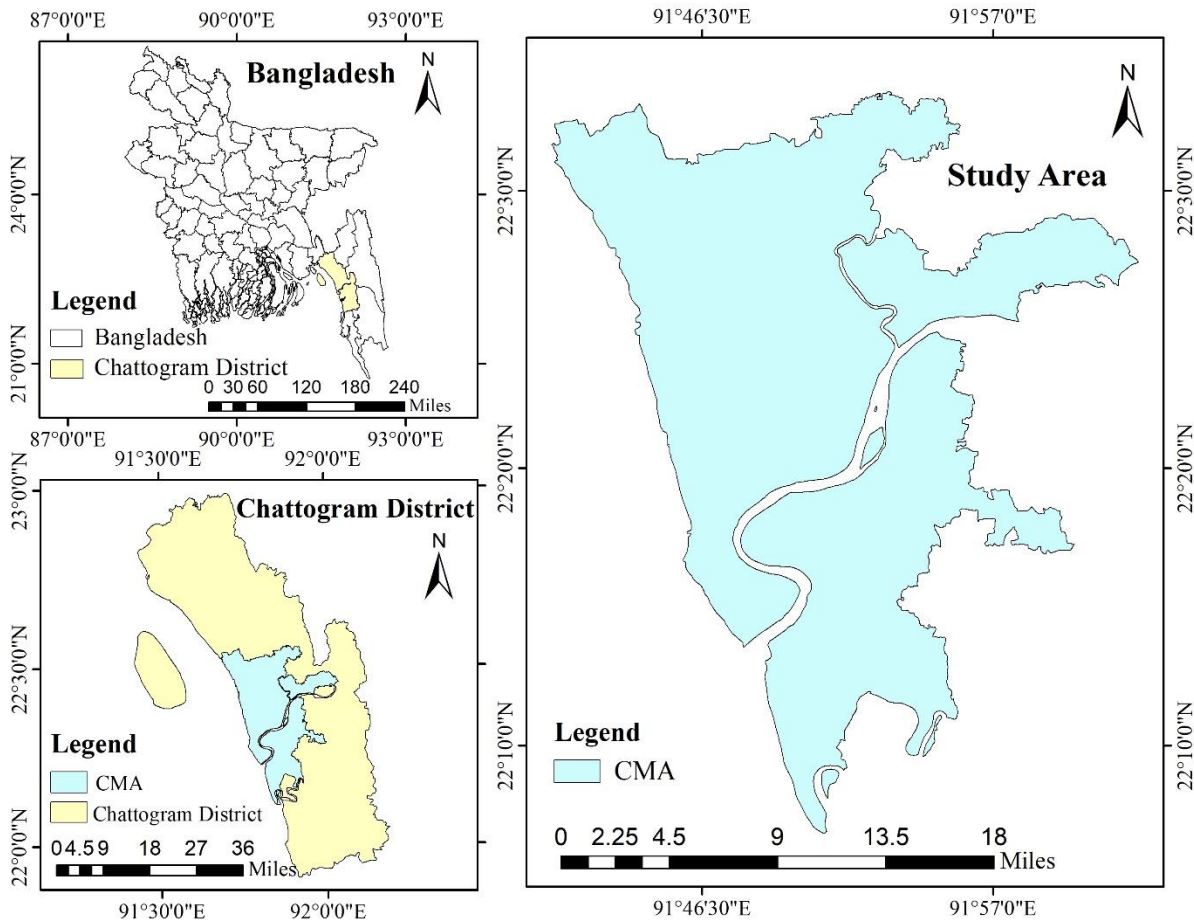
98 However, comprehensive ecological assessments of CMA to understand its transformations  
99 and the impact of LUTs on ecological quality are lacking, both on small and large scales.  
100 Additionally, the existing literature does not provide a clear explanation for the relationship  
101 between the numerous indices examined. This study focuses on the spatiotemporal changes in  
102 EEQ in the CMA with three distinct aims. In the beginning, we used four specific indices,  
103 specifically NDVI, LST, NDBSI, and WET, to conduct a comprehensive spatiotemporal  
104 assessment of ecological quality within the CMA. Then we delved into the exploration of the  
105 impact of land use practices on the area's ecological health. Finally, we examined the  
106 interrelationships between these indices to understand their combined effectiveness in  
107 assessing ecological quality. This paper is innovative for two main reasons. Firstly, while most  
108 previous assessments in Bangladesh relied on unreliable surveys, our study used the RSEI with  
109 Landsat imagery data to evaluate environmental conditions and changing patterns. Secondly,  
110 we leveraged the powerful cloud computing capabilities of GEE to establish a consistent time  
111 threshold for the periods studied. Additionally, we enhanced the quality of the original images  
112 using cloud removal and image median equalization techniques. This comprehensive study is  
113 the first of its kind in CMA as well as in Bangladesh which will aid policymakers in formulating  
114 strategies for ecological and environmental conservation to the greatest extent.

## 115 **2. Materials and Methods**

### 116 **2.1 Study Area**

117 The CMA is situated in the southeastern part of Bangladesh (Fig 1). This area lies between  
118 approximately 22°06' and 22°34' north latitude and 91°40' and 92°2' east longitude, covering  
119 approximately 720 km<sup>2</sup>. CMA is geographically bordered by the Halda River and the Karnafuli  
120 River to the northeast and southwest, respectively, while the Rangamati district to the east and  
121 the Bay of Bengal to the west (Rahman *et al.*, 2017). The region experiences an annual average  
122 temperature varying from 13.5 °C to 32.5 °C, with a mean annual rainfall of 3378 mm (Ahmed  
123 *et al.*, 2018). CMA is characterized by a varied terrain, ranging from coastal plains to hilly  
124 tracts, influencing local climatic patterns and land use dynamics. The landscape features a mix  
125 of natural ecosystems, including coastal belts, low-lying wetlands, and hills covered with

126 vegetation. Urbanized zones comprise a significant portion of the area, with densely populated  
127 neighborhoods, industrial estates, and commercial hubs.



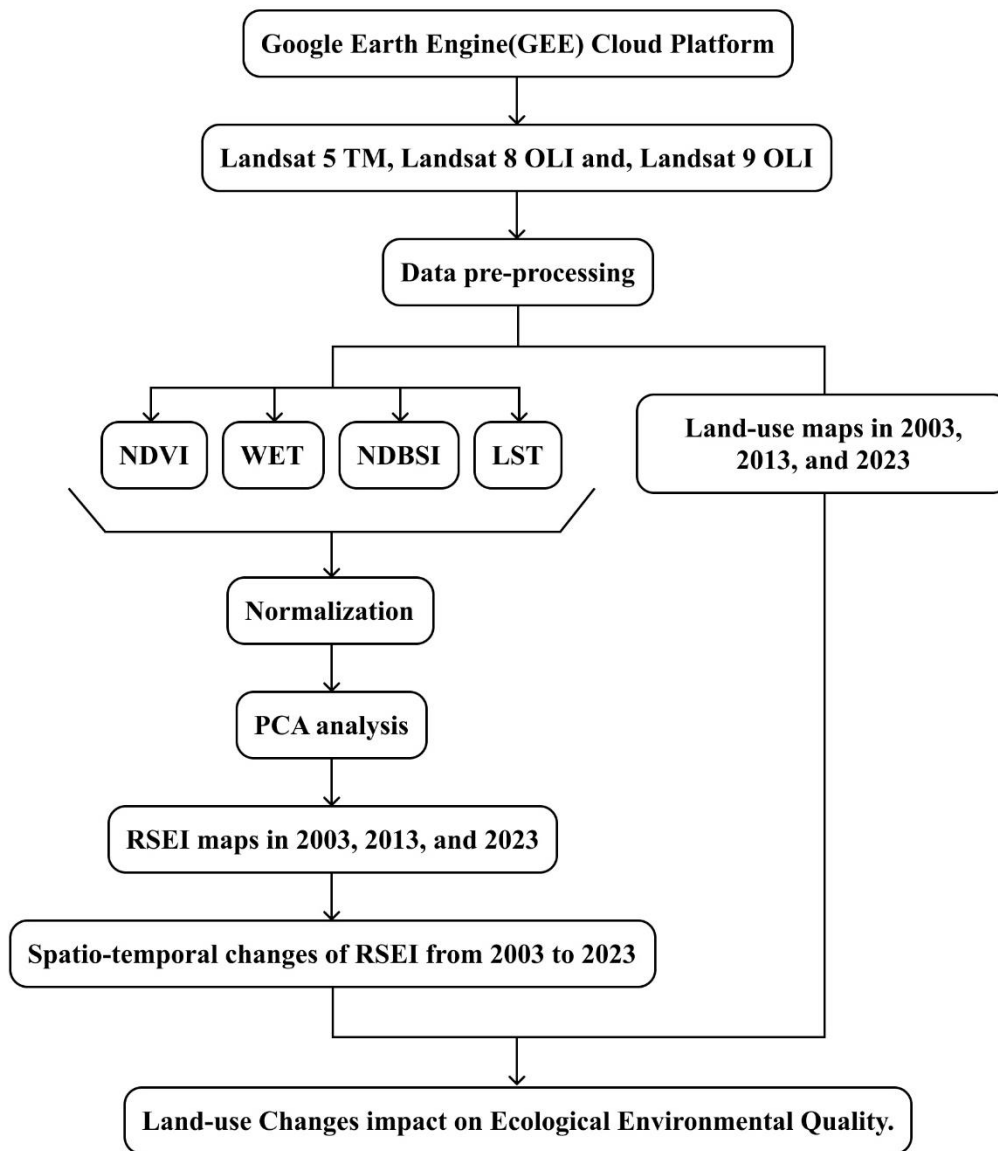
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**Fig. 1** Chattogram metropolitan area in Bangladesh.

## 130 2.2 Data source

131 This paper investigates the characteristics of the spatio-temporal variation of the EEQ in CMA  
132 in 2003, 2013, and 2023. Fig 2 presents a comprehensive outline of the study's workflow. To  
133 achieve this, an RSEI model was developed using principal component analysis (PCA) and  
134 Landsat satellite images from the GEE (<https://earthengine.google.org/>) platform as the data  
135 source. RSEI was constructed as a dependent variable using the four indices: NDVI, LST,  
136 WET, and NDBSI as independent variables. We used Landsat imagery during the best growth  
137 season (March–April) every year to reduce the uncertainty driven by seasonal variations.  
138 Additionally, a quality control band is employed to remove cloud shadows from satellite  
139 images. The water influence was eliminated from each indication by masking the water with  
140 the normalized difference water index (NDWI). To ensure reliability, all raster images were  
141 reprojected to the WGS\_1984\_UTM\_Zone\_46N for further analysis.



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**Fig. 2** Flow diagram of the methodology used for this study.

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### **2.3 Land use land cover classification**

145

Land-use maps of the study area were obtained using supervised classifications performed on Landsat TM+OLI imagery in 2003, 2013, and 2023. The random forest algorithm in the GEE platform was used for this purpose. The land use is categorized into five classes: vegetation, cropland, water body, urban areas, and bare land. To assess the accuracy of the classification process, a confusion matrix was used to calculate the Kappa index, as well as the overall user's and producer's accuracy (Congalton 1991). The overall classification accuracy of each LULC map ranged from 0.76% to 0.81% (Table 2).

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### **2.4 Remote sensing-based ecological index**

153 Relying solely on remote-sensing data, RSEI serves as a comprehensive tool for swift  
 154 assessment of ecological and environmental conditions in any specific geographic area. There,  
 155 we calculate the RSEI index by integrating four indices: NDVI, WET, LST, and NDBSI.  
 156 ArcGIS 10.8 software normalizes each index [0, 1] after calculating these indices in GEE. This  
 157 software then performs PCA, and the output PC1 (first principal component) represents the  
 158 initial value of RSEI. Finally, we generated the normalized RSEI by following Eq. (1) in the  
 159 raster calculator.(Airiken *et al.*, 2022; Mao *et al.*, 2023; Su *et al.*, 2023; Wang *et al.*, 2022).

$$160 \quad RSEI = \frac{(PC1 - PC1_{min})}{(PC1_{max} - PC1_{min})} \quad (1)$$

161 Where  $PC1_{min}$  and  $PC1_{max}$  represent the minimum and maximum value of PC1.

#### 162 **2.4.1 Normalized Difference Vegetation Index**

163 NDVI emerged as the most popular vegetation index in remote sensing applications. This index  
 164 is extensively employed to identify and quantify the presence of green vegetation on terrestrial  
 165 surfaces. In the context of remote sensing ecological assessments, the degree of greenness is  
 166 described by the following eq. (2) in the ecological index proposed by (Rouse *et al.*, 1973).

$$167 \quad NDVI = \frac{(\rho_{NIR} - \rho_{red})}{(\rho_{NIR} + \rho_{red})} \quad (2)$$

168 Here,  $\rho_{NIR}$  and  $\rho_{Red}$  are represented as the near-infrared band and red bands.

#### 169 **2.4.2 Wetness**

170 Tassel cap transformation facilitates the extraction of moisture, which is then expressed through  
 171 the resulting wet component (Zhang *et al.*, 2022). However, one needs to understand that the  
 172 coefficients associated with these components exhibit variability depending on the sensor used.  
 173 The calculation of the wet component in Landsat 5 and Landsat 8 can be performed according  
 174 to the following eq.(3) and eq.(4) (Lobser and Cohen 2007; Wang *et al.*, 2023).

175 Landsat 5 TM,

$$176 \quad Wet = 0.3102\rho_{Red} + 0.1594\rho_{NIR} + 0.0315\rho_{Blue} + 0.2021\rho_{Green} \\ 177 \quad \quad \quad - 0.6806\rho_{SWIR1} - 0.6109\rho_{SWIR2} \quad (3)$$

178 Landsat 8 OLI/TIRS,

$$179 \quad Wet = 0.3283\rho_{Red} + 0.3407\rho_{NIR} + 0.1511\rho_{Blue} + 0.1973\rho_{Green} \\ 180 \quad \quad \quad - 0.7171\rho_{SWIR1} - 0.4559\rho_{SWIR2} \quad (4)$$

181  $\rho_{Red}$ ,  $\rho_{Green}$ , and  $\rho_{Blue}$  represent as red, green, and blue bands.  $\rho_{SWIR}$  and  $\rho_{NIR}$  represent  
 182 as shortwave and near-infrared bands.

### 183 2.4.3 Land surface temperature

184 The single-chain method was used to estimate LST for this study area (Jimenez-Munoz *et al.*,  
 185 2009, 2014; Roy and Bari 2022; Sobrino *et al.*, 2004). The following equations were utilized  
 186 to calculate LST in GEE platform.

$$187 \quad LST = \frac{T_b}{1 + \left(\lambda * \frac{T_b}{\rho}\right) \ln \varepsilon} \quad (5)$$

188  $T_b$  represents the brightness temperature calculated by following the formula,

$$189 \quad T_b = \frac{k_2}{\ln(k_1/L_\lambda + 1)} \quad (6)$$

$$190 \quad L_\lambda = Gain * DN + Bias \quad (7)$$

191  $L_\lambda$  represent the amount of electromagnetic energy emitted from a surface. Gain and Bias are  
 192 calibration coefficients that are specific to each Landsat band. They are applied to the DN value  
 193 to account for variations in sensor response and to convert the DN to a physical unit of radiance.

194  $k_1$  and  $k_2$  are the thermal conversion constant for Landsat 5 TM and Landsat 8 OLI. For the  
 195 TM sensor,  $k_1$  and  $k_2$  respectively 607.76 mW cm<sup>-2</sup> and 1260.56 K. for the OLI sensor,  $k_1$  and  
 196  $k_2$  respectively 774.89 mW cm<sup>-2</sup> and 1321.08 K (Karbalaei Saleh *et al.*, 2021; Yue *et al.*, 2019).  
 197  $\lambda$  indicates the wavelength for thermal band (11.435  $\mu$ m for Landsat 5 TM and 10.9  $\mu$ m for  
 198 Landsat 8 OLI),  $\rho$  is the constant value ( $1.438 \times 10^{-2}$  mK), and  $\varepsilon$  denotes the emissivity which  
 199 is calculated as follows:

$$200 \quad Emissivity (\varepsilon) = 0.004 * pv + 0.986 \quad (8)$$

201 Where,  $pv$  represent the proportion of vegetation. Minimum and maximum value of  $NDVI$  and  
 202 following eq.(9) used to calculate the  $pv$

$$203 \quad Proportion\ of\ vegetation\ (pv) = \left[ \frac{(NDVI - NDVI_{min})}{(NDVI_{max} - NDVI_{min})} \right]^2 \quad (9)$$

### 204 2.4.4 Normalized difference built-up soil index

205 NDBSI recognized for its efficacy in assessing dryness, involves the averaging of two indices:  
 206 The built-up index (IBI) and the soil index (SI) (Liu *et al.*, 2022). These indices characterize

207 the respective conditions of urban areas and bare land (Hu and Xu 2018). NDBSI is calculated  
 208 as follows:

$$209 \quad IBI = \frac{\{2\rho SWIR1 / (\rho SWIR1 + \rho NIR) - [\rho NIR / (\rho NIR + \rho Red) + \rho Green / (\rho Green + \rho SWIR1)]\}}{\{2\rho SWIR1 / (\rho SWIR1 + \rho NIR) + [\rho NIR / (\rho NIR + \rho Red) + \rho Green / (\rho Green + \rho SWIR1)]\}} \quad (10)$$

$$210 \quad SI = \frac{[(\rho SWIR1 + \rho Red) - (\rho Blue + \rho NIR)]}{[(\rho SWIR1 + \rho Red) + (\rho Blue + \rho NIR)]} \quad (11)$$

$$211 \quad NDBSI = \frac{(IBI + SI)}{2} \quad (12)$$

212 In order to uniform these four indices, should be normalized [0,1]. The normalization process  
 213 is as follows:

$$214 \quad Normalized\ Index = \frac{Index - Index_{min}}{Index_{max} - Index_{min}} \quad (13)$$

215 Here,  $Index_{min}$  and  $Index_{max}$  define the minimum and maximum value of the index,  
 216 respectively.

## 217 **2.5 Principal Component Analysis**

218 PCA is a statistical technique that is used for data reduction, identifying the components that  
 219 are responsible for the overall variability within the values taken into account (Akbar *et al.*,  
 220 2022; Li *et al.*, 2021). Additionally, it is used in the scientific arena for various purposes e.g.  
 221 extracting and identifying the initial components, selecting the significant elements that are to  
 222 be retained in the model (Petrişor *et al.*, 2012). In this study, four specific ecological factors  
 223 named NDVI, LST, NDBSI, and WET are used, and then calculate the RSEI through PCA  
 224 using ArcGIS 10.8 software.

## 225 **3. Results**

### 226 **3.1 Spatial-temporal changes of LULC**

227 The use and arrangement of land significantly impacts ecosystem health and overall  
 228 environmental well-being, serving as vital determinants of habitat quality, biodiversity, and  
 229 resource availability (Yang *et al.*, 2019). Over the past two decades, from 2003 to 2023, we  
 230 classified the CMA into five distinct LULC classes. These classes include vegetation, cropland,  
 231 water bodies, urban areas, and bare land. Vegetation primarily occupied the hilly terrains  
 232 located in the northwest and southern regions of the study area, forming dense patches that  
 233 serve as crucial biodiversity hotspots and contribute to ecological stability by providing

234 essential ecosystem services, such as carbon sequestration and soil erosion prevention. The  
 235 study region predominantly found crops in areas adjacent to major rivers, distributing them  
 236 sporadically to capitalize on fertile soils and water access for agricultural productivity. This  
 237 extensive cropland supports the regional economy, providing food security and employment  
 238 opportunities for local communities. Meanwhile, economic activities, population growth, and  
 239 infrastructure development have driven rapid urbanization trends, resulting in a dense  
 240 concentration of urban land in the central part of the CMA. This transformation into built-up  
 241 areas has implications for both environmental health, particularly through increased pollution  
 242 and the urban heat island effect, as well as social dynamics, such as housing demand and  
 243 congestion. We observed scattered water ponds, including lakes, reservoirs, and small ponds,  
 244 throughout the region, providing critical ecosystem services such as freshwater supply, climate  
 245 regulation, and aquatic habitats. However, these water bodies often appeared fragmented, likely  
 246 indicating anthropogenic pressures such as encroachment or over-extraction (Fig 3). Lastly,  
 247 areas unsuitable for either agriculture or urban expansion primarily featured bare land, mostly  
 248 composed of exposed soil and rocky outcrops. These areas, while naturally occurring in some  
 249 cases, could also represent degraded land, potentially due to unsustainable agricultural  
 250 practices, deforestation, or construction activities. The dynamic interplay among these LULC  
 251 classes over time provides insight into the shifting landscape patterns of the CMA and  
 252 highlights the need for sustainable land management practices to ensure balanced  
 253 environmental health.

254 **Table 1** LULC area in CMA from 2003 to 2023.

LULC Types	Years					
	2003		2010		2023	
	Area (km <sup>2</sup> )	Area (%)	Area (km <sup>2</sup> )	Area (%)	Area (km <sup>2</sup> )	Area (%)
Vegetation	334.97	46.25	318.95	44.04	258.06	35.63
Cropland	313.64	43.3	288.19	39.79	307.89	42.51
Waterbody	18.65	2.58	13.96	1.93	15.76	2.18
Urban	34.08	4.71	80.38	11.1	101.98	14.08
Bare Land	22.91	3.16	22.77	3.14	40.56	5.6

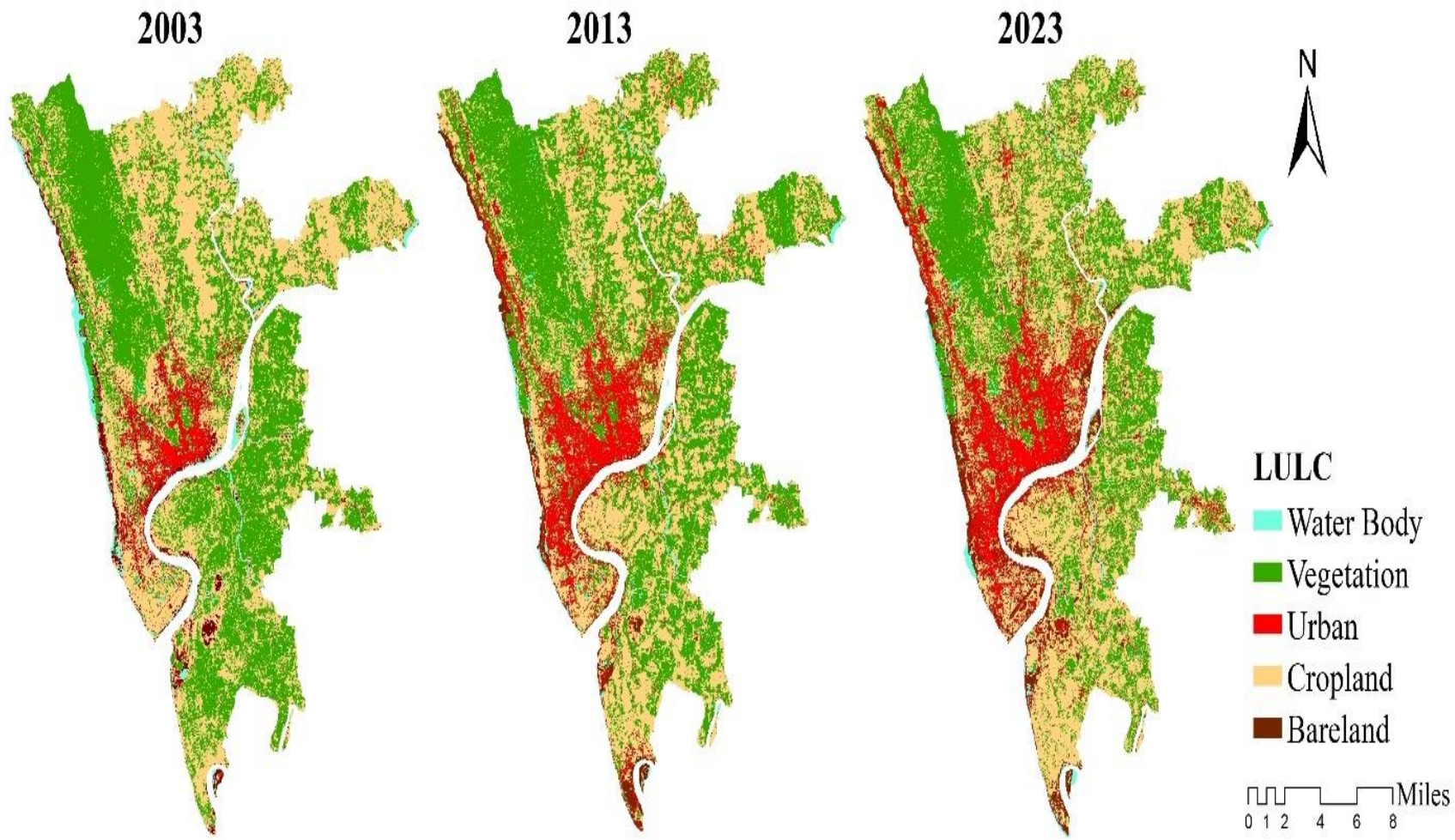
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256 **Table 2** Accuracy assessment of LULC maps.

LULC Types	Years					
	2003		2013		2023	
	P	U	P	U	P	U
Vegetation	0.91	0.79	0.89	0.84	1.00	1.00
Cropland	0.94	0.84	0.74	0.82	0.91	0.78
Waterbody	0.94	1.00	0.93	1.00	1.00	0.89
Urban	0.76	0.78	0.89	0.75	0.69	0.79
Bare Land	0.70	0.74	0.50	0.74	0.72	0.67

<b>Overall accuracy (%)</b>	0.81	-	0.78	-	0.76	-
<b>Kappa coefficient</b>	0.75	-	0.67	-	0.66	-

257 \* p: producer accuracy; U: user accuracy

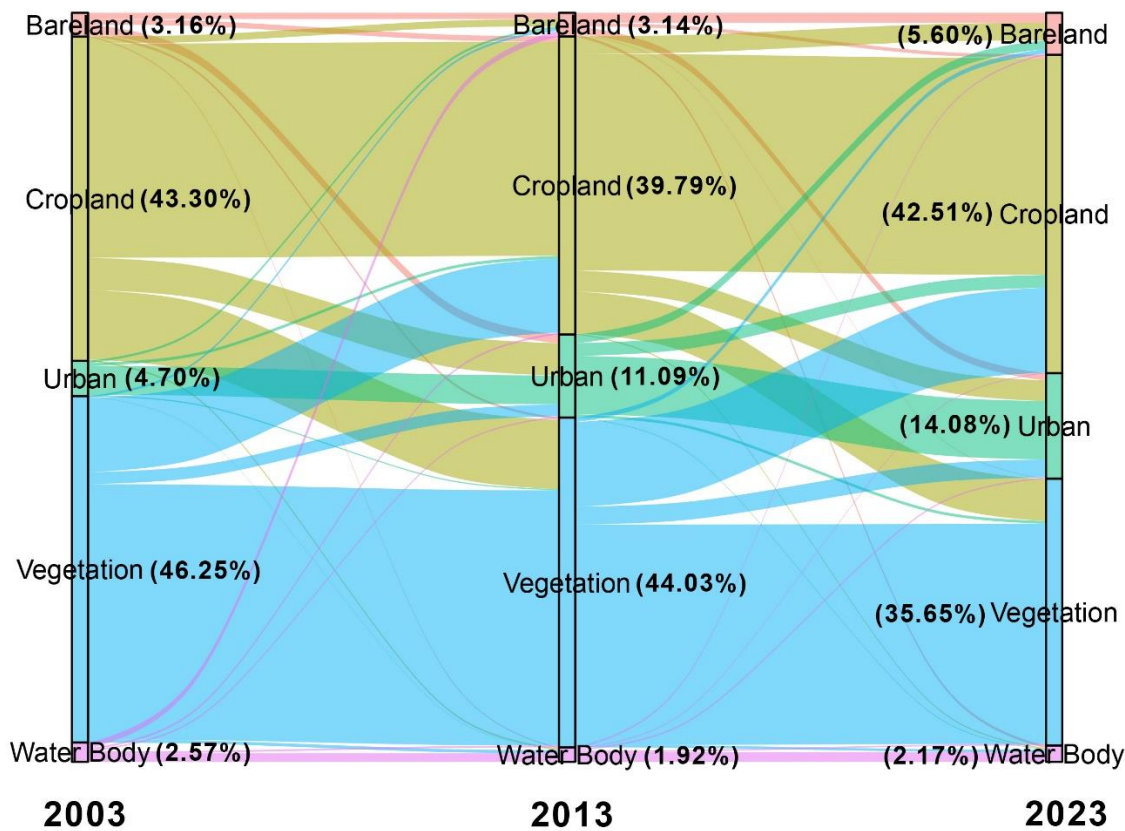


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**Fig 3:** Spatial distribution of LULC in CMA from 2003 to 2023.

260 Table 1 shows that more than 77% of the land was covered by vegetation and cropland, indicating  
 261 the strong presence of forestry and agriculture in the area. The coverage of vegetation and cropland  
 262 decreased significantly from 46.25% (334.98 km<sup>2</sup>) and 43.31% (313.64 km<sup>2</sup>) in 2003 to 35.63%  
 263 (258.06 km<sup>2</sup>) and 42.51% (307.89 km<sup>2</sup>) in 2023, respectively. The LULC classes showed a gradual  
 264 expansion in urban areas over the last two decades, indicating social progress. The urban area  
 265 expanded from 34.08 km<sup>2</sup> (4.71%) in 2003 to 101.99 km<sup>2</sup> (14.08%) in 2023, an increase of 67.91  
 266 km<sup>2</sup> in 20 years. The annual expansion rate was 2.34 km<sup>2</sup> (0.46%) % each year. This growth is  
 267 linked to the city's GDP which increased from \$16 billion in 2005 to \$39 billion in 2020 (Miah  
 268 and Hossain 2022). All classes, excluding urban areas and bare land, experienced a decrease, with  
 269 the vegetation class showing the largest decline.



270

271 **Fig. 4** LULC transition from 2003 to 2023 using Sankey Diagram.

272 The LULC transfer matrix was used to create a Sankey diagram (Fig 4) that effectively illustrates  
 273 the origin and destination of each land use category qualitatively and quantitatively. From the Fig

274 4, it shows that vegetation land holds a dominant position and tends to decline. Moreover, the  
 275 vegetation land is not only transferred within its category but primarily transferred to cropland and  
 276 urban land. During 2013-2023, most of the vegetated land was converted to cropland and urban  
 277 land, covering 81.99 km<sup>2</sup> and 17.34 km<sup>2</sup> respectively. Cropland follows as the second most  
 278 converted land type, with a significant portion being transferred to vegetation and urban land at  
 279 67.13 km<sup>2</sup> and 31.174 km<sup>2</sup>, respectively, from 2003 to 2013. Conversely, urban and bare land,  
 280 along with water bodies, exhibit comparatively lower transfer volumes. From 2003 to 2023, urban  
 281 areas in CMA increased the most among all types of land use. In CMA, urban lands are mostly  
 282 composed of converted cropland, vegetation, and bare land. This rapid urbanization is leading to  
 283 ecological degradation and impacting SDGs in the region.

### 284 3.2 Spatial-temporal changes in ecological quality

285 The research analyzed various ecological factors like NDVI, LST, NDBSI, and WET to calculate  
 286 the RSEI through PCA using ArcGIS 10.8 software. The PCA results of CMA in 2003, 2013, and  
 287 2023 are tabulated. According to Table 3, PC1 eigenvalues had a contribution rate of over 80%.  
 288 This indicates that the first principal component accurately represents the important features of  
 289 each factor. The NDVI and WET index from 2003 to 2023 are positive values, indicative of their  
 290 constructive impact on the local ecological landscape. On the contrary, the NDBSI and LST  
 291 showed adverse values, suggesting a detrimental influence impeding ecological restoration and  
 292 enhancement efforts within the study area.

293 **Table 3:** PCA dataset of RSEI.

Year	Indices	PC1	PC2	PC3	PC4	Mean RSEI
2003	NDVI	0.676	-0.543	0.428	0.254	0.616
	LST	-0.311	0.327	0.892	0.022	
	NDBSI	-0.628	-0.550	-0.031	0.549	
	WET	0.227	0.543	-0.140	0.796	
	<b>Eigenvalue</b>	0.018	0.002	0.001	0.000	
	<b>Percentage</b>	84.91	8.07	6.67	0.35	
2013	NDVI	0.711	-0.010	0.696	0.098	0.627
	LST	-0.353	0.859	0.370	0.015	
	NDBSI	-0.599	-0.495	0.566	0.275	
	WET	0.105	0.130	-0.239	0.956	
	<b>Eigenvalue</b>	0.019	0.001	0.001	0.000	
	<b>Percentage</b>	88.98	6.57	4.37	0.08	
2023	NDVI	0.720	-0.027	0.666	0.194	0.621
	LST	-0.326	0.866	0.377	0.035	
	NDBSI	-0.576	-0.435	0.453	0.524	
	WET	0.209	0.245	-0.458	0.829	
	<b>Eigenvalue</b>	0.017	0.002	0.001	0.000	
	<b>Percentage</b>	82.83	10.40	6.53	0.25	

294 To better understand the ecological environment's changes and attributes within the study area, the  
 295 RSEI has been categorized into five distinct levels based on pertinent literature: worst (0–0.2),  
 296 poor (0.2–0.4), moderate (0.4–0.6), good (0.6–0.8), and excellent (0.8–1) (Su *et al.*, 2023; Sun *et*  
 297 *al.*, 2022; X. Wang *et al.*, 2022). Over the period of 2003 to 2023, the average RSEI in the study  
 298 area was above 0.6. This indicates that all areas in the study were rated as good according to the  
 299 RSEI classification criteria.

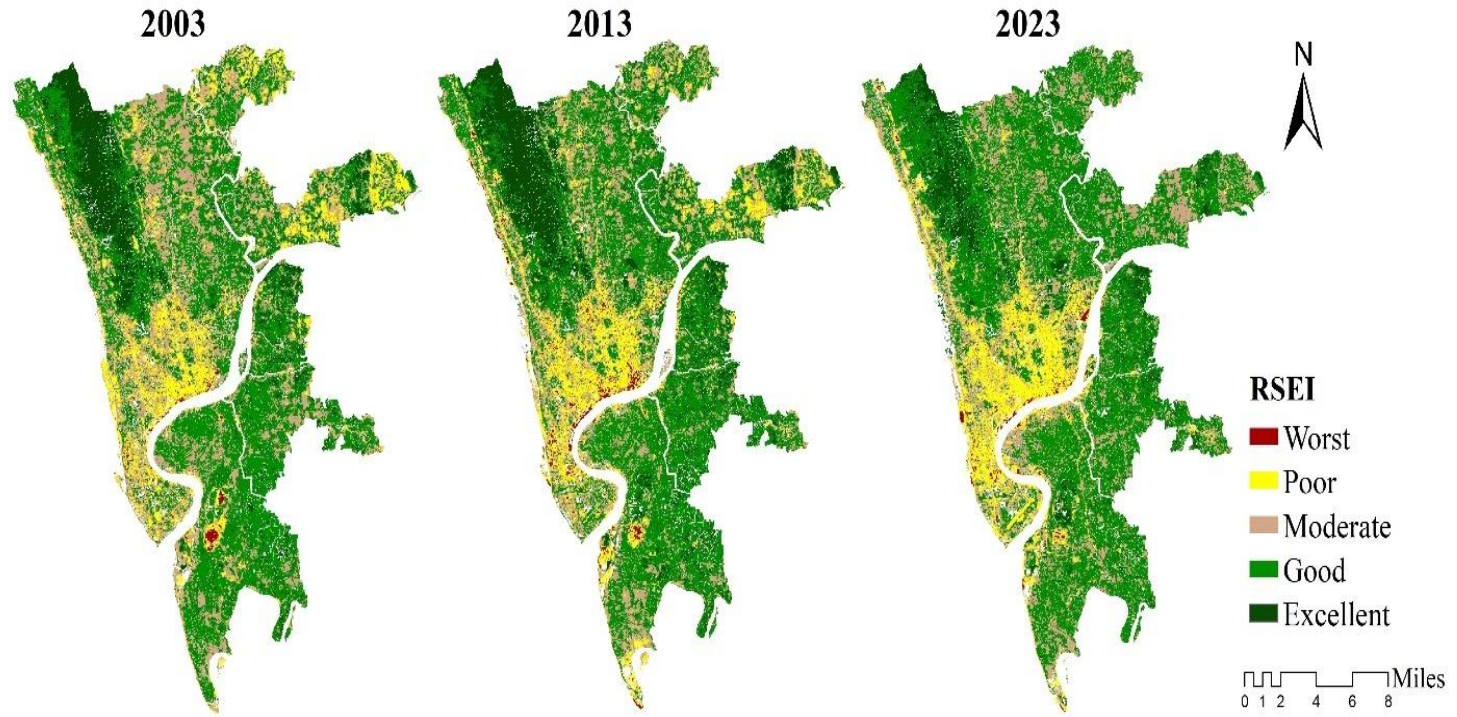
300 The temporal and spatial distribution of the EEQ of the study area has been shown in Fig 5 from  
 301 2003 to 2023. In 2003, there were small and scattered areas with low EEQ, particularly clustered  
 302 near Bangladesh's main seaport. Human activities significantly affected these areas, which were  
 303 mostly on the outskirts of the main urban area. The hilly regions in the northeast and south  
 304 exhibited excellent EEQ. The agricultural lands primarily range from good to moderate EEQ  
 305 levels. The red areas with poor EEQ gradually expanded over the next few years, particularly in  
 306 2008, 2014, and 2018, demonstrating a polycentric clustering expansion trend compared to the  
 307 baseline in 2003.

308 **Table 4** RSEI area calculation.

Grade/years	Worst (0-0.2)		Poor (0.2-0.4)		Moderate (0.4-0.6)		Good (0.6-0.8)		Excellent (0.8-1)	
	Area (km <sup>2</sup> )	%	Area (km <sup>2</sup> )	%	Area (km <sup>2</sup> )	%	Area (km <sup>2</sup> )	%	Area (km <sup>2</sup> )	%
<b>2003</b>	4	0.6	68	10.0	213	31.2	316	46.3	81	11.9
<b>2013</b>	9	1.3	72	10.6	175	25.7	321	47.1	104	15.3
<b>2023</b>	5	0.7	71	10.4	181	26.6	362	53.2	62	9.1

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312

**Fig. 5** Spatial distribution of RSEI in CMA from 2003 to 2023.

313 The study examined the size and proportion of each ecological grade within the study area over  
 314 three different years (Table 4). In 2003, the study area was 681 km<sup>2</sup>, with 71 km<sup>2</sup> classified as  
 315 worst EEQ and 71 km<sup>2</sup> as poor EEQ, making up 10.4% of the total area. By 2013, this percentage  
 316 increased but then decreased to 11.1% by 2023. Moderate EEQ land was 213 km<sup>2</sup> in 2003,  
 317 accounting for 31.3% of the total area. This decreased to 25.7% by 2013 and slightly increased to  
 318 26% by 2023. Land with good EEQ levels showed a gradual increase from 2003 to 2023. Excellent  
 319 EEQ areas were 11.9% in 2003, increased to 15.3% by 2013, but decreased to 9.1% by 2023,  
 320 indicating a conversion of vegetation land to cropland or urban areas.

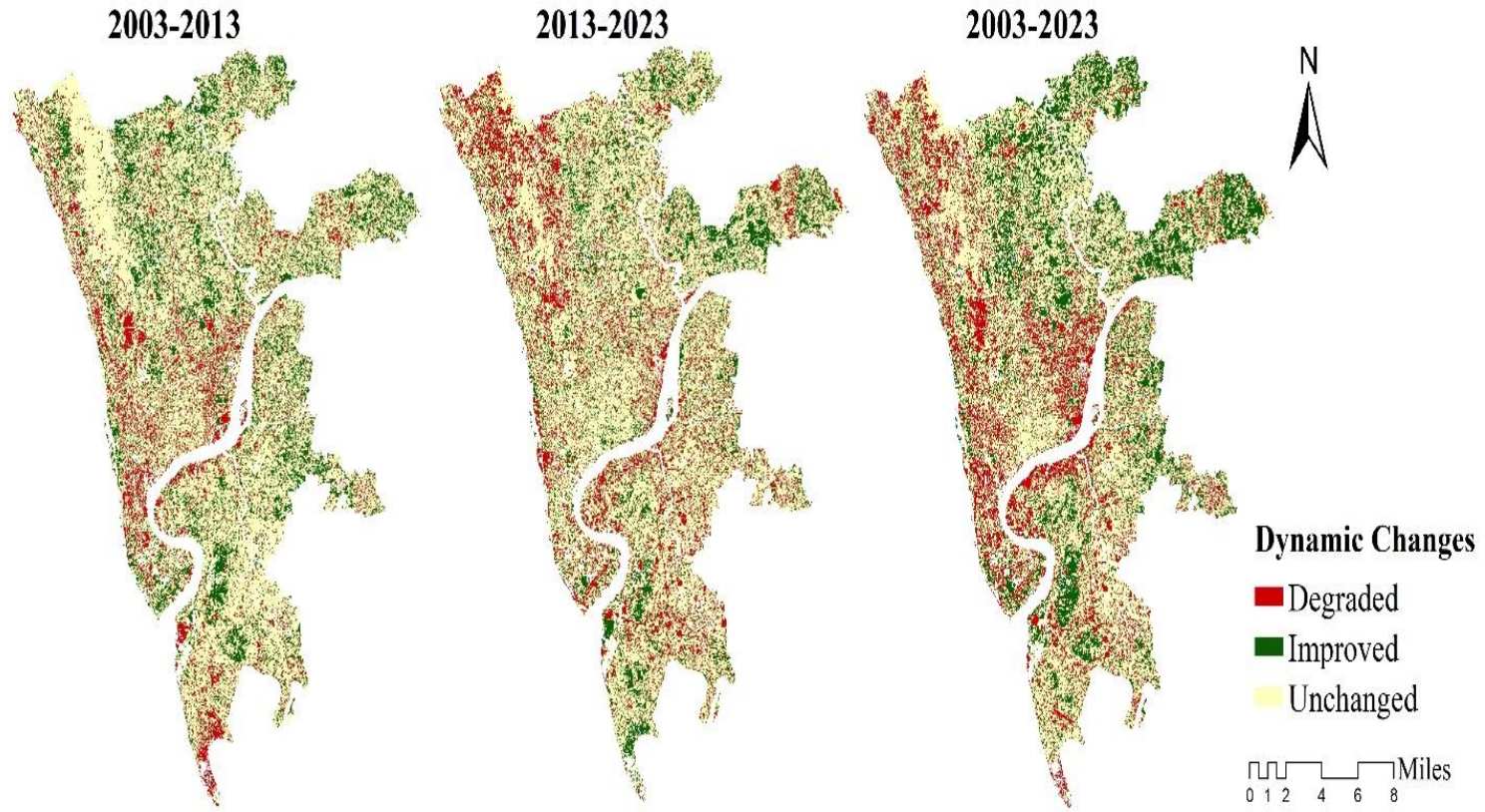
321 The changes in RSEI detection in CMA from 2003 to 2023 are illustrated in Fig 6. The ecological  
 322 environment degradation was primarily focused on the central regions due to increased human  
 323 activities. Urban construction and land-reclamation activities in the central and economic zones  
 324 between 2003 and 2013 damaged the natural landscape and produced pollutants, affecting  
 325 ecological quality.

326 **Table 5** Dynamic change of ecological quality over the years.

Year	Improved (km <sup>2</sup> )	%	Unchanged (km <sup>2</sup> )	%	Degraded (km <sup>2</sup> )	%
2003 – 2013	139	21.4	413	63.4	99	15.2
2013 – 2023	108	16.5	407	62.5	137	21.0
2003 – 2023	151	23.1	355	54.5	146	22.4

327

328 Between 2013 and 2023, urban construction on the island shifted to developing within the existing  
 329 urban area because of the limited availability of appropriate land resources. This change resulted  
 330 in a moderate decline in ecological quality. From 2003 to 2013, 63.4% of the entire area remained  
 331 unchanged in ecological quality, accounting for 413 km<sup>2</sup> (Table 5). During this period, degenerated  
 332 areas accounted for 15.2% of the entire area at 99 km<sup>2</sup>, while improved regions covered 21.4% of  
 333 the total area, which is 139 km. From 2013 to 2023, degraded areas accounted for 21% of the entire  
 334 area, improved regions covered 16.5%, and unchanged areas comprised 62.5%. These findings  
 335 indicate that the study area experienced more ecological degradation than improvement during this  
 336 period. Comparing the entire span from 2003 to 2023, degraded areas constituted 22.4%,  
 337 ecologically improved areas comprised 23.1%, and unchanged areas accounted for 54.5% of the

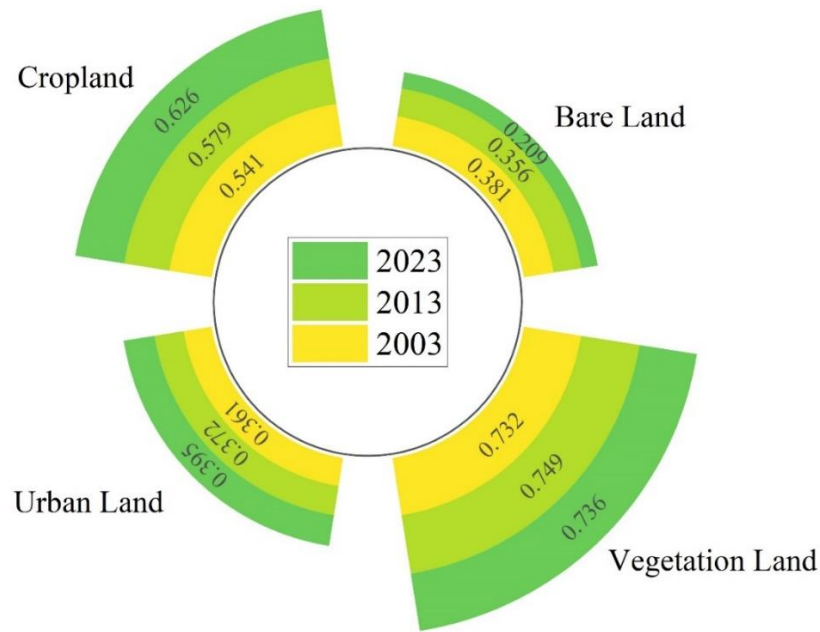


**Fig. 6** Spatial changes in RSEI over 2003 to 2023.

341 study area. From 2003 to 2023, there were significant land transformations from forests and  
342 croplands to urban areas. However, the overall ecological quality did not see a significant decline  
343 during this period.

### 344 3.3 Land use responses to ecological quality

345 The images of Landsat 5 in 2003 and Landsat 8 in 2013 and 2023 were used in our experiment.  
346 They were also used to calculate the indices required to obtain the formulated RSEI and map land  
347 use. The RSEI was greatly impacted using scientific urban planning and ecological environment  
348 planning. We analyzed and processed data on land use type formation and the remote sensing  
349 ecological index to better understand the relationship between land use type and the ecological  
350 environment. As a result, we obtained the RSEI of each land use type (Fig 7). The study found  
351 differences in EEQ among different land use types. Compared to other land use types, vegetated  
352 land had the most RSEI value between 2003 and 2023. Furthermore, the RSEI values on bare and  
353 urban land were among the lowest.



354

355 **Fig. 7** RSEI value of each land use type from 2003 to 2023.

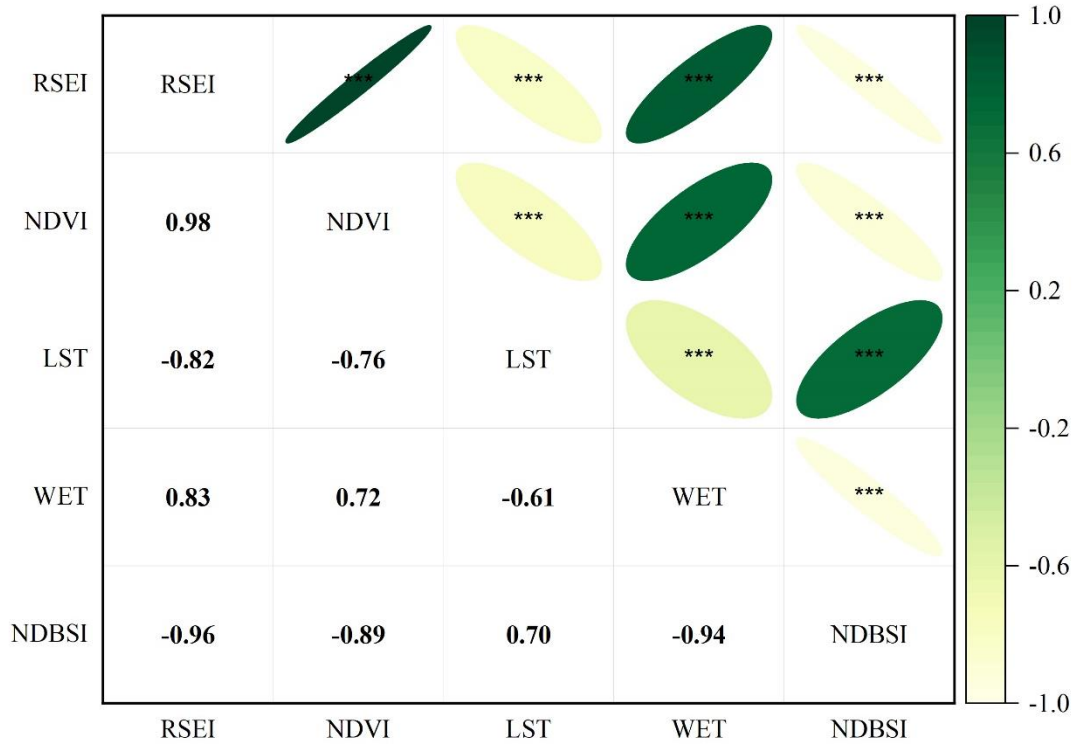
356

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### 359 3.3 Correlation analysis among indices

360 For correlation analysis, the values of RSEI, NDVI, NDBSI, WET, and LST were extracted from  
361 the tabular format of the converted raster-to-point data of each image, using the 2013 Landsat  
362 image data as an example. We used Origin Pro to conduct correlation analysis. The results of the  
363 analysis can be seen in Fig 8. Green shading represents a positive correlation, with darker shades  
364 indicating a stronger positive correlation, while white indicates a negative correlation. The NDVI  
365 and WET indicators demonstrated a strong positive correlation ( $R^2 = 0.98, 0.83$ ) with RSEI,  
366 whereas they exhibited a significant negative correlation ( $R^2 = -0.82, -0.96$ ) with LST and NDBSI.  
367 A significant positive correlation ( $R^2 = 0.72$ ) was observed between WET and NDVI, while a  
368 significant negative correlation ( $R^2 = -0.76, -0.89$ ) was found with LST and NDBSI. WET and  
369 NDBSI displayed significant negative and positive correlations ( $R^2 = -0.61, 0.70$ ) with LST,  
370 respectively. On the other hand, NDBSI exhibited a significant negative correlation ( $R^2 = -0.94$ )  
371 with WET. As a result, the research above suggests that there is a significant relationship between  
372 RSEI and each measure. RSEI is more reflective of overall ecological quality and more effectively  
373 synthesizes information from each variable when compared to the usage of individual remote  
374 sensing-based metrics. Therefore, the RSEI was utilized to represent the studied area's entire  
375 ecological quality.



\* p<=0.05 \*\* p<=0.01 \*\*\* p<=0.001

**Fig. 8** Correlation matrix of RSEI.

#### 4. Discussion

The GEE platform was employed to utilize the cloud-free images, and median equalization techniques were applied to enhance the original image quality. The platform allows for convenient access to extensive data at any time (Kumar and Mutanga 2018; Shelestov *et al.*, 2017; Wang *et al.*, 2022). The application of the Random Forest algorithm for land use classification demonstrated robust performance and high accuracy in distinguishing between different land cover types. During training, Random Forest creates several decision trees as a method of collaborative learning and outputs the mode of classes for each tree. This method successfully reduces the overfitting issue that distinct decision trees can have, improving the model's predictability (Gislason *et al.*, 2006; Svoboda *et al.*, 2022; Tikuye *et al.*, 2023). In this study, the Random Forest algorithm was particularly adept at handling the complexity and heterogeneity inherent in land use data. Unfortunately, to eradicate the detrimental effect of enormously water regions on the normalization process and eigenvalue of WET and NDBSI in the RSEI calculation, the NDWI was utilized in

391 this work to mask the large water areas (Acharya *et al.*, 2018). In other words, our research  
392 concentrated on assessing ecological quality in land regions rather than water bodies. PC1  
393 eigenvalues had a contribution rate of over 80%. This indicates that the first principal component  
394 accurately represents the important features of each factor, which were like previous studies (Maity  
395 *et al.*, 2022; Wang *et al.*, 2022; Xu *et al.*, 2019). Our study showed that RSEI negatively correlated  
396 with LST and NDBSI, suggesting detrimental impacts, while NDVI and WET showed positive  
397 correlations, indicating beneficial effects. These findings are consistent with previous studies  
398 (Airiken *et al.*, 2022; Karbalaie Saleh *et al.*, 2021; Maity *et al.*, 2022; Mao *et al.*, 2023; Su *et al.*,  
399 2023; Wang *et al.*, 2022; Xu *et al.*, 2019). The impact of the vegetation indication outweighed the  
400 other indicators (Table 2). The CMA's ecological quality ranged from 0.616 to 0.621, which is  
401 considered good quality. Ecological quality remained mostly stable and even showed a slight  
402 improvement. The RSEI was greatly impacted of scientific urban planning and ecological  
403 environment planning.

404 The RSEI values differed in various landscapes, such as 0.21-0.38 in barren areas and 0.36-0.39  
405 in urban areas (Guo *et al.*, 2017; Yue *et al.*, 2019), greater than 0.73 in vegetation regions (Xu *et al.*  
406 *et al.*, 2019), and even 0.54-0.63 in cropland (Shan *et al.*, 2019). The result of the RSEI value of each  
407 land use type supports that forests generally enhance the ecological quality of the land by  
408 supporting biodiversity, regulating climate and water cycles, and providing various ecosystem  
409 services. In contrast, construction land often degrades ecological quality through habitat  
410 destruction, soil degradation, increased stormwater runoff, heat island effects, and loss of green  
411 space. Sustainable land use planning and development practices are essential to mitigate the  
412 negative impacts of construction activities and preserve the ecological integrity of landscapes.  
413 There are two possible ways to apply the RSEI to SDG 15. I) RSEI is represented as an indicator  
414 for the construction industry which can be applied to make buildings more sustainable and affect  
415 the protection of wildlife and the environment. II) The RSEI is also used as an indicator for  
416 terrestrial ecosystems, and it has the potential to substitute the improved vegetation index and  
417 water network density that are included in the biological diversity indicator system to meet  
418 sustainable development goals.

419

420

## 421 **5. Conclusion**

422 This study examined the patterns of land use changes and their effects on the ecological quality of  
423 CMA over the two decades (2003-2023), which provide valuable insights for the sustainable  
424 development of the region's society, economy, and ecology. The key findings are as follows:

425 I. The study found significant decreases in vegetation and cropland, alongside substantial  
426 increases in urban areas over the past two decades, indicating social progress but potential  
427 negative impacts on ecological health due to urban expansion. The urban areas expanded  
428 annually at a rate of 2.34 km<sup>2</sup>. The Sankey diagram showed that most new urban lands  
429 originated from cropland, vegetation, and bare land, with vegetation land largely being  
430 converted to cropland and urban land.

431 II. The NDVI and WET indices showed positive values from 2003 to 2023, indicating a  
432 positive impact on the local ecological landscape, while NDBSI and LST showed negative  
433 values, indicating a negative impact. Vegetation land had the highest RSEI values during  
434 this period, suggesting better ecological quality, whereas bare and urban land had the  
435 lowest RSEI values, suggesting worse ecological quality. Land with vegetation led to  
436 ecological improvement, while land converted to urban land led to ecological degradation.

437 III. All the indicators exhibited a significant linear correlation with RSEI and with each other.  
438 Furthermore, NDVI and NDBSI showed higher correlation coefficients than LST and  
439 WET.

440 This study helps policymakers understand the trade-offs and synergism among different  
441 development objectives. It highlights potential conflicts between economic growth and  
442 environmental sustainability, allowing policymakers to devise strategies that balance competing  
443 interests and maximize sustainable outcomes.

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### 447 **Declaration of competing interests**

448

449 The authors declare no conflicts of interest.

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453 **Data Availability**

454 Data will be made available on request.

455 **References**

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